INTERSTATE 5 COLUMBIA RIVER CROSSING

Ecosystems Technical Report for the Final Environmental Impact Statement



May 2011



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Cover Sheet

Interstate 5 Columbia River Crossing

Ecosystems Technical Report for the Final Environmental Impact Statement:

Submitted By:

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Interstate 5 Columbia River Crossing Ecosystems Technical Report for the Final Environmental Impact Statement		
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ACRONYMS

Acronym Description

μPa 1 micropascal

API area of potential impact

BA biological assessment

BES City of Portland Bureau of Environmental Services

BGEPA Bald and Golden Eagle Protection Act

BLM Bureau of Land Management
BMP best management practice

BNSF Burlington Northern Santa Fe Railroad

BOD biological oxygen demand

BPA Bonneville Power Administration

CAO critical areas ordinance

CD collector-distributor

CIA contributing impervious area

cfs cubic feet per second

COD chemical oxygen demand

COP City of Portland
CR Columbia River

CRC Columbia River Crossing

CTR Commute Trip Reduction (Washington)

C-TRAN Clark County Public Transportation

dB decibel

DDE dichlorodiphenyldichloroethylene

DEIS Draft Environmental Impact Statement

DEQ Oregon Department of Environmental Quality

DOT U.S. Department of Transportation

DPS distinct population segment

DSL Oregon Department of State Lands

ECO Employee Commute Options (Oregon)

Ecology Washington Department of Ecology

ESC erosion and spill control

EFH essential fish habitat

EPA United States Environmental Protection Agency

environmental impact statement

EIS

ESA Endangered Species Act

ESEE economic, social, environmental, and energy

ESH essential salmonid habitat
ESU evolutionarily significant unit

FEIS Final Environmental Impact Statement

FEMA Federal Emergency Management Agency

FHWA Federal Highway Administration

fps feet per second

FTA Federal Transit Administration

FPC Fish Passage Center

GIS Geographic Information System

GMA Growth Management Act
GMAV Genus Mean Acute Values
HPA hydraulic project approval

I-5 Interstate 5

InterCEP Interstate Collaborative Environmental Process

LCR Lower Columbia River

LPA Locally Preferred Alternative

LRV light rail vehicle

MAX Metropolitan Area Express
MBTA Migratory Bird Treaty Act

MCDD Multnomah County Drainage District

Metro Metropolitan Regional Government

MHCC Mount Hood Community College

MMPA Marine Mammal Protection Act

MSFCMA Magnuson-Stevens Fisheries Conservation and Management Act

NAS National Academy of Sciences

NMFS National Marine Fisheries Service

NOAA National Oceanic and Atmospheric Administration

NPDES National Pollutant Dishcharge Elimination System

NR Natural Resource

NRCS Natural Resources Conservation Service

NRMP natural resource management plan

OAR Oregon Administrative Rule

ODA Oregon Department of Agriculture

ODFW Oregon Department of Fish and Wildlife

ODOT Oregon Department of Transportation

OHW ordinary high water line

ORNHIC Oregon Natural Heritage Information Center

ORS Oregon Revised Statutes

OTC Oregon Transportation Commission

PAH polycyclic aromatic hydrocarbon

PCE primary constituent element
PDX Portland International Airport

PGIS pollution-generating impervious surface

RM river mile

RMS root-mean-square

ROD Record of Decision

RTC Regional Transportation Council

sq. ft. square feet

SEPA State Environmental Policy Act
SMA Shoreline Management Act

SOC species of concern
SOI species of interest

SPCC spill prevention, control, and countermeasures

SPUI single-point urban interchange

SWCD East Multnomah Soil and Water Conservation District

TDM transportation demand management

Tri-County Metropolitan Transportation District

TSM transportation system management

TMDL total maximum daily load

TPH total petroleum hydrocarbons

UCR Upper Columbia River

USACE United States Army Corps of Engineers
USFWS United States Fish and Wildlife Service

UWR Upper Willamette River

WAC Washington Administrative Code

WDFW Washington Department of Fish and Wildlife

WDFW-PHS Washington Department of Fish and Wildlife, Priority Habitats and Species
WDNR-NHP Washington Department of Natural Resources, Natural Heritage Program

WNHP Washington Natural Heritage Program

WNWCB Washington Noxious Weed Control Board

Interstate 5 Columbia River Crossing Ecosystems Technical Report for the Final Environmental Impact Statement

WSDOT Washington State Department of Transportation

WTC Washington Transportation Commission

1. Summary

1.1 Introduction

For this report, ecosystem resources include fish, wildlife, and plants, and their habitats, within and around the Interstate 5 (I-5) Columbia River Crossing (CRC) project area. The key issues that are addressed in this report are listed below:

- The potential for impacts to special-status species.
- The potential for impacts to habitats that support fish, wildlife, and plants.
- The potential for impacts to protected habitats.
- The potential for impacts to other ecosystem resources, including migratory birds, marine mammals, rare plants, and noxious weeds.

Impacts and effects may be beneficial or adverse. This report addresses how each alternative may differ in its effect on ecosystems, as well as how regional conditions may be affected by the project overall.

1.2 Description of Alternatives

This technical report evaluates the CRC project's locally preferred alternative (LPA) and the No-Build Alternative. The LPA includes two design options: The preferred option, LPA Option A, which includes local vehicular access between Marine Drive and Hayden Island on an arterial bridge; and LPA Option B, which does not have arterial lanes on the light rail/multi-use path bridge, but instead provides direct access between Marine Drive and the island with collector-distributor (CD) lanes on the two new bridges that would be built adjacent to I-5. In addition to the design options, if funding availability does not allow the entire LPA to be constructed in one phase, some roadway elements of the project would be deferred to a future date. This technical report identifies several elements that could be deferred, and refers to that possible initial investment as LPA with highway phasing. The LPA with highway phasing option would build most of the LPA in the first phase, but would defer construction of specific elements of the project. The LPA and the No-Build Alternative are described in this section.

1.2.1 Adoption of a Locally Preferred Alternative

Following the publication of the Draft Environmental Impact Statement (DEIS) on May 2, 2008, the project actively solicited public and stakeholder feedback on the DEIS during a 60-day comment period. During this time, the project received over 1,600 public comments.

During and following the public comment period, the elected and appointed boards and councils of the local agencies sponsoring the CRC project held hearings and workshops to gather further public input on and discuss the DEIS alternatives as part of their efforts to determine and adopt a locally preferred alternative. The LPA represents the alternative preferred by the local and regional agencies sponsoring the CRC project. Local agency-elected boards and councils determined their preference based on the results of the evaluation in the DEIS and on the public and agency comments received both before and following its publication.

In the summer of 2008, the local agencies sponsoring the CRC project adopted the following key elements of CRC as the LPA:

- A replacement bridge as the preferred river crossing,
- Light rail as the preferred high-capacity transit mode, and
- Clark College as the preferred northern terminus for the light rail extension.

The preferences for a replacement crossing and for light rail transit were identified by all six local agencies. Only the agencies in Vancouver – the Clark County Public Transit Benefit Area Authority (C-TRAN), the City of Vancouver, and the Regional Transportation Council (RTC) – preferred the Vancouver light rail terminus. The adoption of the LPA by these local agencies does not represent a formal decision by the federal agencies leading this project – the Federal Highway Administration (FHWA) and Federal Transit Administration (FTA) – or any federal funding commitment. A formal decision by FHWA and FTA about whether and how this project should be constructed will follow the FEIS in a Record of Decision (ROD).

1.2.2 Description of the LPA

The LPA includes an array of transportation improvements, which are described below. When the LPA differs between Option A and Option B, it is described in the associated section. For a more detailed description of the LPA, including graphics, please see Chapter 2 of the FEIS.

1.2.2.1 Multimodal River Crossing

Columbia River Bridges

The parallel bridges that form the existing I-5 crossing over the Columbia River would be replaced by two new parallel bridges. The eastern structure would accommodate northbound highway traffic on the bridge deck, with a bicycle and pedestrian path underneath; the western structure would carry southbound traffic, with a two-way light rail guideway below. Whereas the existing bridges have only three lanes each with virtually no shoulders, each of the new bridges would be wide enough to accommodate three through-lanes and two add/drop lanes. Lanes and shoulders would be built to full design standards.

The new bridges would be high enough to provide approximately 95 feet of vertical clearance for river traffic beneath, but not so high as to impede the take-offs and landings by aircraft using Pearson Field or Portland International Airport to the east. The new bridge structures over the Columbia River would not include lift spans, and both of the new bridges would each be supported by six piers in the water and two piers on land.

North Portland Harbor Bridges

The existing highway structures over North Portland Harbor would not be replaced; instead, they would be retained to accommodate all mainline I-5 traffic. As discussed at the beginning of this chapter, two design options have emerged for the Hayden Island and Marine Drive interchanges. The preferred option, LPA Option A, includes local vehicular access between Marine Drive and Hayden Island on an arterial bridge. LPA Option B does not have arterial lanes on the light rail/multi-use path bridge, but instead provides direct access between Marine Drive and the island with collector-distributor lanes on the two new bridges that would be built adjacent to I-5.

LPA Option A: Four new, narrower parallel structures would be built across the waterway, three on the west side and one on the east side of the existing North Portland Harbor bridges. Three of

the new structures would carry on- and off-ramps to mainline I-5. Two structures west of the existing bridges would carry traffic merging onto or exiting off of I-5 southbound. The new structure on the east side of I-5 would serve as an on-ramp for traffic merging onto I-5 northbound.

The fourth new structure would be built slightly farther west and would include a two-lane arterial bridge for local traffic to and from Hayden Island, light rail transit, and a multi-use path for pedestrians and bicyclists. All of the new structures would have at least as much vertical clearance over the river as the existing North Portland Harbor bridges.

LPA Option B: This option would build the same number of structures over North Portland Harbor as Option A, although the locations and functions on those bridges would differ, as described below. The existing bridge over North Portland Harbor would be widened and would receive seismic upgrades.

LPA Option B does not have arterial lanes on the light rail/multi-use path bridge. Direct access between Marine Drive and the island would be provided with collector-distributor lanes. The structures adjacent to the highway bridge would carry traffic merging onto or exiting off of mainline I-5 between the Marine Drive and Hayden Island interchanges.

1.2.2.2 Interchange Improvements

The LPA includes improvements to seven interchanges along a 5-mile segment of I-5 between Victory Boulevard in Portland and SR 500 in Vancouver. These improvements include some reconfiguration of adjacent local streets to complement the new interchange designs, as well as new facilities for bicyclists and pedestrians along this corridor.

Victory Boulevard Interchange

The southern extent of the I-5 project improvements would be two ramps associated with the Victory Boulevard interchange in Portland. The Marine Drive to I-5 southbound on-ramp would be braided over the I-5 southbound to the Victory Boulevard/Denver Avenue off-ramp. The other ramp improvement would lengthen the merge distance for northbound traffic entering I-5 from Denver Avenue. The current merging ramp would be extended to become an add/drop (auxiliary) lane which would continue across the river crossing.

Potential phased construction option: The aforementioned southbound ramp improvements to the Victory Boulevard interchange may not be included with the CRC project. Instead, the existing connections between I-5 southbound and Victory Boulevard could be retained. The braided ramp connection could be constructed separately in the future as funding becomes available.

Marine Drive Interchange

All movements within this interchange would be reconfigured to reduce congestion for motorists entering and exiting I-5 at this location. The interchange configuration would be a single-point urban interchange (SPUI) with a flyover ramp serving the east to north movement. With this configuration, three legs of the interchange would converge at a point on Marine Drive, over the I-5 mainline. This configuration would allow the highest volume movements to move freely without being impeded by stop signs or traffic lights.

The Marine Drive eastbound to I-5 northbound flyover ramp would provide motorists with access to I-5 northbound without stopping. Motorists from Marine Drive eastbound would access I-5

southbound without stopping. Motorists traveling on Martin Luther King Jr. Boulevard westbound to I-5 northbound would access I-5 without stopping at the intersection.

The new interchange configuration changes the westbound Marine Drive and westbound Vancouver Way connections to Martin Luther King Jr. Boulevard and to northbound I-5. These two streets would access westbound Martin Luther King Jr. Boulevard farther east. Martin Luther King Jr. Boulevard would have a new direct connection to I-5 northbound.

In the new configuration, the connections from Vancouver Way and Marine Drive would be served, improving the existing connection to Martin Luther King Jr. Boulevard east of the interchange. The improvements to this connection would allow traffic to turn right from Vancouver Way and accelerate onto Martin Luther King Jr. Boulevard. On the south side of Martin Luther King Jr. Boulevard, the existing loop connection would be replaced with a new connection farther east.

A new multi-use path would extend from the Bridgeton neighborhood to the existing Expo Center light rail station and from the station to Hayden Island along the new light rail line over North Portland Harbor.

LPA Option A: Local traffic between Martin Luther King Jr. Boulevard/Marine Drive and Hayden Island would travel via an arterial bridge over North Portland Harbor. There would be some variation in the alignment of local streets in the area of the interchange between Option A and Option B. The most prominent differences are the alignments of Vancouver Way and Union Court.

LPA Option B: With this design option, there would be no arterial traffic lanes on the light rail/multi-use path bridge over North Portland Harbor. Instead, vehicles traveling between Martin Luther King Jr. Boulevard/ Marine Drive and Hayden Island would travel on the collector-distributor bridges that would parallel each side of I-5 over North Portland Harbor. Traffic would not need to merge onto mainline I-5 to travel between the island and Martin Luther King Jr. Boulevard/Marine Drive.

Potential phased construction option: The aforementioned flyover ramp could be deferred and not constructed as part of the CRC project. In this case, rather than providing a direct eastbound Marine Drive to I-5 northbound connection by a flyover ramp, the project improvements to the interchange would instead provide this connection through the signal-controlled SPUI. The flyover ramp could be constructed separately in the future as funding becomes available.

Hayden Island Interchange

All movements for this interchange would be reconfigured. The new configuration would be a split tight diamond interchange. Ramps parallel to the highway would be built, lengthening the ramps and improving merging speeds. Improvements to Jantzen Drive and Hayden Island Drive would include additional through, left-turn, and right-turn lanes. A new local road, Tomahawk Island Drive, would travel east-west through the middle of Hayden Island and under the I-5 interchange, improving connectivity across I-5 on the island. Additionally, a new multi-use path would be provided along the elevated light rail line on the west side of the Hayden Island interchange.

LPA Option A: A proposed arterial bridge with two lanes of traffic, one in each direction, would allow vehicles to travel between Martin Luther King Jr. Boulevard/ Marine Drive and Hayden Island without accessing I-5.

LPA Option B: With this design option there would be no arterial traffic lanes on the light rail/multi-use path bridge over North Portland Harbor. Instead, vehicles traveling between Martin Luther King Jr. Boulevard/Marine Drive and Hayden Island would travel on the collector-distributor bridges that parallel each side of I-5 over North Portland Harbor.

SR 14 Interchange

The function of this interchange would remain largely the same. Direct connections between I-5 and SR 14 would be rebuilt. Access to and from downtown Vancouver would be provided as it is today, but the connection points would be relocated. Downtown Vancouver I-5 access to and from the south would be at C Street rather than Washington Street, while downtown connections to and from SR 14 would be made by way of Columbia Street at 4th Street.

The multi-use bicycle and pedestrian path in the northbound (eastern) I-5 bridge would exit the structure at the SR 14 interchange, and then loop down to connect into Columbia Way.

Mill Plain Interchange

This interchange would be reconfigured into a SPUI. The existing "diamond" configuration requires two traffic signals to move vehicles through the interchange. The SPUI would use one efficient intersection and allow opposing left turns simultaneously. This would improve the capacity of the interchange by reducing delay for traffic entering or exiting the highway.

This interchange would also receive several improvements for bicyclists and pedestrians. These include bike lanes and sidewalks, clear delineation and signing, short perpendicular crossings at the ramp terminals, and ramp orientations that would make pedestrians highly visible.

Fourth Plain Interchange

The improvements to this interchange would be made to better accommodate freight mobility and access to the new park and ride at Clark College. Northbound I-5 traffic exiting to Fourth Plain would continue to use the off-ramp just north of the SR 14 interchange. The southbound I-5 exit to Fourth Plain would be braided with the SR 500 connection to I-5, which would eliminate the non-standard weave between the SR 500 connection and the off-ramp to Fourth Plain as well as the westbound SR 500 to Fourth Plain Boulevard connection.

Additionally, several improvements would be made to provide better bicycle and pedestrian mobility and accessibility, including bike lanes, neighborhood connections, and access to the park and ride.

SR 500 Interchange

Improvements would be made to the SR 500 interchange to add direct connections to and from I-5. On- and off-ramps would be built to directly connect SR 500 and I-5 to and from the north, connections that are currently made by way of 39th Street. I-5 southbound traffic would connect to SR 500 via a new tunnel underneath I-5. SR 500 eastbound traffic would connect to I-5 northbound on a new on-ramp. The 39th Street connections with I-5 to and from the north would be eliminated. Travelers would instead use the connections at Main Street to connect to and from 39th Street.

Additionally, several improvements would be made to provide better bicycle and pedestrian mobility and accessibility, including sidewalks on both sides of 39th Street, bike lanes, and neighborhood connections.

Potential phased construction option: The northern half of the existing SR 500 interchange would be retained, rather than building new connections between I-5 southbound to SR 500 eastbound and from SR 500 westbound to I-5 northbound. The ramps connecting SR 500 and I-5 to and from the north could be constructed separately in the future as funding becomes available.

1.2.2.3 Transit

The primary transit element of the LPA is a 2.9-mile extension of the current Metropolitan Area Express (MAX) Yellow Line light rail from the Expo Center in North Portland, where it currently ends, to Clark College in Vancouver. The transit element would not differ between LPA and LPA with highway phasing. To accommodate and complement this major addition to the region's transit system, a variety of additional improvements are also included in the LPA:

- Three park and ride facilities in Vancouver near the new light rail stations.
- Expansion of Tri-County Metropolitan Transportation District's (TriMet's) Ruby Junction light rail maintenance base in Gresham, Oregon.
- Changes to C-TRAN local bus routes.
- Upgrades to the existing light rail crossing over the Willamette River via the Steel Bridge.

Operating Characteristics

Nineteen new light rail vehicles (LRV) would be purchased as part of the CRC project to operate this extension of the MAX Yellow Line. These vehicles would be similar to those currently used by TriMet's MAX system. With the LPA, LRVs in the new guideway and in the existing Yellow Line alignment are planned to operate with 7.5-minute headways during the "peak of the peak" (the two-hour period within the 4-hour morning and afternoon/evening peak periods where demand for transit is the highest) and 15-minute headways during off-peak periods.

Light Rail Alignment and Stations

Oregon Light Rail Alignment and Station

A two-way light rail alignment for northbound and southbound trains would be constructed to extend from the existing Expo Center MAX station over North Portland Harbor to Hayden Island. Immediately north of the Expo Center, the alignment would curve eastward toward I-5, pass beneath Marine Drive, then rise over a flood wall onto a light rail/multi-use path bridge to cross North Portland Harbor. The two-way guideway over Hayden Island would be elevated at approximately the height of the rebuilt mainline of I-5, as would a new station immediately west of I-5. The alignment would extend northward on Hayden Island along the western edge of I-5, until it transitions into the hollow support structure of the new western bridge over the Columbia River.

Downtown Vancouver Light Rail Alignment and Stations

After crossing the Columbia River, the light rail alignment would curve slightly west off of the highway bridge and onto its own smaller structure over the Burlington Northern Santa Fe (BNSF) rail line. The double-track guideway would descend on structure and touch down on Washington Street south of 5th Street, continuing north on Washington Street to 7th Street. The elevation of 5th Street would be raised to allow for an at-grade crossing of the tracks on Washington Street. Between 5th and 7th Streets, the two-way guideway would run down the center of the street.

Traffic would not be allowed on Washington between 5th and 6th Streets and would be two-way between 6th and 7th Streets. There would be a station on each side of the street on Washington between 5th and 6th Streets.

At 7th Street, the light rail alignment would form a couplet. The single-track northbound guideway would turn east for two blocks, then turn north onto Broadway Street, while the single-track southbound guideway would continue on Washington Street. Seventh Street will be converted to one-way traffic eastbound between Washington and Broadway with light rail operating on the north side of 7th Street. This couplet would extend north to 17th Street, where the two guideways would join and turn east.

The light rail guideway would run on the east side of Washington Street and the west side of Broadway Street, with one-way traffic southbound on Washington Street and one-way traffic northbound on Broadway Street. On station blocks, the station platform would be on the side of the street at the sidewalk. There would be two stations on the Washington-Broadway couplet, one pair of platforms near Evergreen Boulevard, and one pair near 15th Street.

East-west Light Rail Alignment and Terminus Station

The single-track southbound guideway would run in the center of 17th Street between Washington and Broadway Streets. At Broadway Street, the northbound and southbound alignments of the couplet would become a two-way center-running guideway traveling east-west on 17th Street. The guideway on 17th Street would run until G Street, then connect with McLoughlin Boulevard and cross under I-5. Both alignments would end at a station east of I-5 on the western boundary of Clark College.

Park and Ride Stations

Three park and ride stations would be built in Vancouver along the light rail alignment:

- Within the block surrounded by Columbia, Washington 4th and 5th Streets, with five floors above ground that include space for retail on the first floor and 570 parking stalls.
- Between Broadway and Main Streets next to the stations between 15th and 16th Streets, with space for retail on the first floor, and four floors above ground that include 420 parking stalls.
- At Clark College, just north of the terminus station, with space for retail or C-TRAN services on the first floor, and five floors that include approximately 1,910 parking stalls.

Ruby Junction Maintenance Facility Expansion

The Ruby Junction Maintenance Facility in Gresham, Oregon, would need to be expanded to accommodate the additional LRVs associated with the CRC project. Improvements include additional storage for LRVs and other maintenance material, expansion of LRV maintenance bays, and expanded parking for additional personnel. A new operations command center would also be required, and would be located at the TriMet Center Street location in Southeast Portland.

Local Bus Route Changes

As part of the CRC project, several C-TRAN bus routes would be changed in order to better complement the new light rail system. Most of these changes would re-route bus lines to downtown Vancouver where riders could transfer to light rail. Express routes, other than those

listed below, are expected to continue service between Clark County and downtown Portland. The following table (Exhibit 1-1) shows anticipated future changes to C-TRAN bus routes.

Exhibit 1-1. Proposed C-TRAN Bus Routes Comparison

C-TRAN Bus Route	Route Changes
#4 - Fourth Plain	Route truncated in downtown Vancouver
#41 - Camas / Washougal Limited	Route truncated in downtown Vancouver
#44 - Fourth Plain Limited	Route truncated in downtown Vancouver
#47 - Battle Ground Limited	Route truncated in downtown Vancouver
#105 - I-5 Express	Route truncated in downtown Vancouver
#105S - I-5 Express Shortline	Route eliminated in LPA (The No-Build runs articulated buses between downtown Portland and downtown Vancouver on this route)

Steel Bridge Improvements

Currently, all light rail lines within the regional TriMet MAX system cross over the Willamette River via the Steel Bridge. By 2030, the number of LRVs that cross the Steel Bridge during the 4-hour PM peak period would increase from 152 to 176. To accommodate these additional trains, the project would retrofit the existing rails on the Steel Bridge to increase the allowed light rail speed over the bridge from 10 to 15 mph. To accomplish this, additional work along the Steel Bridge lift spans would be needed.

1.2.2.4 Tolling

Tolling cars and trucks that use the I-5 river crossing is proposed as a method to help fund the CRC project and to encourage the use of alternative modes of transportation. The authority to toll the I-5 crossing is set by federal and state laws. Federal statutes permit a toll-free bridge on an interstate highway to be converted to a tolled facility following the reconstruction or replacement of the bridge. Prior to imposing tolls on I-5, Washington and Oregon Departments of Transportation (WSDOT and ODOT) would have to enter into a toll agreement with U.S. Department of Transportation (DOT). Recently passed state legislation in Washington permits WSDOT to toll I-5 provided that the tolling of the facility is first authorized by the Washington legislature. Once authorized by the legislature, the Washington Transportation Commission (WTC) has the authority to set the toll rates. In Oregon, the Oregon Transportation Commission (OTC) has the authority to toll a facility and to set the toll rate. It is anticipated that prior to tolling I-5, ODOT and WSDOT would enter into a bi-state tolling agreement to establish a cooperative process for setting toll rates and guiding the use of toll revenues.

Tolls would be collected using an electronic toll collection system: toll collection booths would not be required. Instead, motorists could obtain a transponder that would automatically bill the vehicle owner each time the vehicle crossed the bridge, while cars without transponders would be tolled by a license-plate recognition system that would bill the address of the owner registered to that license plate.

The LPA proposes to apply a variable toll on vehicles using the I-5 crossing. Tolls would vary by time of day, with higher rates during peak travel periods and lower rates during off-peak periods. Medium and heavy trucks would be charged a higher toll than passenger vehicles. The traffic-related impact analysis in this FEIS is based on toll rates that, for passenger cars with transponders, would range from \$1.00 during the off-peak to \$2.00 during the peak travel times (in 2006 dollars).

1.2.2.5 Transportation System and Demand Management Measures

Many well-coordinated transportation demand management (TDM) and transportation system management (TSM) programs are already in place in the Portland-Vancouver Metropolitan region and supported by agencies and adopted plans. In most cases, the impetus for the programs is from state-mandated programs: Oregon's Employee Commute Options (ECO) rule and Washington's Commute Trip Reduction (CTR) law.

The physical and operational elements of the CRC project provide the greatest TDM opportunities by promoting other modes to fulfill more of the travel needs in the project corridor. These include:

- Major new light rail line in exclusive right-of-way, as well as express bus and feeder routes;
- Modern bicycle and pedestrian facilities that accommodate more bicyclists and pedestrians, and improve connectivity, safety, and travel time;
- Park and ride lots and garages; and
- A variable toll on the highway crossing.

In addition to these fundamental elements of the project, facilities and equipment would be implemented that could help existing or expanded TSM programs maximize capacity and efficiency of the system. These include:

- Replacement or expanded variable message signs or other traveler information systems in the CRC project area;
- Expanded incident response capabilities;
- Queue jumps or bypass lanes for transit vehicles where multi-lane approaches are provided at ramp signals for entrance ramps;
- Expanded traveler information systems with additional traffic monitoring equipment and cameras, and
- Active traffic management.

1.2.3 LPA Construction

Construction of bridges over the Columbia River is the most substantial element of the project, and this element sets the sequencing for other project components. The main river crossing and immediately adjacent highway improvement elements would account for the majority of the construction activity necessary to complete this project.

1.2.3.1 Construction Activities Sequence and Duration

The following table (Exhibit 1-2) displays the expected duration and major details of each element of the project. Due to construction sequencing requirements, the timeline to complete the initial phase of the LPA with highway phasing is the same as the full LPA.

Exhibit 1-2. Construction Activities and Estimated Duration

Element	Estimated Duration	Details
Columbia River bridges	4 years	 Construction is likely to begin with the bridges. General sequence includes initial preparation, installation of foundation piles, shaft caps, pier columns, superstructure, and deck.
Hayden Island and SR 14 interchanges	1.5 - 4 years for each interchange	 Each interchange must be partially constructed before any traffic can be transferred to the new structure. Each interchange needs to be completed at the same time.
Marine Drive interchange	3 years	 Construction would need to be coordinated with construction of the southbound lanes coming from Vancouver.
Demolition of the existing bridges	1.5 years	 Demolition of the existing bridges can begin only after traffic is rerouted to the new bridges.
Three interchanges north of SR 14	4 years for all three	 Construction of these interchanges could be independent from each other or from the southern half of the project. More aggressive and costly staging could shorten this timeframe.
Light rail	4 years	 The river crossing for the light rail would be built with the bridges. Any bridge structure work would be separate from the actual light rail construction activities and must be completed first.
Total Construction Timeline	6.3 years	 Funding, as well as contractor schedules, regulatory restrictions on in-water work, weather, materials, and equipment, could all influence construction duration. This is also the same time required to complete the smallest usable segment of roadway – Hayden Island through SR 14 interchanges.

1.2.3.2 Major Staging Sites and Casting Yards

Staging of equipment and materials would occur in many areas along the project corridor throughout construction, generally within existing or newly purchased right-of-way or on nearby vacant parcels. However, at least one large site would be required for construction offices, to stage the larger equipment such as cranes, and to store materials such as rebar and aggregate. Suitable sites must be large and open to provide for heavy machinery and material storage, must have waterfront access for barges (either a slip or a dock capable of handling heavy equipment and material) to convey material to the construction zone, and must have roadway or rail access for landside transportation of materials by truck or train.

Three sites have been identified as possible major staging areas:

- 1. Port of Vancouver (Parcel 1A) site in Vancouver: This 52-acre site is located along SR 501 and near the Port of Vancouver's Terminal 3 North facility.
- 2. Red Lion at the Quay hotel site in Vancouver: This site would be partially acquired for construction of the Columbia River crossing, which would require the demolition of the building on this site, leaving approximately 2.6 acres for possible staging.

3. Vacant Thunderbird hotel site on Hayden Island: This 5.6-acre site is much like the Red Lion hotel site in that a large portion of the parcel is already required for new right-of-way necessary for the LPA.

A casting/staging yard could be required for construction of the overwater bridges if a precast concrete segmental bridge design is used. A casting yard would require access to the river for barges, including either a slip or a dock capable of handling heavy equipment and material; a large area suitable for a concrete batch plant and associated heavy machinery and equipment; and access to a highway and/or railway for delivery of materials.

Two sites have been identified as possible casting/staging yards:

- 1. Port of Vancouver Alcoa/Evergreen West site: This 95-acre site was previously home to an aluminum factory and is currently undergoing environmental remediation, which should be completed before construction of the CRC project begins (2012). The western portion of this site is best suited for a casting yard.
- 2. Sundial site: This 50-acre site is located between Fairview and Troutdale, just north of the Troutdale Airport, and has direct access to the Columbia River. There is an existing barge slip at this location that would not have to undergo substantial improvements.

1.2.4 The No-Build Alternative

The No-Build Alternative illustrates how transportation and environmental conditions would likely change by the year 2030 if the CRC project is not built. This alternative makes the same assumptions as the build alternatives regarding population and employment growth through 2030, and also assumes that the same transportation and land use projects in the region would occur as planned. The No-Build Alternative also includes several major land use changes that are planned within the project area, such as the Riverwest development just south of Evergreen Boulevard and west of I-5, the Columbia West Renaissance project along the western waterfront in downtown Vancouver, and redevelopment of the Jantzen Beach shopping center on Hayden Island. All traffic and transit projects within or near the CRC project area that are anticipated to be built by 2030 separately from this project are included in the No-Build and build alternatives. Additionally, the No-Build Alternative assumes bridge repair and continuing maintenance costs to the existing bridge that are not anticipated with the replacement bridge option.

1.3 Long-term Effects

1.3.1 Aquatic Resources

Long-term impacts to ecosystem resources as a result of the CRC project are likely for aquatic resources, including federally listed fish species and riverine habitat in the Columbia River and North Portland Harbor. Long-term effects include those related to direct effects to species and habitat.

Indirect effects are those caused by the action and may occur later in time or are farther removed in distance, but are still reasonably foreseeable. Two elements of the CRC project are likely to result in indirect effects. Increases to impervious surface area within drainages, and the consequent increase in stormwater runoff volumes and pollutant loads, would cause ongoing effects to the project area water bodies. Long-term net benefits to water quality may result from project improvements to stormwater treatment in the project area. Increased capacity of the

highway system and light rail transit network could potentially lead to changes in land use or traffic patterns for years to come.

1.3.1.1 Long-term Effects to Species and Habitat

Long-term effects to aquatic species could include increased risk of predation to juvenile salmonids due to in-water shading and flow refuge associated with bridge structures. Bridge piers constructed in the channel may provide refugia via shade and protection from the river current for piscivorous fish species that could feed on out-migrating juvenile salmonids, thereby impacting overall juvenile survival rates. Holding and rearing behavior of salmonids as they pass through the project could also be affected by these localized changes in habitat. See section 4.1.3.1 for a detailed discussion of predation risk.

Long-term effects to listed salmonids would be consistent with current conditions with respect to the presence of human-made structures in a highly modified urban setting; that is, the continued presence of bridge pier elements in the river and a major highway system over the river. Bridge piers in the river, particularly in near-shore and shallow-water areas, can have long-term impacts to aquatic habitat and channel dynamics as a result of sediment deposition and alteration of flow patterns. The project would have permanent impacts to shallow-water habitat (water less than 20 feet deep) in the Columbia River and North Portland Harbor, including the addition of in-water and overwater bridge elements and the removal of existing in-water and overwater structures. Permanent impacts to deep-water habitat in the Columbia River would include a net physical gain of habitat area (due to removal of existing bridge piers), and an increase in overwater coverage. See section 4.1 for a detailed discussion of effects to species and habitat.

Due to the depth of the water and active riverbed in the Columbia River, benthic organisms (e.g., aquatic macroinvertebrates) are not likely to be present at the majority of the pier construction locations. Two piers would be located in shallow water (20 feet or less)/near-shore areas, where habitat for benthic organisms would be displaced by the new structures.

1.3.1.2 Long-term Effects to Water Quality

Addition of impervious surface to a watershed has the potential to affect fish by altering water quality in the receiving water bodies. Stormwater runoff flows over the roadway, picking up contaminants from impervious surfaces and delivering them to the roadside drainage system and eventually to surface water bodies (Pacific EcoRisk 2007). Sources of these contaminants include vehicles, atmospheric deposition, roadway maintenance, and pavement wear (Pacific EcoRisk 2007). Contaminants that may be present in stormwater runoff associated with highways include suspended sediments, nutrients, polycyclic aromatic hydrocarbons (PAHs), oils and grease, antifreeze from leaks, cadmium and zinc from tire wear, and copper from wear and tear of brake pads, bearings, metal plating, and engine parts.

The CRC Project is a bi-state initiative and it is important to note that the implementation of water management objectives differ significantly between Oregon and Washington State. The primary differences involve how areas that require pollutant reduction are calculated. These differences, which are described in the following paragraphs, can have an impact of the sizes of water quality facility required, especially for projects like the CRC that involve significant areas of impervious pavement.

Oregon requires runoff from the entire contributing impervious area (CIA) be treated to reduce pollutants regardless of degree to which the surfaces would contribute pollutants to runoff. Using this approach, runoff from highways would be required to be treated in the same manner as runoff

from bike-pedestrian paths. In contrast, Washington State focuses on requiring treatment for runoff from the pollutant-generating impervious surfaces (PGIS).

ODOT defines the CIA as consisting of all impervious surfaces within the strict project limits, plus impervious surface owned or operated by ODOT outside the project limits that drain to the project via direct flow or discrete conveyance. NMFS has expanded this definition to also include impervious areas that are not owned by ODOT but drain onto the project footprint.

WSDOT and Ecology define PGIS as surfaces that are considered a significant source of pollutants in stormwater runoff including:

- highways, ramps and non-vegetated shoulders;
- light rail transit guideway subject to vehicular traffic;
- streets, alleys, and driveways; and
- bus layover facilities, surface parking lots, and the top floor of parking structures.

The following types of impervious area are considered non-PGIS:

- light rail transit guideway not subject to vehicular traffic except the occasional use by emergency or maintenance vehicles (referred to as an exclusive guideway);
- light rail transit stations; and
- bicycle and pedestrian paths.

The project CIA currently contains 256 acres of impervious surface and would add a net 42 acres, resulting in a post-project net total of 298 acres. The increase in CIA would likely have effects on stormwater runoff draining from the project area into all of the project area water bodies: the Columbia River, North Portland Harbor, Columbia Slough, and Burnt Bridge Creek. It would not be expected to have effects on Fairview Creek since all impervious surfaces would be treated onsite and impervious surface within the watershed would actually decrease by 0.5 acre as a result of the expansion of the Ruby Junction maintenance facility.

The project would install stormwater treatment facilities to treat or sequester pollutants and to provide flow control (where required) before runoff enters any surface water body. The completed project will provide treatment not only for the new and rebuilt impervious surface, but also for existing impervious surface that is not currently treated. The completed project would treat more than 8 times the amount of new PGIS. The CRC project occurs within several different state and local jurisdictions, each of which has different stormwater treatment standards. The CRC project team agreed to incorporate the most restrictive water quality requirements of all these standards, as embodied in the ODOT stormwater best management practices (BMP) selection tool (ODOT 2008). Furthermore, the conceptual stormwater management design provides treatment and infiltration of the entire project CIA, to the maximum extent possible, in response to the requirements of NMFS and DEQ for the CRC project. The extent of treatment would likely result in a net benefit to water quality and water quantity in the project area water bodies during the majority of storm events. See section 4.1.1 for a detailed discussion of stormwater treatment and effects to aquatic resources.

Pollutant loading within the project corridor would be expected to decline within the Columbia River and Burnt Bridge Creek compared to pollutant loads expected under the No-Build alternative because stormwater treatment would be provided where treatment would otherwise not exist. Pollutant loads would decline in the Columbia Slough watershed with the exception of

dissolved copper. Pollutant loads of dissolved copper are projected to increase slightly (5-6 percent) in the Columbia Slough as a result of the LPA.

1.3.2 Terrestrial Resources

Potential long-term effects to peregrine falcon habitat may occur if the existing bridge is removed and structures that are currently used by this species are demolished. No other significant long-term impacts are anticipated to terrestrial resources. See Sections 4 and 5 for additional effects analysis.

1.3.3 Botanical Resources

No long-term effects are expected to botanical resources. Although trees and other vegetation may be removed within the project footprint for new permanent and temporary structures, revegetation with native plants to meet local criteria would occur within or adjacent to the project footprint. Noxious weeds would be removed in accordance with state transportation department policies.

1.3.4 Regional Resources

Long-term regional effects would be seen primarily in effects to listed fish and aquatic habitat, especially water quality. The Columbia River in the project area is a major waterway through which at least 14 salmonid stocks, as well as lamprey, sturgeon, and other native fish, pass during various portions of their life cycles. Salmonids are present in the project area during adult migration upriver to spawn, juvenile outmigration, and rearing; therefore, impacts to these species at these life stages could have substantial implications for survival and reproduction of these populations of salmonids. However, long-term impacts from project activities are likely to be consistent with existing conditions for aquatic species (e.g., the presence of a major artificial structure in the mainstem of the river). Impacts of a large bridge structure in the mainstem river could be reduced to some extent, relative to existing conditions, by several design elements. For example, the new bridge would have fewer piers in shallow-water habitat. In addition, water quality in the mainstem Columbia River and North Portland Harbor may be improved, at least in the immediate project area, through improvements to stormwater collection and treatment.

Long-term regional effects to terrestrial species and habitats would likely be consistent with existing conditions. Migratory birds would likely use the new bridge designs and the natural habitat in the project area for roosting, foraging, and potentially for nesting, similar to their use of the existing elements. Wildlife passage would be likely to remain limited in the project area due to the highly urbanized setting.

Regional traffic patterns would likely change as a result of improvements to the I-5 bridge crossing, potentially resulting in negative impacts to ecosystem resources in some areas and positive effects on water quality in other areas. These effects are addressed in the Indirect Effects Technical Report.

1.4 Temporary Effects

1.4.1 Aquatic Resources

Temporary effects to aquatic habitat and aquatic species are anticipated from in-water work. Inwater work may include removing existing bridge piers, constructing new piers, and installing and removing temporary in-water work structures. In-water work is likely to include coffer dams, barges, drilling equipment, impact pile drivers, vibratory pile drivers, and other construction vehicles in and near the water. Construction activities may cause injury or death to aquatic species.

In-water work would likely cause localized increases in underwater noise, turbidity, artificial lighting, avian predation, hydraulic shadowing, and shading. Specific effects could include potential sub-lethal injury due to hydroacoustic impacts associated with pile driving and fish handling; increased risk of predation due to in-water shading during construction; and potential mortality associated with hydroacoustic impacts and fish handling. Effects to habitat include turbidity, loss of shallow-water habitat, and obstructions to migration.

Water quality could be adversely impacted by accidental contaminant spills (e.g., barge and heavy equipment fuel, oil), erosion, turbidity, and sediment. Current riparian vegetative structure provides negligible benefits for regulating water temperature in the Columbia River and North Portland Harbor. Only small amounts of riparian vegetation may be removed during the project and would not be expected to affect aquatic habitats. See Section 5 for additional effects analysis.

1.4.2 Terrestrial Resources

Temporary impacts to terrestrial resources, specifically to migratory birds and peregrine falcons, are likely to occur. Modifications to migratory bird habitats are likely because existing vegetation, as well as the bridge structures themselves (on which birds may roost or nest), are expected to be removed. Vegetation, including potential nesting habitat such as trees and shrubs, would be replanted and would replace the temporarily impacted habitat. Construction noise may also disturb or prevent nesting.

Temporary effects to terrestrial species are anticipated from construction noise and impacts to vegetation. Construction activity causing noise disturbance could result in reduced nesting success for migratory birds. Trees, shrubs, and other vegetation serving as cover, nesting, roosting, and perching habitat may be removed during construction. Such vegetation removal could also impact terrestrial wildlife using such habitat structure for cover, feeding, breeding, and dispersal. See Section 5 for additional effects analysis.

1.4.3 Botanical Resources

Temporary impacts to vegetation in the project area may result from grading, staging, realignment of the main bridge structure, and other project-related activities. Disturbed vegetated areas would be replanted according to site restoration plans. No effects to sensitive plant species are expected because no sensitive plants are known to occur within the project area. Noxious weeds would be removed in accordance with state transportation department policies.

1.5 Proposed Mitigation

Mitigation for impacts to aquatic, terrestrial, and botanical resources include BMPs, conservation measures, and avoidance and minimization measures.

The LPA would impact fish species by the presence of large piers in the river that could provide habitat for piscivorous fish, and that could alter stream flow. In addition, riparian fringe habitat may be altered. Mitigation measures to address these impacts include impact avoidance and impact minimization. Revegetation of riparian areas and limited use of riprap would be employed to limit negative long-term effects. Long-term impacts to terrestrial resources, such as migratory

birds, are relatively minimal and would not require extensive mitigation. Refer to Section 6 for a more detailed discussion of proposed mitigation approaches.

During construction, the LPA would impact fish species through in-water work that could result in increased turbidity, hydroacoustic impacts, temporary localized dewatering, and potential contaminant spills. Mitigation measures to address these impacts include impact avoidance and impact minimization. Impact avoidance has been addressed through project design alternatives that were considered but not advanced due to impacts to ecosystem and other resources. Certain design alternatives have also been modified to reduce impacts to resources. Impact minimization would be addressed through implementing BMPs (e.g., sediment and erosion control, no-work zones, appropriate flagging and fencing), monitoring project activities, timing restrictions for inwater work to avoid impacts to fish runs, using cofferdams around select in-water work sites, and using bubble curtains around impact pile driving that may cause adverse impacts from noise.

The LPA would impact terrestrial resources, such as migratory birds and species of interest (SOI) (defined for the purposes of this document as species which are not protected by federal statute but which are locally rare or have special habitat requirements), through noise impacts and removal or degradation of habitat. Mitigation measures to address these impacts include impact avoidance and impact minimization. For example, to avoid direct impacts to active peregrine falcon and other migratory bird nests, demolition of existing structures would be scheduled outside of nesting seasons, and/or management plans would be developed to provide guidance on ways to avoid violation of the Migratory Bird Treaty Act (MBTA).

Stormwater collections and treatment would occur to treat for metals, biosolids, and other contaminants. Methods used would be more effective and efficient than current treatment, and should result in improved water quality in the project area.

2. Methods

2.1 Introduction

Methods used to collect data and analyze effects included:

- Collecting a list of federally listed species, potential species of interest¹ (SOI), and their habitats from local, state, and federal resource and management agencies.
- Determining species life history and habitat requirements.
- Conducting field surveys with accepted protocols during appropriate seasons.
- Examining existing Geographic Information System (GIS) data layers.
- Discussing potential impacts to resources with species experts, local resource managers, and agency biologists.

Refer also to the Ecosystems Methods and Data Report for additional details.

2.2 Project Area

The project area is defined as all areas that would be directly impacted by the project, including the footprint of the permanent and temporary structures, the widened highway segments, the new interchanges, city street realignments, associated road shoulder excavation and fill areas, stormwater facilities, wetland mitigation sites, and staging and access areas, including those in the Columbia River and North Portland Harbor where work would occur from barges and temporary structures.

Along the I-5 corridor, the project area extends 5 miles from north to south, beginning at the I-5/SR 500 interchange in Vancouver, Washington and extending to the I-5/Victory Boulevard in Portland, Oregon (Exhibit 2-1). At its northern end, the project area extends west into downtown Vancouver and east near Clark College to include high-capacity transit alignments, transit stations, park-and-ride locations, and city road improvements included as part of this project. Heading south along the existing overwater bridge alignments, the project area extends 0.25 mile on either side of the existing bridges to include the new river and harbor bridges, as well as the areas where construction and demolition activities would occur. Continuing south, the project extends east to include city road improvements along Victory Boulevard.

The project area also includes those portions of the Columbia River and North Portland Harbor that would be affected by underwater noise. In the Columbia River and North Portland Harbor, hydroacoustic impacts from impact pile driving are the farthest reaching extent of project aquatic impacts. Due to the curvature of the river and islands present, underwater noise from impact pile driving is expected to reach land before it reaches ambient levels. Noise from impact pile driving is not expected to extend beyond Sauvie Island, approximately 5.5 miles downstream and Lady Island, 12.5 miles upstream. This distance encompasses the Columbia River from approximately

¹ SOI are not a specific category of governmental or NGO-designated species, but are referred to here as those identified through tribal, local, state, and federal coordination as those species that are locally rare and have special habitat considerations.

river mile (RM) 101 to 118. Within North Portland Harbor, underwater noise is expected to extend 3.5 miles downstream and 1.9 miles upstream.

The project area includes potential staging and casting yards at the Port of Vancouver, Alcoa/Evergreen, Sundial, Red Lion at the Quay, and Thunderbird Hotel staging sites.

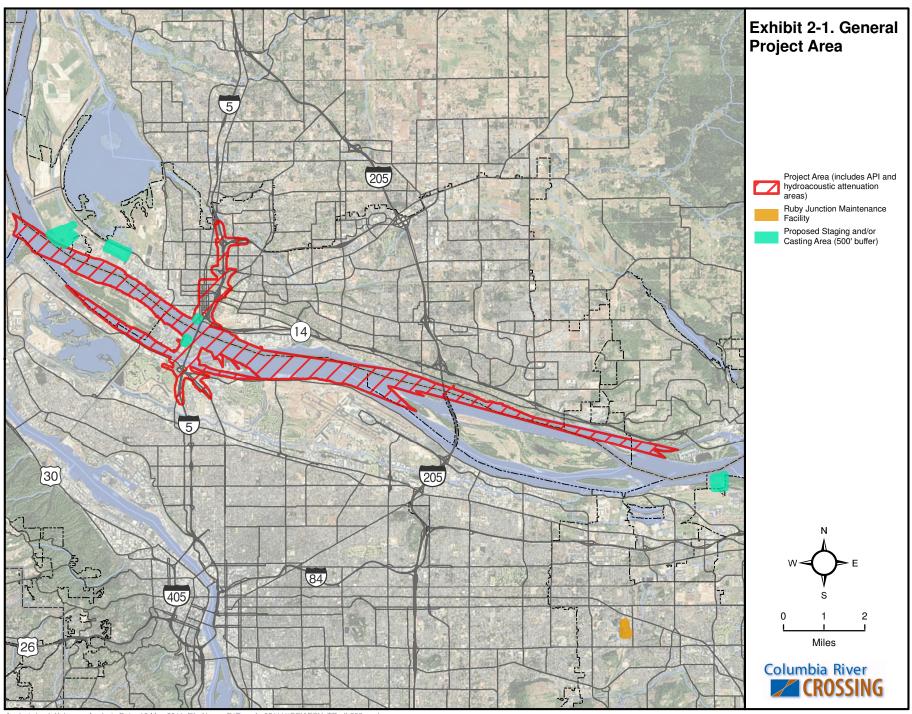
In downtown Portland, the project area includes the upper deck of the Steel Bridge where minor rail improvements would take place.

In Gresham, Oregon, the project area includes a 10.5-acre expansion of the Ruby Junction Maintenance Facility. This includes all associated cut and fill slopes and stormwater treatment facilities.

2.3 Effects Guidelines

Local, state, and federal agencies provide guidance in determining impacts to ecosystem resources. The impact assessment considered effects to species and habitats, taking into consideration federal and state protected status, impacts to species' ecology and critical life stages (e.g., breeding), primary constituent elements (PCEs) where applicable (e.g., critical habitat), and other relevant factors. The following factors were considered in determining the type and degree of impacts:

- Effects to listed species analyzed in Section 7 of the Endangered Species Act (ESA) consultations conducted with the United States Fish and Wildlife Service (USFWS) and NMFS; consultation has been completed with FHWA, FTA, ODOT, and WSDOT.
- Effects to Essential Fish Habitat (EFH) under the Magnuson-Stevens Fishery Conservation Management Act (MSFCMA).
- Effects to threatened or endangered species recovery potential as described in a USFWS or NMFS recovery plan or other guidance if a recovery plan is not available.
- Extent of impacts to existing wildlife corridors (which could be either further degraded or improved by this project).
- Impacts to fish passage for all life stages of listed and non-listed native fish (e.g., physical barriers).
- Effects to high quality habitat, such as fragmentation, degradation, or impairment that would reduce its capacity to provide vital functions for species; "high quality" habitat is defined in Oregon Department of Fish and Wildlife's (ODFW) Habitat Mitigation Policy and Washington Department of Fish and Wildlife's (WDFW) Priority Habitats.
- Effects to migratory birds, as defined under the MBTA, such as take of active nests and/or eggs.
- Effects to marine mammals, as defined under the Marine Mammal Protection Act (MMPA), such as harassment or injury.
- Effects to species under state regulatory statutes governing "take," such as the Oregon and Washington statutory authorities protecting endangered, threatened, and sensitive species.
- Effects to state and locally protected habitats (e.g., impacts that would remove or degrade habitats to the point that they can no longer provide vital functions for the species dependent on these habitats).



Analysis by J. Koloszar; Analysis Date: 16 May 2011; File Name: F:\Transfer051111\ESY\ESY_TR_JL238.mxd

2.4 Data Collection Methods

The project team conducted field reviews of SOI and aquatic, riparian, and terrestrial habitat features and conditions within the project area. Existing data, including previously prepared environmental reviews, were also gathered and incorporated into the analysis.

The following process was used to collect fish, wildlife, and botanical resource data:

- Collected a list of potential SOI and their habitats. These data were obtained from the Oregon Natural Heritage Information Center (ORNHIC), USFWS, NMFS, WDFW, and the Washington Department of Natural Resources, Natural Heritage Program (WDNR-NHP).
 - Contacted federal, state, and local agencies, and local biologists and experts.
 - Examined studies, plans, and reports prepared by local, state, and federal agencies
 and private organizations for information on species and habitats that may occur
 within the project area. These studies included the ODOT Peregrine Falcon
 Management Plan 2002 through 2007, annual peregrine falcon monitoring reports for
 the Portland metropolitan area, and the Lower Columbia Salmon Recovery and Fish
 and Wildlife Subbasin Plan.
- 2. Determined SOI habitat requirements.
 - Examined studies, plans, and reports and consulted with local biologists and federal, state, and local agencies.
 - Determined if critical habitat has been designated for listed species potentially found within the project area. Examined PCEs for species with designated critical habitat.
- 3. Determined potential habitat types and their associated species.
 - Obtained aerial photography to identify habitat types.
 - Obtained GIS maps of habitats, documented species locations, locally protected zones, critical habitats, and other ecological features. Such resource classifications include EFH (NMFS), regionally significant habitat (Metro)², essential salmonid habitat (ESH) (Oregon Department of State Lands [DSL]), priority habitats (WDFW), critical area ordinances (City of Vancouver) and environmental zones (City of Portland).
- 4. Conducted field reconnaissance in the appropriate season(s) to assess the presence of listed botanical species and all species' associated habitats within the project area and, if present, the role the habitats play in the species' life histories.
 - Ground-truthed habitat types and boundaries. Quantified habitat types within the project area based on GIS data.
 - Used Johnson and O'Neil's (2001) species/habitat matrix to determine the species most likely to be present in these habitats.

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² Metro is the directly elected regional government that serves the residents of Clackamas, Multnomah and Washington counties, and the 25 cities in the Portland, Oregon metropolitan area.

- Determined SOI habitat use within the project area and identified wildlife passage opportunities.
- Conducted rare plant surveys using the intuitive controlled method (BLM 1998).
 Conducted noxious weed surveys and mapped results based on Oregon Department of Agriculture (ODA) and Washington Noxious Weed Control Board (WNWCB) status.
- Inspected bridges for bridge-nesting species and bats, and identified potential
 migratory bird habitat. Visual inspections for these species were conducted during
 nesting seasons.
- 5. Characterized aquatic and terrestrial habitats found during field surveys for features important to fish, wildlife, and plants. All species seen during field surveys were recorded.
 - Aquatic characteristics of interest included water quality, substrate composition, bank stability, channel condition, fish passage, bathymetric characteristics, and riparian conditions. Streams were evaluated for their potential to support fish and other aquatic resources. These characteristics were evaluated qualitatively (e.g., visual observation) during the field survey, and supported by technical reports from appropriate agencies (e.g., Ecology, DEQ, WDFW, ODFW).
 - Riparian corridors were surveyed for fish and wildlife habitat elements at the I-5 crossings of the Columbia River, North Portland Harbor, and Columbia Slough. Burnt Bridge Creek was surveyed where it runs parallel to I-5 at the northern boundary of the project area. Surveyed habitat elements include vegetation type and density, stream characteristics, and piers, footings, riprap, and other structures below the ordinary high water line (OHW).
 - Terrestrial characteristics of interest included opportunities for wildlife passage, habitat distribution, structure, and composition, and habitat fragmentation or connectivity.
- 6. Compiled lists and maps of observed SOI, habitats, protected habitats, rare plants, and noxious weeds.
- 7. Analyzed data to determine potential project impacts on ecosystem resources.
 - Used agency-approved documents to determine the potential impacts from proposed alternatives on ecosystem resources.
 - Determined potential impacts to listed species and designated critical habitat.
 - Identified other resources, such as SOI or protected habitats, which might be impacted.
 - Identified habitats that provide connectivity at a landscape scale.
- 8. Conducted windshield surveys for habitats classified as non-urban based on the Johnson and O'Neil's (2001) species/habitat matrix. Special consideration was given to habitats that provide connectivity within the project area. Used species/habitat matrix to determine the species most likely to be present in habitats identified from existing data.
- 9. Compiled a list of observed habitats and potential SOI, rare plants, and noxious weeds.
- 10. Analyzed data to determine the potential for indirect impacts to ecosystem resources.

- Determined potential indirect impacts to listed species and designated critical habitat.
- Identified other resources, such as SOI or protected habitats that might be indirectly impacted.
- Identified habitats that provide connectivity at a landscape scale.

2.5 Analysis Methods

Potential cumulative effects from this project are evaluated in the Cumulative Effects Technical Report. Please refer to this report for an evaluation of possible cumulative effects.

2.5.1 Aquatic Resource Impacts

The following process was used to determine short- and long-term operational impacts on aquatic resources:

- Evaluated impacts to fish passage by comparing structural designs of the various alternatives.
- Used maps of protected habitats to determine sensitive areas that may be impacted by the project and to quantify the impact area relative to existing habitat.
- Evaluated and quantified the potential for effects to critical habitat, suitable habitat, or "take" of listed fish.

2.5.2 Terrestrial Resource Impacts

The following process was used to determine short- and long-term operational impacts on terrestrial resources, including botanical resources:

- Evaluated and quantified the potential for destruction or adverse modification of critical habitat, suitable habitat, or "take" of listed wildlife and plants.
- Evaluated and quantified impacts to species and resources not listed under the ESA based on the amount of habitat modification, destruction, or increased levels of disturbance from project operation.
- Evaluated and quantified impacts to wildlife passage based on changes to existing wildlife corridors or fragmentation of existing habitat.
- Used maps of protected habitats to determine sensitive areas that may be impacted by the project and to quantify the impact area relative to undisturbed habitat.

2.5.3 Species of Interest Impacts

The following process was used to determine short- and long-term operational impacts on special-status species:

- Evaluated the potential for adverse effects to listed species, relative to their survival and recovery, under the federal ESA.
- Used maps of special-status species locations to determine habitats that may be impacted by the project and to quantify the impact area relative to undisturbed habitat.

In addition, local, state, and federal biologists were interviewed and beneficial impacts were identified and evaluated.

2.6 Mitigation Measures Approach

Bi-state coordination is occurring to best mitigate for impacts to ecosystem resources. The intent is to provide mitigation measures that are consistent with the mitigation policies of local, state, and federal governments. The mitigation measures approach was guided by the following actions:

- Avoiding impact through design modification or by not taking a certain action or parts of the action.
- Identifying and evaluating ways to minimize impacts to ecosystem resources.
- Researching and identifying BMPs to minimize and avoid impacts.
- Discussing BMPs and potential mitigation needs with local, state, and federal agencies.
- Rectifying temporary impacts by repairing, rehabilitating, or restoring the affected resource.
- Reducing or eliminating the impact over time by preservation and maintenance operations.
- Compensating for permanent impacts by replacing, enhancing, or providing substitute resources or environments. Compensation for unavoidable impacts is consistent with state and federal mitigation rules and guidance. Priority was placed on on-site compensatory mitigation first, but considers off-site mitigation options where appropriate. In choosing between mitigation options, the likelihood for success, ecological sustainability, practicability of long-term monitoring and maintenance, and relative costs is evaluated. The mitigation goal is to replace the lost or impaired ecosystem functions in accordance with applicable laws, regulations, and procedures.
- Short-term impacts to water quality would be addressed through a Stormwater Pollution Prevention Plan, which would include construction BMPs, such as appropriate measures to prevent accidental spills of chemicals and materials and ways to minimize vegetation removal and/or replant disturbed areas.
- Long-term impacts to water quality would be addressed through local, state, and federal
 requirements for the prevention of increases to pollutant loads and for standards and
 requirement for stormwater treatment.

Refer to the Wetlands Technical Report for further details on wetland compensatory mitigation needs and requirements.

2.7 Coordination

This technical report was developed in collaboration with federal, state, and local agencies, including the Environmental Protection Agency (EPA), USFWS, NMFS, ODFW, WDFW, DSL, Washington Department of Ecology (Ecology), the City of Vancouver, Metropolitan Regional Government (Metro), and the City of Portland. Regular meetings were held, beginning in 2005, with representatives from the federal and state environmental regulatory agencies (a group formed specifically to provide input on this project, and known as the Interstate Collaborative Environmental Process [InterCEP]).

Working groups for fisheries and water quality also met to discuss specific project elements. These meetings occurred between 2006 and 2010.

Native American tribes with resource interests relevant to this project also provided input and guidance in developing this report.

Interstate 5 Columbia River Crossing Ecosystems Technical Report for the Final Environmental Impact Statement

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3. Affected Environment

3.1 Introduction

The I-5 bridges connect two major metropolitan areas, and therefore the surrounding landscape is characterized by urban development interspersed with remnant natural habitat areas in the form of riparian buffers, open space and parks, and the mainstem Columbia River. All natural areas have been modified to suit the urban landscape and the needs of the urban population. Wildlife species that currently utilize the project area appear to have become relatively habituated to ambient levels of noise, light, and activities associated with large urban centers, at least for portions of their life cycles. City and county zoning and planning for habitat protection have maintained areas (albeit small and disjointed) of aquatic and riparian habitat that support listed and non-listed fish, sensitive reptiles and amphibians, mammals, and migratory birds.

3.2 Regional Conditions

Compared to historical conditions, the availability and quality of fish, wildlife, and plant habitat in the project area has been reduced by human settlement and development.

3.2.1 Regional Aquatic Conditions

The Columbia River and its tributaries are the dominant aquatic system in the Pacific Northwest. The Columbia River originates on the west slope of the Rocky Mountains in Canada and flows approximately 1,200 miles to the Pacific Ocean, draining an area of approximately 258,000 square miles. The ocean influence reaches 23 miles upstream from the river mouth in the form of salt water intrusion from the Columbia River estuary. Coastal tides influence the flow rate and river level up to Bonneville Dam at RM 146.1 (USACE 1989). Levees, built along the river between 1919 and 1921, and dams built between the 1930s and 1970s, have significantly altered hydrologic flow and reduced the abundance and quality of fish and wildlife habitat in the project area. The lower Columbia River is used for transport of commercial goods, irrigation, power generation, and recreation. The banks in many portions, particularly those in the urbanized area around the project area, have been armored for flood and erosion control. Channel dredging occurs periodically to ensure passage for commercial vessels.

Aquatic habitats in the project area, in general, support populations of native, non-native, and listed fish species in rivers, backwater areas, small creeks, ponds, and sloughs. Aquatic habitats have been subject to human modifications (e.g., dredging, filling, armoring) to accommodate commercial and residential development, and few (if any) of these habitats are in pristine condition. The North Portland Harbor connects to the mainstem Columbia River and shares many of the same attributes. Additional aquatic habitats of note in the project area include Burnt Bridge Creek in Washington, and the Columbia Slough and Fairview Creek in Oregon.

3.2.2 Regional Terrestrial Conditions

The region is classified within the western forest ecoregion (Omernik 1987), with elevations ranging from sea level to 11,240 feet. The Pacific Northwest temperate rainforest is one of the most productive forest regions in the world. Forest types of this ecoregion include old-growth conifer (e.g., Douglas-fir (*Pseudotsuga menziesii*), spruce (*Picea* sp.), hemlock), remnant

hardwoods (e.g., Oregon oak woodlands), alpine communities (e.g., montane grasslands), and riparian, wetland, and aquatic systems. The project area was historically closed upland forest/woodland with patches of grassland savannah and prairie in lowland areas near water (e.g., present-day Hayden Island) (Hulse et al. 2002).

The suite of wildlife species originally inhabiting the project area and surrounding landscape in the Lower Columbia basin included at least 18 amphibian species (e.g., Pacific treefrog), 15 reptile species (e.g., western pond turtles), 154 bird species (woodpeckers, owls, songbirds, waterfowl), and 69 mammal species (e.g., elk, cougar, coyote, bobcat) (Hulse et al. 2002). The project area is located within the Pacific Flyway, the major north-south route for migratory birds that extends from Patagonia to Alaska. Migratory birds use the area for resting, feeding, and breeding. Species that once occurred in the area but have since been extirpated, largely due to human influence, include the grizzly bear (*Ursus arctos*), California condor (*Gymnogyps californianus*), and gray wolf (*Canis lupus*). Abundance and distribution of other species have sharply declined, some to the point of requiring legal protection (e.g., northern spotted owl [*Strix occidentalis*]). Other species have adapted to the conversion in land and habitat cover, persisting or even benefiting from the changes (e.g., raccoons [*Procyon lotor*], red-tailed hawks [*Buteo jamaicensis*]).

Native Americans lived in the region for 11,000 years before the arrival of Euro-American settlers. However, human populations were very low in the region prior to settlement (Hulse et al. 2002). As the region became settled by mineral and timber prospectors in the 1840s and 1850s, the area grew into a major West Coast port, and urban areas gradually displaced wildlife habitat. Current urbanized conditions have impacted habitat, making the project area unsuitable for historic population sizes of large mammals and many native amphibians, reptiles, birds, and other wildlife that were once common in the project area. Most of these native species still occur in the project vicinity but in reduced numbers.

Terrestrial wildlife species that currently occur in the project area—for example, bald eagles (Haliaeetus leucocephalus) and peregrine falcons (Falco peregrinus anatu)—have adapted to some extent to the urban environment and are able to nest and forage in and near the project area. Large- and medium-sized mammals (e.g., ungulates, carnivores) may occasionally be seen near these urban environments, and some have adapted to living in developed urban areas (e.g., red fox [Vulpes vulpes], coyote [Canis latrans]). However, for the most part, these species no longer occur in the project area. Terrestrial habitat is limited to relatively small, patchy areas protected by city regulations (e.g., wetlands, forested park areas, open spaces, and riparian buffers) and currently support species with relatively small home ranges and restricted habitat requirements (e.g., turtles). Portions of the region adjacent to the project area (e.g., Forest Park, Vanport Wetlands, and the western end of Hayden Island) retain forested and wetland habitats capable of supporting native wildlife.

3.2.3 Regional Botanical Conditions

Due to the highly urbanized character of the project area, most natural habitat for native plants has been lost or highly degraded through land use conversion from natural to urban use. Remaining habitat for botanical resources, particularly for rare plants, is restricted to open space, wetlands, riparian buffers, and park lands managed under protective mandate. These habitats tend to be relatively small and isolated from each other, limiting the distribution of native plants. Nonnative plants and noxious weeds are ubiquitous in the project area and further limit the ability of native plants to persist in most of the remaining suitable habitat.

3.3 Aquatic Resources

In this technical report, the term "aquatic resources" refers primarily to fish species and their habitat. Wetlands are discussed in the Wetlands and Waters Technical Report. Water quality is an important component of habitat for listed and non-listed aquatic (and terrestrial) species.

3.3.1 Summary of Aquatic Habitats

The project area contains the following water bodies: the Lower Columbia River, North Portland Harbor, Burnt Bridge Creek, Columbia Slough, and Fairview Creek. These are described individually in more detail below.

3.3.1.1 Columbia River and North Portland Harbor

The I-5 bridges are located at RM 106 of the Columbia River. The project area within the Columbia River extends from RM 101 to 118 (see description of the project area in section 2.2). The Columbia River within the project area is highly altered by human disturbance. Urban development extends up to the shoreline. There has been extensive removal of historic streamside forests and wetlands. Riparian areas have been further degraded by the construction of dikes and levees and by the placement of stream bank armoring. For several decades, industrial, residential, and upstream agricultural sources have contributed to water quality degradation in the river. Additionally, the river receives high levels of disturbance in the form of heavy barge traffic.

The 12 major dams located in the Columbia Basin are the dominant forces controlling flow in the project area. Bonneville Dam in particular influences flow in the project area, and all the dams buffer temporary hydrologic effects within the basin. Consequently, the Columbia River upstream of Bonneville Dam is a highly managed stream that resembles a series of slack-water lakes rather than its original free-flowing state. The major second factor regulating stream flow in the project area is tidal influence from the Pacific Ocean. Saltwater intrusion from the Pacific Ocean extends approximately 23 miles upstream from the river mouth at Astoria. Coastal tides influence the flow rate and river level up to Bonneville Dam at RM 146.1 (USACE 1989).

The Columbia River estuary is generally considered to be the portion of the Columbia River extending from the mouth to all tidally influenced areas (that is, to Bonneville Dam) (NMFS 2007). Therefore, the project area is part of the estuary.

The substrate of the river within the project area is predominantly composed of sand, with relatively small percentages of fine sediments and organic material (DEA 2006; NMFS 2002). A bathymetric study completed in 2006 found significant scouring on the upstream side of each bridge pier, and scour channels on the downstream side (DEA 2006). The scouring ranged from approximately 10 to 15 feet deep. Bedload transport patterns were evident in the form of sandwaves, a natural feature of the river bottom that indicates the influence of the currents and that continuously moves and shifts with the currents. The sandwaves observed in this study were especially distinct on the downstream side of the bridge. The sandwaves in the middle of the river were regular, while the sandwaves on the northern downstream side were larger and more irregular. The northern upstream side of the bridge was relatively smooth and had few to no sandwaves, while the southern upstream side had irregular sandwaves. Average river depth was approximately 27 feet (DEA 2006). Shallow-water habitat (defined as 20 feet deep or less) is present along both banks, but is more abundant along the Oregon bank (Exhibit 3-2).

Shallow and near-shore habitat is present in the project area on both the Oregon and Washington shores and is influenced by flow and sediment input from tributaries and the mainstem river,

which eventually settles to form shoals and shallow flats. This shallow-water habitat is used extensively by juvenile and adult salmonids for migrating, feeding, and holding. Phytoplankton, microdetritus, and macroinvertebrates are present in shallow areas and serve as the prey base for salmonids and other native fish (USACE 2001). A recent study along Hayden Island documented suitable rearing habitat for naturally spawned juvenile Chinook salmon, and to a lesser extent, for naturally produced juvenile chum, coho, and steelhead in the vicinity of the project area (NOAA 2009).

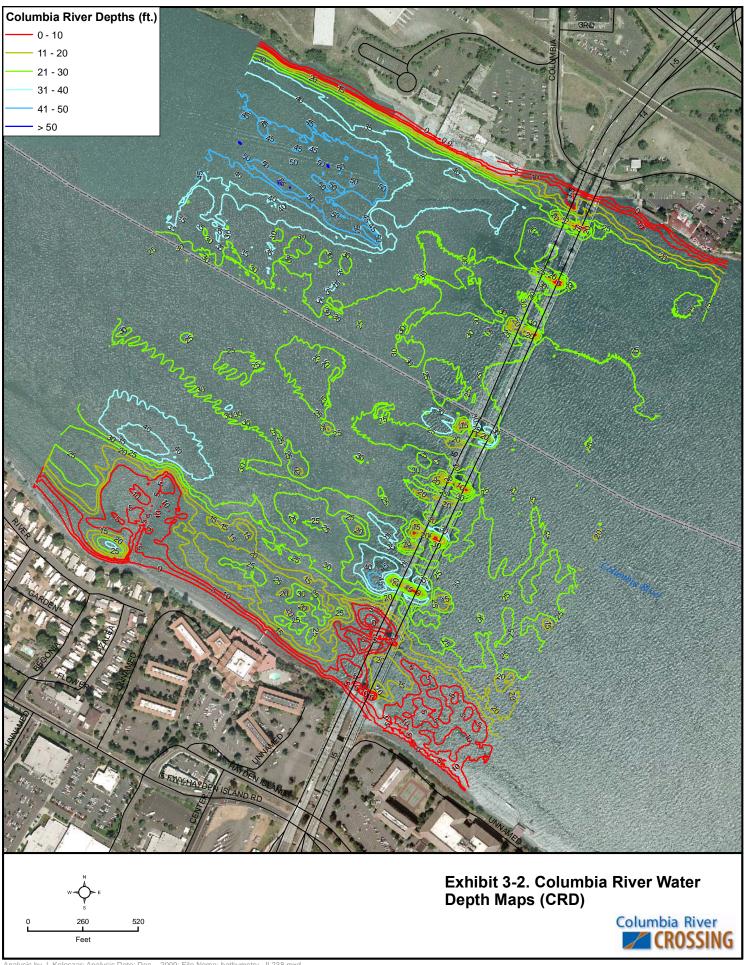
Hydrology has been significantly altered from historical conditions. Landform and bridge footings are the dominant and subdominant floodplain constrictions, respectively. Ten bridge footings are currently located below OHW. A flood control levee runs along the south bank of North Portland Harbor and forms a boundary between the adjacent neighborhoods and the harbor. Numerous upstream dams, shoreline levees, and channel dredging have restricted habitat-forming processes such as sediment transport and deposition, erosion, and natural flooding. Therefore, habitat complexity is reduced, and shallow-water habitat areas can no longer form. For this reason, these habitats are particularly lacking in the project area. Shoreline erosion rates are likely slower than they were historically due to flow regulation. The river channel is deeper and narrower than historical conditions.

Sand and gravel mining routinely occurs in several locations in the Columbia River portion of the project area. Multnomah County has issued seven permits for sand and gravel mining from September 2006 to May 2019 (Exhibit 3-1).

Exhibit 3-1. Multnomah County Sand and Gravel Permits

File Number	Name	TRS	Approximate River Mile	Tax Lot	Issued	Expires
SG-17209	Pacific Rock Products LLC	01N03E22	119	500	9/28/2006	9/27/2009
SG-16094	Columbia River Sand and Gravel, Inc.	02N01W24	102		8/1/2007	7/31/2010
SG-17111	Northwest Aggregates Co.	01N03E20A	117.5	100	7/1/2007	6/30/2010
SG-7174	Northwest Aggregates Co.	01N03E21	118	200,300	7/1/2007	6/30/2010
23300-SG	Morse Brothers, Inc.	01N03E23	120	100, 200, 300	12/15/2007	12/14/2010
24822-SG	Morse Brothers, Inc.	01N03E23C	120	100	2/19/2008	2/18/2011
25030-SG	Rose City Yacht Club, Inc.	01N01E01	109	100 & 200	6/1/2009	5/31/2019

Some high-quality backwater and side channel habitats have persisted along the Lower Columbia River banks and near undeveloped islands (USACE 2001) outside of the project area. These habitats contain high-quality wetlands and riparian vegetation, such as emergent plants and low herbaceous shrubs.



Typical riparian vegetation in the project area is shown in Exhibit 3-3 and Exhibit 3-4. Exhibit 3-5 provides riparian vegetation area estimates for the Columbia River. Data were collected from the banks of the stream within 500 feet upstream and downstream of the bridge crossing (1,000 feet total). The riparian vegetation was visually surveyed.



Exhibit 3-3. North Bank of Columbia River Looking Downstream Toward Existing I-5 Bridges



Exhibit 3-4. South Bank of Columbia River (foreground) Upstream and Downstream of Existing I-5 Bridges

Tree canopy in the project riparian areas is generally absent or sparse. Where present, typical canopy dominants include native willows (*Salix* spp.) and black cottonwood (*Populus balsamifera*) species and non-native species such as ailanthus (*Ailanthus altissima*). The understory is typically dominated by non-native species such as Himalayan blackberry (*Rubus armeniacus*) and ailanthus, and native species such as roses (*Rosa* sp.) and willows. Ground cover is typically dominated by non-natives such as English ivy (*Hedera helix*), reed canarygrass (*Phalaris arundinacea*) and Himalayan blackberry.

The riparian area within the project area is relatively degraded. As a result, shallow-water habitat has only sparse vegetative cover. Because riparian areas are limited in size and are unlikely to support productive vegetative communities in this urban setting, there is little potential for future large wood recruitment. Fish cover elements are generally sparse to absent in the project area, although some boulders and artificial structures are present.

Exhibit 3-5. Riparian Vegetation Cover Estimate in the Project Area for the Columbia River

	Vegetation Type and Density (% cover)						
	North Bank Upstream	North Bank Downstream	South Bank Upstream	South Bank Downstream			
Canopy (> 15 ft high)							
Vegetation Type	None	Deciduous	Deciduous	None			
Big trees (Trunk > 1 ft dbh)	Absent (0%)	Absent (0%)	Absent (0%)	Absent (0%)			
Small trees (Trunk < 1 ft dbh)	Absent (0%)	Sparse (< 10%)	Sparse (< 10%)	Absent (0%)			
Understory (1.5 to 15 ft high)							
Vegetation Type	Mixed	Mixed	Mixed	Mixed			
Woody Shrubs & Saplings	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)			
Non-Woody Herbs, Grasses & Forbs	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)			
Invasive Species	Heavy (40-75%)	Heavy (40-75%)	Heavy (40-75%)	Heavy (40-75%)			
Ground Cover (0.0 to 1.5 ft high)							
Vegetation Type	Mixed	Mixed	Deciduous	Mixed			
Woody Shrubs & Saplings	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)			
Non-Woody Herbs, Grasses & Forbs	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)	Sparse (< 10%)			
Barren, Bare Dirt, or Duff	Heavy (40-75%)	Heavy (40-75%)	Heavy (40-75%)	Heavy (40-75%)			
Invasive Species	Moderate (10-40%)	Moderate (10-40%)	Sparse (< 10%)	Moderate (10-40%			

Water temperatures of the Columbia River at Washougal, Washington range from approximately 6°C in early spring to approximately 22°C in late summer (USGS 2007). Temperatures in the project area are assumed to be similar to this sample site, which is 2 miles upstream of the confluence of the Columbia and Washougal Rivers. Desirable water temperatures for young salmonids during downstream migration range from 6.7 to 13.3°C. In freshwater, temperatures greater than 23°C are lethal for juvenile salmonids, and temperatures greater than 21°C are lethal for adult salmonids (USACE 2001).

As discussed in the Water Quality and Hydrology Technical Report, the Columbia River does not meet Oregon Department of Environmental Quality (DEQ) standards (and is 303(d) listed) for the following parameters: temperature, PCBs, PAHs, DDT metabolites (DDE), and arsenic (DEQ 2007). The DEQ does not differentiate between the North Portland Harbor and Columbia River when compiling the 303(d) list; therefore, these listings also apply to the North Portland Harbor. The Columbia River is not on Washington State's 303(d) list for any parameters (Ecology

2009a). In addition to the 303(d) listings, EPA has approved total maximum daily loads (TMDLs) for the Columbia River for dioxin and total dissolved gas (DEQ 1991, 2002).

As discussed in the Water Quality and Hydrology Technical Report, runoff from the I-5 bridges over Hayden Island discharges directly to the Columbia River through roadside grates located along the entire span. Runoff from the bridges is not treated prior to release to the river.

Refer to the Water Quality and Hydrology Technical Report for a description of the Columbia River floodplain, hydrology, and details on stormwater outfalls.

The North Portland Harbor is a large side channel of the Columbia River located along the southern banks of Hayden Island. The harbor branches off the Columbia River upstream (east) of the existing bridges and flows approximately 5 miles downstream (west) before rejoining the mainstem Columbia. I-5 crosses the North Portland Harbor at approximately RM 4, and this crossing is referred to as the North Portland Harbor bridges (Exhibit 3-6).



Exhibit 3-6. North Portland Harbor Bridges

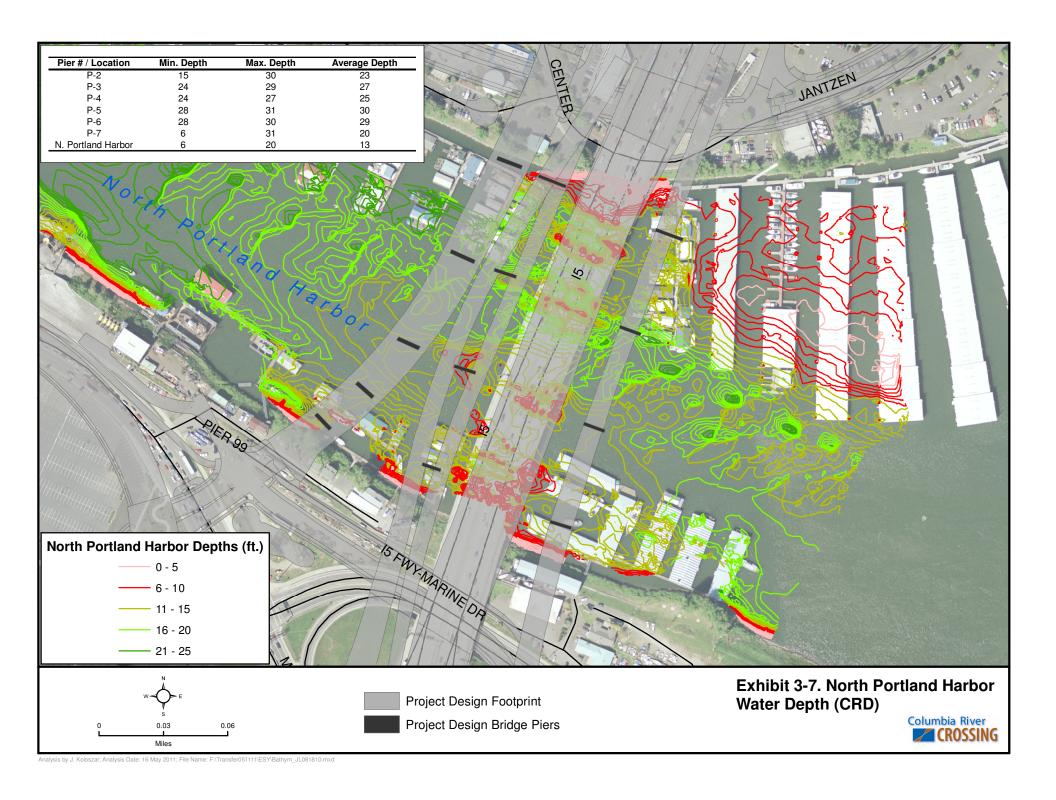
The aquatic description of the Columbia River also applies to North Portland Harbor. Much of the fish cover provided in North Portland Harbor consists of permanently moored floating homes and boathouses. Landform, specifically levees, and bridge footings are the dominant and subdominant floodplain constrictions, respectively.

The substrate of the harbor within the project area is predominantly composed of sand, with relatively small percentages of fine sediments and organic material. A bathymetric study completed in 2006 (DEA 2006) found deep scouring near the ends of the downstream piers on the north bank of the slough, with scour holes approximately 8 to 10 feet deep. Scouring around the upstream piers was approximately 3 to 7 feet. Scouring was more pronounced around the northern piers than the southern piers. A particularly deep (approximately 21 feet) area on the south side of the channel, downstream of the existing bridge, is indicative of a fast-moving current through the harbor. The average depth of the harbor was approximately 14 feet (Exhibit

3-7). Shallow-water habitat (defined as 20 feet deep or less) is present throughout the project area in North Portland Harbor.

The Columbia River and North Portland Harbor provide holding, migration, and limited rearing habitat for Chinook salmon (*Oncorhynchus tshawytscha*), chum salmon (*O. keta*), sockeye salmon (*O. nerka*), steelhead trout (*O. mykiss*), coho salmon (*O. kisutch*), and eulachon (*Thaleichthys pacificus*), as well as for species of concern (SOC) such as coastal cutthroat trout (*Oncorhynchus clarkii* clarkii) and Pacific lamprey (*Entosphenus tridentatus*). Additional native fish that are known to occur in the Columbia River and North Portland Harbor include white sturgeon (*Acipenser transmontanus*), green sturgeon (*Acipenser medirostris*), suckers (*Catostomus* spp.), sticklebacks (*Gasterosteus* spp.), starry flounder (*Platichthys stellatus*), sculpin (*Cottus* spp.), northern pikeminnow (*Ptychocheilus oregonensis*), shiners (*Cyprinidae*), peamouth (*Mylocheilus caurinus*), and chiselmouth (*Acrocheilus alutaceus*).

Aquatic organisms that constitute the prey base for salmonids and other fish in the Lower Columbia River include invertebrates such as sand shrimp, mysids, crabs, zooplankton (e.g., daphnids, chironomid larvae), and floating insect larvae and adults. Benthic species present in the Columbia River and North Portland Harbor include mussels (e.g., *Anodonta* spp.). Native species share aquatic habitat with listed salmonids and other aquatic species; therefore, habitat description, habitat quality parameters, and project impacts described for listed aquatic species also apply to populations of non-listed native species that occur within the project area.



3.3.1.2 Columbia Slough

The Columbia Slough (also known as the Slough) is a slow-moving, low-gradient drainage canal running nearly 19 miles from Fairview Lake in the east to the Willamette River in the west. Running roughly parallel to the Columbia River, the Slough is a remnant of the historic system of lakes, wetlands, and channels that dominated the south floodplain of the river. Drainage and flood control in the Slough is provided via a system of dikes, pumps, weirs, and levees (CH2M Hill 2005). The Columbia Slough watershed drains approximately 37,741 acres of land in portions of Portland, Troutdale, Fairview, Gresham, Maywood Park, Wood Village, and Multnomah County (unincorporated areas). The Slough and surrounding area were historically used by Native Americans for fishing, hunting, and gathering food (COP 2009).

The Slough is divided into upper, middle, and lower reaches. The Upper and Middle Sloughs receive water inputs from Fairview Lake, groundwater, and stormwater from the Portland International Airport (PDX) and other industrial, commercial, and residential sites in the surrounding area. Water levels in the Upper and Middle Sloughs are managed to provide adequate flows for pollution reduction (e.g., during PDX de-icing/anti-icing events) and surface water withdrawals, flood control, and recreation (COP 2009). PDX is constructing new facilities to control releases of de-icing/anti-icing materials. Upgrades to the City of Portland sewer system were done in 2000 to control combined sewer overflows to the Slough.

The project area crosses the Lower Slough at RM 6.5 (CH2M HILL 2005) The predominant land use around the Slough in the project vicinity is light industrial, with some residential. The Lower Slough extends from the Peninsula Drainage Canal (which is the border between Peninsula Drainage District No. 1 and Multnomah County Drainage District No. 1) to the Willamette River. The Lower Slough connects to the Willamette River approximately 6.5 miles west of the project area, within a mile of the confluence of the Columbia and Willamette Rivers (COP 2009). The Lower Slough experiences from 1 to 3 feet of tidal fluctuation in water surface daily. Water levels are generally unmanaged in this portion of the Slough, but are affected by the management of the dams on the Columbia and Willamette Rivers. Water depth in the Lower Slough ranges from 2.0 to 4.5 feet NGVD. The slough is generally between 100 and 200 feet wide. The Lower Slough receives water inputs from combined sewer overflows, stormwater, Smith and Bybee Lakes, leachate from the St John's Landfill, and the Upper Columbia Slough (COP 2009).

The water column in the Columbia Slough often contains algal and aquatic macrophyte growth, especially in summer months when flow is low and temperatures are high. The Slough is a lentic (still water) system with low dissolved oxygen levels. However, the Slough provides habitat for many fish and wildlife species. As of 2004, 19 species of fish (including juvenile salmonids, lamprey, sculpins, threespine sticklebacks, and suckers), freshwater shrimp, and crawfish, had been identified in the Lower Slough. It provides some of the only remaining off-channel and refugia habitat in the Lower Willamette River area (COP 2009).

Anadromous fish can access the Lower Columbia Slough up to an impassable levee near NE 18th Avenue (RM 8.3). At Smith and Bybee Lakes, a water control structure allows fish passage. Recent genetic analyses of juvenile Chinook in the Slough show that juveniles originating from Middle and Upper Columbia River ESA-listed ESUs are present in the Slough from January to June (Teel et al. 2009). The Slough is accessible to and provides potentially suitable habitat for juveniles of most upper Columbia River and Willamette River salmonid runs. Juveniles are not likely to be present in the Slough during summer months (approximately June through September, depending on the year) as water temperatures are often too high to support juvenile salmonids (COP 2009).

Benthic habitat in the Lower Slough is dominated by sand, is extremely low in dissolved oxygen, and contains some toxic pollutants (COP 2009). Generally, the benthic community, including 36 taxa, increases in abundance from the Lower to the Upper Slough. This increase in species abundance is correlated to an increase in silt dominance, which increases with distance upstream in the Slough. Most of the species are adapted to low dissolved oxygen levels and still water conditions. The benthic community in the Columbia Slough appears to be similar in species richness and density to other similar aquatic habitats in the region (COP 2009).

Riparian habitat along the Slough has been significantly impacted by urban development along most portions of the Slough. Remaining areas of vegetation generally occur in a narrow band along Slough banks and are dominated by black cottonwood, Oregon ash (*Fraxinus latifolia*), willows, red osier dogwood (*Cornus sericea*), Himalayan blackberry, common snowberry (*Symphoricarpos albus*), and reed canarygrass. Both Himalayan blackberry and reed canarygrass are aggressive, non-native species. However, riparian areas in some portions along the Slough provide microclimate and shade, bank stabilization and sediment control, pollution control, streamflow moderation, organic matter input, large woody debris, and wildlife travel corridors.

Much of the Slough's wetland habitat has been filled, dredged, channelized, and/or degraded by current and past land uses. Remnant wetlands and restored wetlands do exist in the Slough watershed and provide habitat for wildlife, thermoregulation, nutrient removal, and other important ecosystem functions. The Oregon DEQ has listed irrigation, domestic and industrial water supply, livestock watering, anadromous fish passage, salmonid fish rearing, salmonid fish spawning, resident fish and aquatic life, wildlife use, hunting, fishing, boating, water contact recreation, aesthetic quality, and hydropower as beneficial uses of the Columbia Slough (Oregon Administrative Rules 340-041-0340, Table 340A; COP 2009).

Several restoration efforts are ongoing in the Columbia Slough area. The City of Portland's Watershed Revegetation Program and its community partners are conducting non-native species removal and native plantings in many areas along the Slough. The Multnomah County Drainage District (MCDD) now uses in-channel equipment to perform repairs and maintenance of channel and bank areas. Formerly, MCDD cleared vegetation to access these areas from the shore. Both vegetation enhancement and MCDD's alteration of maintenance practices have resulted in an increase in native plant diversity and cover in the Slough watershed. The City of Portland Bureau of Environmental Services (BES) has been involved in revegetation efforts in the Slough watershed since 1996 and has successfully re-established native vegetation along more than 40 miles of Slough streambank within the City of Portland (COP 2009).

3.3.1.3 Burnt Bridge Creek

Burnt Bridge Creek is a small perennial tributary to the Lower Columbia River. It originates near the Mill Plain suburb east of Vancouver, Washington and flows west (roughly paralleling SR 500 for approximately 5 miles) to its outlet at Vancouver Lake. The lake then drains into the Lower Columbia River via Lake River. I-5 crosses Burnt Bridge Creek at approximately RM 2.

Within and upstream of the project area (between Leverich Park and Fourth Plain Boulevard), canopy cover increases and many pools are present, providing good rearing and spawning habitat (WDFW 2007). Portions of the creek that are designated as riparian protection zone (e.g., within Leverich Park) tend to have retained characteristics of higher quality habitat, such as functioning pool and marsh habitat.

Within the project area, the stream passes through a valley constrained by surrounding land uses (primarily residential development). Stream slope is between 0 and 2 percent, but approximately 80 percent of the stream has a gradient of less than 0.1 percent (PBS 2003).

Burnt Bridge Creek enters the project area in Leverich Park, northeast of the I-5/SR 500 interchange. In the park area, the creek has substantial overhead cover from large-diameter trees and shrubs in some areas, and sparse cover by widely spaced large-diameter trees in areas maintained by park staff. In the more open areas within the park, the banks are highly eroded by regular visitor usage and mowing of herbaceous vegetation in the vicinity of the channel. Substrate within the park consists of fine sediments and gravels. Both riffles and pools are present within the park channel (WDFW/MHCC 1999).

Dominant tree species in this portion include natives such as Douglas-fir, black cottonwood, willow, and ash. The understory is dominated by non-native Himalayan blackberry and natives such as red alder (*Alnus rubra*), red osier dogwood, and beaked hazelnut (*Corylus cornuta*). Ground cover is typically dominated by non-native species such as Himalayan blackberry, reed canarygrass, and teasel (*Dipsacus sylvestris*). Riparian vegetation cover near Burnt Bridge Creek within the project area is summarized in Exhibit 3-8.

Exhibit 3-8. Riparian Vegetation Cover Estimate within the Project Area for Burnt Bridge Creek

	Vegetation Type and Density Left Bank (% cover)	Vegetation Type and Density Right Bank (% cover)
_	Mixed	Mixed
Canopy (> 15 ft high)		
Big trees (Trunk > 1 ft dbh ^a)	Moderate (10-40%)	Moderate (10-40%)
Small trees (Trunk < 1 ft dbh)	Moderate (10-40%)	Moderate (10-40%)
Understory (1.5 to 15 ft high)		
Woody Shrubs & Saplings	Very Heavy (> 75%)	Very Heavy (> 75%)
Non-Woody Herbs, Grasses & Forbs	Sparse (< 10%)	Heavy (40-75%)
Invasive Species	Very Heavy (> 75%)	Very Heavy (> 75%)
Ground Cover (0.0 to 1.5 ft high)		
Woody Shrubs & Saplings	Heavy (40-75%)	Heavy (40-75%)
Non-Woody Herbs, Grasses & Forbs	Sparse (< 10%)	Very Heavy (> 75%)
Barren, Bare Dirt, or Duff	Very Heavy (> 75%)	Very Heavy (> 75%)
Invasive Species	Absent (0%)	Absent (0%)

a Diameter at breast height.

From Leverich Park, the Burnt Bridge Creek channel passes under Leverich Park Way through a cement culvert and onto City of Vancouver property adjacent to I-5. The channel is armored for approximately 100 feet, after which it continues north, parallel to I-5 and Leverich Park Way, through a silt-dominated channel. The vegetation surrounding this portion of the channel is dominated by reed canarygrass with some overhanging Himalayan blackberry and red osier dogwood. Site observations indicate that the channel banks are undercut due to the growth habit of reed canarygrass and eroded due to the presence of nutria (*Myocastor coypus*).

Approximately 500 feet north of the cement culvert, Leverich Park Way bends to the west and the Burnt Bridge Creek channel passes under the roadway through a large corrugated metal pipe culvert. The channel continues north through a densely vegetated, privately owned area for about 200 feet. No permission to enter this area was granted during field visits to assess habitat and site characteristics. The channel then flows through a culvert under I-5, continuing north alongside a WSDOT wetland mitigation site to the west and Bonneville Power Administration (BPA) property and private land to the east. From the second culvert under Leverich Park Way to the point where Burnt Bridge Creek exits the project area, the channel is dominated by fine sediments

(PBS 2003) and has moderate to dense overhanging vegetation consisting of deciduous and coniferous tree and shrub species.

Between I-5 and Vancouver Lake, the creek is low gradient with moderate canopy cover, and contains marsh and pool features that provide good rearing habitat for juvenile salmonids.

Within the project area, Burnt Bridge Creek is on Ecology's 303(d) list for fecal coliform and temperature (DEQ 2007). Ecology has not approved any TMDLs for Burnt Bridge Creek (Ecology 2009b). Some stormwater runoff is routed to the creek through pipes and ditches, but most runoff is discharged into the ground through buried infiltration facilities. Three stormwater outfalls from I-5 discharge treated water into Burnt Bridge Creek: one on the east side of I-5 and two on the west side of I-5. Runoff from I-5 at the north of the SR 500 interchange area is routed to a retention pond east of I-5 and south of the Main Street interchange. Retained runoff usually evaporates or infiltrates, and releases to Burnt Bridge Creek only occur during peak runoff events. Runoff from SR 500 east of I-5 flows to a detention pond located at NE 15th Avenue before being released to Burnt Bridge Creek.

All freshwater life stages of coho, Chinook, and steelhead are potentially present in Burnt Bridge Creek (Weinheimer 2007 pers. comm.). Native resident fish, including dace (Cyprinidae), threespine stickleback, redside shiners, suckers, sculpin, and lamprey, are also present in the creek (PBS 2003).

3.3.1.4 Fairview Creek

Fairview Creek is a 5-mile urban stream whose headwaters consist of a wetland near Grant Butte in Gresham. The creek drains to Fairview Lake, a tributary to the eastern portion of the Columbia Slough. Historically, the creek had been a tributary of the Columbia River, but the water from the wetlands was diverted into an artificial channel that drained into the Columbia Slough. In 1960, water managers built a dam along Fairview Creek to create Fairview Lake for water storage and recreation. Fairview Creek has two named tributaries, No Name Creek and Clear Creek (COP 2009).

Fairview Creek receives stormwater runoff from Gresham, Wood Village, and Fairview, an area of about 6.5 square miles. Average flow in Fairview Creek at the USGS gauging station near Glisan Street, approximately 1.4 miles downstream of the Ruby Junction Operations Facility, was 6.39 cfs from 1992 to 1999 (Metro 2003). The 100-year floodplain for Fairview Creek is approximately 1,288 feet wide at its widest point, adjacent to the proposed maintenance facility expansion area (Metro 2003).

DEQ has placed Fairview Creek on its 303(d) list for *E. coli* (year-round) and fecal coliform (fall/winter/spring); it also has approved TMDLs for bacteria and spring/summer temperature (COP 2008a; DEQ 2009).

Excessive fine sediments have been shown to settle in the streambeds of Fairview Creek. This is caused by the erosion of upland areas and deposition of sediments by stormwater discharged to the creek. These sediments degrade native fish spawning areas and limit suitable habitat for benthic organisms (COP 2009).

Some stream restoration has occurred along Fairview Creek. The East Multnomah Soil and Water Conservation District (SWCD), Smith Presbyterian Church, ODFW, Fairview Village, and the City of Gresham have planted riparian areas, limited human access to sections of the stream, and installed large woody debris and boulders as in-stream habitat structures. The stream system also

includes approximately 1 mile of undeveloped land, which includes parks and green spaces (Brick 2008 pers. comm.).

Fairview Creek between 185th and Marine Drive to Burnside Street has 21 stream crossings. All crossings between the Columbia Slough confluence and Glisan Street appear to be fish-passable, though some may be slightly undersized (Brick 2008 pers. comm.). Fairview Creek habitat upstream of Glisan Street was not assessed for this project.

Anadromous salmonids are not currently present in Fairview Creek. There is an impassable barrier between the lower and middle sections of the Columbia Slough located approximately at stream mile 8.3 (near NE 18th Avenue), approximately 10 miles downstream of Fairview Creek. At one time, Johnson Creek connected to the wetlands that serve as the headwaters of Fairview Creek. The two streams are not currently connected. However, it is possible that on rare occasions during extreme flood events, coho salmon or steelhead trout could enter Fairview Creek via Johnson Creek. These fish would most likely become trapped in the Fairview Creek system (Brick 2008 pers. comm.).

The creek may currently support resident fish species. Native cutthroat trout presence has been documented in only two of the remaining tributaries of the Columbia Slough: Fairview Creek and Osborn Creek (City of Portland 2009).

3.3.2 Threatened, Endangered, and Proposed Species

"Listed" species refer to those with federal and/or state threatened, endangered, or proposed status. Data on listed species were obtained from NMFS, USFWS, ORNHIC, WDNR-NHP, and WDFW-PHS. The Columbia River and North Portland Harbor are known to support listed anadromous salmonids, including Chinook salmon (*Oncorhynchus tshawytscha*), chum salmon (*O. keta*), sockeye salmon (*O. nerka*), steelhead trout (*O. mykiss*), and coho salmon (*O. kisutch*), as well as species of concern (SOC) such as Pacific lamprey (*Entosphenus tridentatus*), river lamprey(*Lampetra ayresi*), coastal cutthroat trout, and the northern distinct population segment (DPS) of green sturgeon (*Acipenser medirostris*) (Exhibit 3-9). Habitat use for these species is primarily migration, holding, and rearing. Chum salmon are known to spawn in the Columbia River upstream of the project area, near the mouth of Camas Creek (FPC 2009) and a recent study showed that Hayden Island provides limited rearing habitat for chum in the project vicinity (NOAA 2009). The southern DPS of eulachon (*Thaleichthys pacificus*) occurs in the Columbia River during migration and spawning.

Bull trout (*Salvelinus confluentus*) are federally threatened and have been documented in the Lower Columbia River at very low abundance (Gray 2007). Bull trout use of the Lower Columbia may include overwintering and feeding; the Bull Trout Lower Columbia Recovery Team considers the mainstem Columbia River to contain core habitat necessary for full recovery of the species (USFWS 2002). Critical habitat has been proposed in the Columbia River within the project area.

NMFS has determined that the southern DPS of green sturgeon may occur in Washington coastal waters and below RM 35 of the Columbia River (74 FR 52299). Northern and southern DPSs were delineated in 2003; in 2006, the southern DPS was listed as threatened, while the northern DPS was classified as a SOC. Southern green sturgeon spawn in the Sacramento River, California, while northern green sturgeon spawn in the Klamath and Rogue Rivers in Oregon. Genetic and tagging data indicate that the stocks commingle in the Columbia River estuary during the summer as subadults and adults.

Steller sea lions are listed as threatened under the federal ESA as well as by both Oregon and Washington. California sea lions (*Zalophus californianus*) and harbor seals (*Phoca vitulina*) are not listed under the ESA, but like Steller sea lions, they are protected under the MMPA.

Pacific lamprey (*Entosphenus tridentatus*, formerly *Lampetra tridentata*) have significant cultural, spiritual, ceremonial, medicinal, subsistence, and ecological value for many Native American tribes in the Pacific Northwest (Archuleta 2005, CRITFC 2008). Lamprey play a key role in the aquatic and terrestrial food web and are an indicator species for anthropogenic impacts to ecological systems (Close et al. 2002). Pacific lampreys are thought to have been historically distributed wherever salmon and steelhead occurred (USFWS 2010). However, current data indicate that distribution and abundance of Pacific lamprey have been significantly reduced by the construction of dams, water diversions, and by degradation of spawning and rearing habitat (Quigley et al. 1996). Pacific lamprey are a federal Species of Concern. For a full discussion of Pacific lamprey in the project area, refer to Appendix A.

Exhibit 3-9. Protected Aquatic/Fish Species Potentially Occurring within the Project Area

ESU/DPS (Where Appropriate) ^a Species Common Name Species Scientific Name	Federal Status ^b	OR Status ^c	WA Status ^d	Critical Habitat Present	EFH Present in Project Area ^e	ESH Present in Project Area ^f	Presence Documented in Project Area ⁹	Habitat Use within Project Area ^h
Lower Columbia River ESU Chinook salmon Oncorhynchus tshawytscha	LT	SC	SC	Yes	Yes	No	Yes	M/R/H
Upper Columbia River- Spring Run Chinook salmon Oncorhynchus tshawytscha	LE	N/A	SC	Yes	Yes	No	Yes	M/R/H
Snake River Fall-Run Chinook salmon Oncorhynchus tshawytscha	LT	LT	SC	Yes	Yes	No	Yes	M/R/H
Snake River Spring/Summer-Run Chinook salmon Oncorhynchus tshawytscha	LT	LT	SC	Yes	Yes	No	Yes	M/R/H
Lower Columbia River DPS Steelhead trout Oncorhynchus mykiss	LT	SC	SC	Yes	No	No	Yes	M/R/H
Middle Columbia River Steelhead trout Oncorhynchus mykiss	LT	SC	SC	Yes	No	No	Yes	M/R/H
Upper Columbia River Steelhead trout Oncorhynchus mykiss	LE	N/A	SC	Yes	No	No	Yes	M/R/H
Snake River Basin Steelhead trout Oncorhynchus mykiss	LT	SV	SC	Yes	No	No	Yes	M/R/H
Snake River Sockeye salmon Oncorhynchus nerka	LE	None	SC	Yes	No	No	Yes	M/R/H

ESU/DPS (Where Appropriate) ^a Species Common Name Species Scientific Name	Federal Status ^b	OR Status ^c	WA Status ^d	Critical Habitat Present	EFH Present in Project Area ^e	ESH Present in Project Area ^f	Presence Documented in Project Area ^g	Habitat Use within Project Area ^h
Lower Columbia River Coho salmon Oncorhynchus kisutch	LT	LE	None	N/A	Yes	No	Yes	M/R/H
Columbia River ESU Chum salmon Oncorhynchus keta	LT	sc	sc	Yes	No	No	Yes	M/R/H
Southwestern Washington/Columbia River Coastal cutthroat trout Oncorhynchus clarki clarki	SOC	SV	N/A	N/A	N/A	No	Yes	Unknown
Columbia River DPS Bull trout Salvelinus confluentus	LT	SC	SC	Yes	N/A	No	Yes	Unknown; potentially overwintering and feeding
Southern DPS Eulachon Thaleichthys pacificus	LT	None	sc	N/A	N/A	N/A	Yes	M,S
Pacific lamprey Lampetra tridentata	SOC	SV	None	N/A	N/A	N/A	Yes	Unknown
River lamprey Lampetra ayresi	SOC	None	sc	N/A	N/A	N/A	Unconfirmed	Unknown
Northern DPS Green sturgeon Acipenser medirostris	soc	None	None	N/A	N/A	N/A	Yes	Unknown
Southern DPS Green sturgeon Acipenser medirostris	LT	None	None	No	N/A	N/A	Unlikely	Unknown
Steller sea lion Eumetopias jubatus	LT	LT	LT	No	N/A	N/A	Yes	Transiting, Foraging
California sea lion Zalophus californianus	Protected (MMPA)	None	None	N/A	N/A	N/A	Yes	Transiting, Foraging
Harbor seal (Phoca vitulina)	Protected (MMPA)	None	None	N/A	N/A	N/A	Yes	Transiting, Foraging

- a ESU = Evolutionarily Significant Unit; DPS = Distinct Population Segment (USFWS 2008).
- b Federal status: LT = Listed Threatened, LE = Listed Endangered, P = Proposed, C = Candidate, SOC = Species of Concern, N/A = Not Applicable (USFWS 2008).
- c OR State status: LT = Listed Threatened, SC = Sensitive Critical, SV = Sensitive Vulnerable, None = No status designated, N/A = Not Applicable (Oregon Threatened and Endangered Species List).
- d WA state status: SC=state candidate, N/A = Not Applicable (WDFW-PHS).
- e EFH = Essential Fish Habitat, per the MSFCMA.
- f ESH = Essential Salmonid Habitat, per DSL and ODFW.
- g Source = StreamNet (2005).
- h Habitat uses: S = Spawning, M/R/H = Migration/Limited Rearing/Holding (StreamNet 2005, NOAA 2009).

NMFS has designated critical habitat for several of the listed salmonid evolutionarily significant units (ESUs) (or DPS for steelhead) that occur in the Columbia River and North Portland Harbor. Chinook and coho salmon habitat is also managed under the MSFCMA. The MSFCMA requires cooperation among NMFS, the Regional Fishery Management Councils, fishing participants,

federal and state agencies, and others in achieving the EFH goals of habitat protection, conservation, and enhancement. EFH comprises those waters and substrates necessary to fish for spawning, breeding, feeding, or growth to maturity, and includes the Columbia River and North Portland Harbor (NMFS 2008a). Of the fish species present in the project area, EFH applies only to Chinook and coho.

DSL, in coordination with other agencies, also designates ESH. ESH is defined as the habitat necessary to prevent the depletion of native salmon species during their life history stages of spawning and rearing (Oregon Administrative Rule [OAR] 141-102-0000). Aquatic habitats within the project area are not designated as ESH.

Formal consultation under Section 7 of the ESA to analyze effects to listed species and EFH has been completed. NMFS issued a Biological Opinion with a "not likely to jeopardize" determination for 13 salmonid stocks, southern green sturgeon, eulachon, Steller sea lion, and relevant critical habitat on January 19, 2011. NMFS also concurred with the determination that the proposed action is "not likely to adversely affect" the southern resident killer whale (*Orcinus orca*). USFWS issued a concurrence letter for a "not likely to adversely affect" determination for bull trout and designated critical habitat on August 27, 2010.

3.3.2.1 Fish Passage

There are no known fish passage barriers within the project area. Barriers to fish passage are present on tributaries to the Columbia River along its entire length, and dams are present on the river upstream of the project area. Off-channel habitat along the North Portland Harbor is extremely limited compared to likely historic conditions and has been degraded along most of the North Portland Harbor. Levees act as fish passage barriers to historic off-channel habitat in both the Columbia River and North Portland Harbor. Insufficient passage facilities (e.g., undersized and failing culverts) are present in Burnt Bridge Creek downstream of the project area but do not act as complete passage barriers. There are no known passage barriers in the Lower Columbia Slough within the project area. As noted above in section 3.3.1.4, fish passage barriers are present downstream of Fairview Creek in the Middle Columbia River Slough; however, none are present in the creek within the project area.

3.4 Terrestrial Resources

3.4.1 Summary of Terrestrial Resources

Two recently federally delisted species that may occur in or near the project area are bald eagles and peregrine falcons. Although both species have been delisted from the federal ESA, the bald eagle is still listed as threatened by Oregon and Washington, and both species' populations will be closely monitored in the near future. Bald eagles will continue to be protected by Washington and Oregon, as well as by the federal Bald and Golden Eagle Protection Act (BGEPA) and the MBTA. No known or potential bald eagle nesting or communal roosting areas exist within the project area. Bald eagles likely forage along the Columbia River and North Portland Harbor.

The peregrine falcon was federally delisted in August 1999, and was delisted by the State of Oregon in April 2007. The species is listed by the State of Washington as sensitive. Peregrine falcons are known to nest in the project area. In addition to being protected under state law, the peregrine falcon is protected by the MBTA. The State of Oregon has not prepared a conservation plan for the peregrine falcon, although their presence in the project area has been monitored since 2001 (Casey 2011).

Bridges are also home to other SOI, including bats and native birds such as swallows (also protected under the MBTA). Bats may use bridge structures for day roosts, and swallows may nest on bridge structures, particularly those over or adjacent to water.

Five terrestrial habitat types exist within the project area. These are described further in Section 3.4.2. Two of these habitat types, Westside Riparian Wetland and Herbaceous Wetlands, are priority habitats for the area (Johnson and O'Neil 2001). In addition to protecting species, local, state, and federal laws protect terrestrial habitats (Section 7).

The terrestrial habitats in the project area support rare species as well as more common native mammals, birds, amphibians, and reptiles, including but not limited to salamanders (e.g., *Batrachoseps* spp.), frogs (e.g., red-legged frogs [*Rana aurora*], tree frogs [*Hyla* spp.]), painted turtles (*Chrysemys picta*), pond turtles (*Emys marmorata*), ospreys (*Pandion haliaetus*), red-tailed hawks (*Buteo jamaicensis*), grebes (*Aechmophorus* spp.), finches (*Carpodacus* spp.), blackbirds (*Agelaius* spp.), geese (*Branta* spp.), native squirrels (*Sciurus* spp.), and raccoons (*Procyon lotor*). Many of these species are discussed under SOI (Section 4.3.4).

3.4.2 Habitat Occurrence

Habitat is the area where wildlife nest, feed, roost, and raise their young. The analysis in this document uses Johnson and O'Neil (2001) Habitat Types classification to classify the different habitats located within the project area (Exhibit 3-11).

- Open Water Lakes, Rivers, and Streams
- Urban and Mixed Environs
- Westside Lowlands Conifer-Hardwood Forest
- Herbaceous Wetlands
- Westside Riparian Wetlands

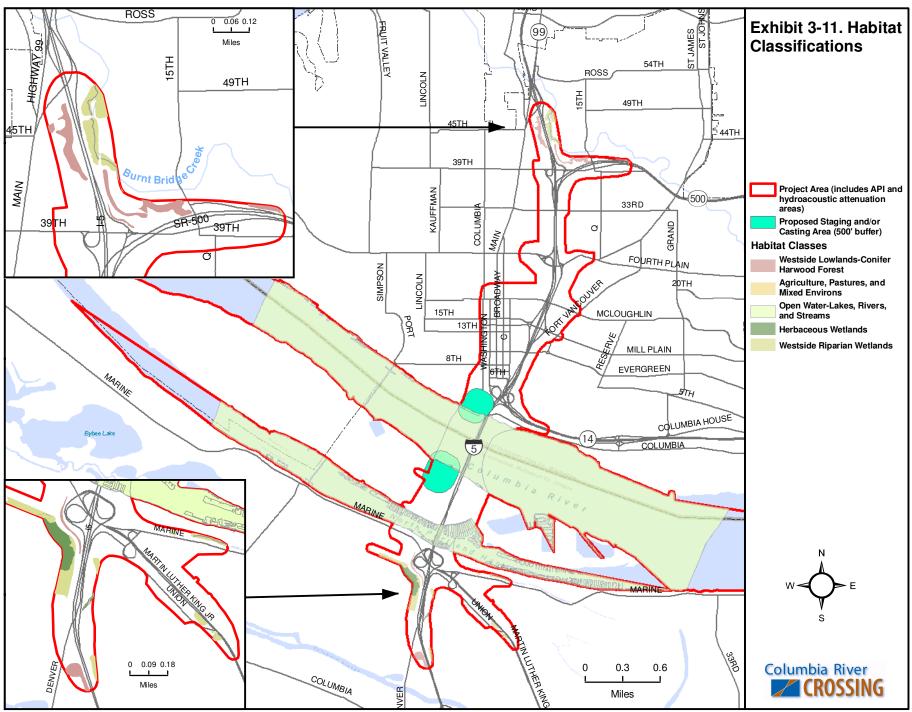
Exhibit 3-10 lists the acres of each Johnson and O'Neil habitat type occurring in the project area.

Exhibit 3-10. Acres of Habitat Classification within the Project Area

Habitat Classification	Acres
Johnson and O'Neil classifications:	
Westside Lowland Conifer-Hardwood Forest	16.9
Lakes, Rivers, Ponds, and Reservoirs	1722.8
Herbaceous Wetlands	9.1
Westside Riparian – Wetlands	30
Urban and Mixed Environs	1179.8
Т	otal: 2958.6

a Includes staging and casting areas, noise attenuation distances, and Ruby Junction (see areas shown on Exhibit 3-11).

Each of the five habitat types provide nesting, breeding, foraging, and/or dispersal habitat for migratory birds, small mammals, amphibians, reptiles, and other native species (Johnson and O'Neil 2001).



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3.4.2.1 Open Water - Lakes, Rivers, and Streams

This habitat includes all areas of open freshwater and shorelines, gravel bars, and sand bars associated with these habitats throughout the region (Johnson and O'Neil 2001). Species of interest associated with this habitat type include the bald eagle, peregrine falcon, osprey, geese and other waterfowl, migratory songbirds, Townsend's big-eared bat, purple martin, Pacific pond turtle, and northern painted turtle (Johnson and O'Neil 2001).

Within the project area, this habitat type includes the Columbia River and the North Portland Harbor.

3.4.2.2 Urban and Mixed Environs (High-density)

This habitat type consists of land containing built structures and impervious surfaces such as buildings, houses, parking lots, and roads. This habitat type is found throughout the project area and occurs within or adjacent to nearly every other habitat type. Land use types may include a mix of commercial, residential, and transportation developments. Many vegetative structural features typical of the historical vegetation have been removed. However, some remaining vegetative structures can provide habitat for nesting or roosting, and landscaping may provide foraging or nesting opportunities. High-density urban landscapes are covered with 60 to 100 percent impervious surfaces. Examples of SOI associated with this habitat type include the bald eagle, peregrine falcon, red-tailed hawk, migratory songbirds, kingfishers, and Townsend's bigeared bat (Johnson and O'Neil 2001).

These environs include core downtown areas, commercial areas, shopping malls, industrial areas, high-density housing, and transportation corridors such as I-5 (Johnson and O'Neil 2001).

3.4.2.3 Westside Lowlands Conifer – Hardwood Forest

This lowland to low montane upland forest occurs over most of western Washington, the Coast Range of Oregon, the western slopes of the Cascades in Oregon, and around the margins of the Willamette Valley. This forest is dominated by one or more of the following species: Western hemlock, Western red cedar, Douglas-fir, Sitka spruce (*Picea sitchensis*), red alder, Port-Orford cedar, and bigleaf maple. This habitat type does not include dry Douglas-fir forests where western hemlock is not able to grow (Johnson and O'Neil 2001). Examples of SOI associated with this habitat type include the bald eagle, peregrine falcon, migratory songbirds, Townsend's big-eared bat, and, historically, the yellow-billed cuckoo and purple martin (Johnson and O'Neil 2001).

Only a small portion of the project area is composed of the Westside Lowlands Conifer-Hardwood Forest habitat type; this portion consists of very small, isolated patches surrounded by urban and mixed environs (e.g., Leverich Park).

3.4.2.4 Herbaceous Wetlands

This habitat type is composed of wet meadows, marshes, fens, and aquatic beds. These habitats are wetlands or riverine floodplains that are dominated by herbaceous vegetation. Common dominants include cattails, sedges, grasses, bulrushes, and various forbs. Aquatic rooted plants that extend to the surface or floating aquatic plants are also dominants (Johnson and O'Neil 2001). Examples of SOI associated with this habitat type include the bald eagle, peregrine falcon, Townsend's big-eared bat, purple martin, and painted turtle (Johnson and O'Neil 2001).

The Herbaceous Wetlands habitat type can be found at the Vanport Wetlands complex (located west of I-5 and south of Marine Drive), immediately surrounding the open water pond/wetland

system east of I-5 Delta Park, and the closed slough east of I-5 along Whitaker Road. Please refer to the Wetlands and Waters Technical Report for more detailed information on wetlands in the project area.

3.4.2.5 Westside Riparian - Wetlands

This habitat includes all freshwater wetlands and riverine floodplains that are dominated by trees or shrubs at low elevations on the west side of the Cascades. Typical dominant species include Sitka spruce, Western red cedar, Western hemlock, red alder, black cottonwood, Oregon ash, willows, and spirea. Also included are all sphagnum bogs (forested, shrub, and herb-dominated) (Johnson and O'Neil 2001). SOI associated with this habitat type include the bald eagle, peregrine falcon, Townsend's big-eared bat, purple martin, willow flycatcher, migratory songbirds, pond turtles, and painted turtle (Johnson and O'Neil 2001).

The Westside Riparian – Wetlands habitat type is found scattered in small patches along the Columbia River and North Portland Harbor, and along the Oregon side of the Columbia River. This habitat type can also be found within the Vanport Wetlands complex. Very little riparian vegetation exists along the Columbia River. Human activities, urban development, and the absence of a riparian corridor cause the riparian area along the Columbia River to be highly disturbed. Please refer to the Wetlands and Waters Technical Report for more detailed information on wetlands in the project area.

3.5 Regional and Local Resource Protection

The project area is located within several governmental jurisdictions, and resource protection regulations vary with each jurisdiction. With the exception of Multnomah County, each of these jurisdictions has established habitat classifications that include lands within the project area. A summary of regional and local resource protection is found in Exhibit 3-12. Note that some of these areas overlap. For example, the open water habitat of the Columbia River is located within a Washington priority habitat, a Portland environmental zone (E-Zone), and a Goal 5 habitat for Metro. Refer to Section 7 for permits and approvals that may be associated with these resource protection areas.

Exhibit 3-12.	Regional a	ind Local	l Resource	Protection	in the	Project Area ^a
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Agency	Jurisdiction	Program	Habitat Protected	Acres in Project Area
WDFW	Washington State	Priority Habitats	Riparian, Urban Natural Open Space, Oak Woodland	2315.6
City of Vancouver	City of Vancouver	Critical Areas Protection Ordinance	Fish and wildlife habitat conservation areas, wetlands, frequently flooded areas, critical aquifer recharge areas, and geologic hazard areas	2375.4 ^b
Clark County	Unincorporated areas of Clark County	Critical Areas Protection Ordinance	Riparian Priority Habitat, Other Priority Habitats and Species, and Locally Important Habitats and Species	See City of Vancouver acres above
City of Portland	City of Portland	Environmental Zones	Important natural resource areas	1977.9
Metro	Portland metropolitan area	Goal 5	Regionally significant fish and wildlife habitat; Riparian habitat; Upland habitat	5106.9

a Includes staging and casting areas, noise attenuation distances, and Ruby Junction (see areas shown on Exhibit 3-13, 3-14, and 3-16).

b City of Vancouver and Clark County critical lands are merged for mapping purposes; these figures represent critical areas for both City of Vancouver and Clark County.

3.5.1 Washington

3.5.1.1 Washington Department of Fish and Wildlife

WDFW is responsible for protecting fish and wildlife species. In order to address the protection of these habitats, WDFW publishes a Priority Habitats and Species List that identifies those habitats and species that should be a priority for management and conservation. This list is largely created to inform the management and conservation efforts of landowners, agencies, governments, and members of the public who, according to WDFW, "have a shared responsibility to protect and maintain these resources" (WDFW 2008).

Priority habitats are those habitats with "unique or significant value to a diverse assemblage of species," including but not limited to a "unique vegetation type or dominant plant species, a described successional stage, or a specific structural element." One or more of the following habitat characteristics are used by WDFW to identify a priority habitat:

- Comparatively high fish and wildlife density
- Comparatively high fish and wildlife species diversity
- Important fish and wildlife breeding habitat
- Important fish and wildlife seasonal ranges
- Important fish and wildlife movement corridors
- Limited availability
- High vulnerability to habitat alteration
- Unique or dependent species

Washington classifies 18 priority habitat types, three of which occur within the project area: Riparian, Urban Natural Open Space, and Oak Woodland. These are mapped in Exhibit 3-13 as riparian and non-riparian conservation areas. These priority habitats were not field-verified during the September 2005 surveys. See Exhibit 3-12 for a summary of acreage of this habitat type occurring in the project area.

3.5.1.2 City of Vancouver

As mandated by the Growth Management Act (GMA) (RCW 36.70A), the City of Vancouver designates and protects through ordinance ecologically sensitive and hazardous areas, termed here "critical areas," as well as their functions and values. Critical areas include wetlands, fish and wildlife habitat conservation areas, geologically hazardous areas, frequently flooded areas, and areas with critical effects on aquifers providing potable water.

Fish and wildlife habitat conservation areas include lakes, streams, rivers, naturally occurring ponds, riparian buffers, and any habitat that serves any life stage of state of federally designated endangered, threatened, or sensitive fish or wildlife species. These conservation areas can also include habitats of Local Importance—habitats that are not designated as Priority Habitat by WDFW but that serve a local importance as recognized by the City.

Frequently flooded areas have been identified as having special flood hazards by the Federal Insurance Administration and the Federal Emergency Management Agency (FEMA) in scientific and engineering reports entitled *The Flood Insurance Study for the City of Vancouver*, *Washington, Clark County* (1981) and *The Flood Insurance Study for Clark County*, *Washington*

(1991), respectively, and accompanying Flood Insurance Rate Maps, Flood Boundary-Floodway Maps, and any revisions thereto.

Geologic hazard areas include landslide, seismic, and erosion hazard areas. Landslide hazard areas include areas where slopes on the property are greater than 25 percent, and areas of historic or active landslides, potential instability, or older landslide debris. Seismic hazard areas include areas subject to liquefaction or dynamic, ground-shaking amplification, and fault rupture hazard areas as identified in previous scientific studies. Erosion hazard areas include areas of potential severe soil or bank erosion as determined by previous Natural Resources Conservation Service (NRCS) studies. Further details on geologic hazards are discussed in the Geology and Soils Technical Report.

The Critical Areas Ordinance (CAO) requires that development in these critical areas result in no net loss of function, including, but not limited to, water quality protection and enhancement, fish and wildlife habitat, and ground water recharge and discharge. This CAO is also intended to "protect residents from hazards and minimize risk of injury or property damage" (City of Vancouver, Municipal Code, Chapter 20.740).

A small portion of the project area is identified as Critical Sensitive Lands. Along the Columbia River in Washington, the riparian area is designated as a critical area (Exhibit 3-13). In addition, under Vancouver Municipal Code 14.26, (Water Resources Protection), the entire City is considered a critical area for the purpose of keeping the City's water resources from being contaminated. The City of Vancouver has jurisdiction over critical areas within the City boundaries. Clark County has jurisdiction over critical areas in the unincorporated area of the County. This discussion of critical areas refers to critical areas within the City of Vancouver. See Exhibit 3-12 for a summary of acreage of this habitat type occurring in the project area.

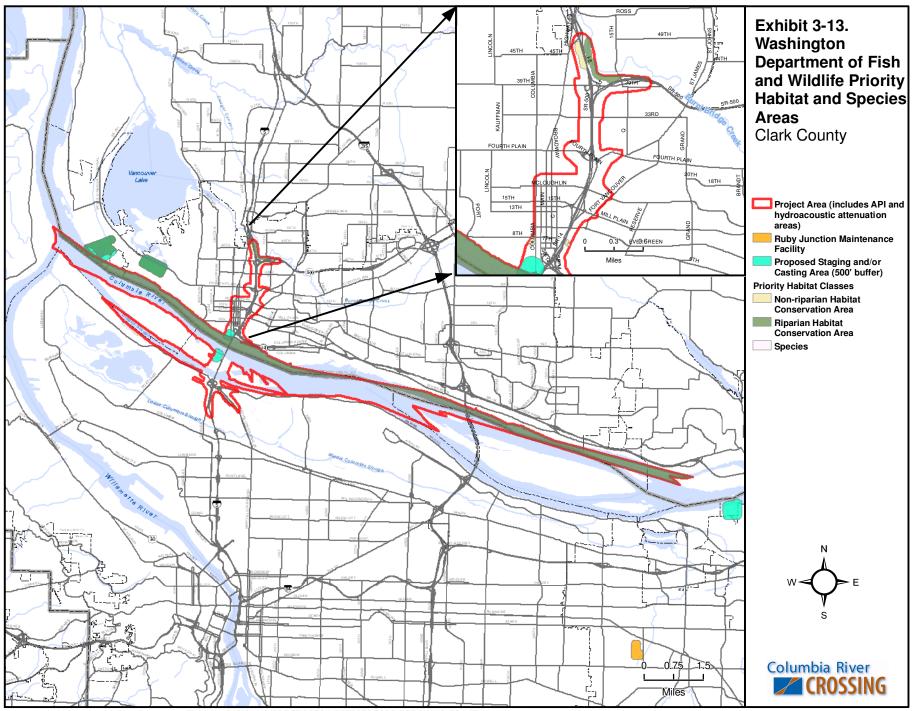
3.5.2 Oregon

3.5.2.1 City of Portland

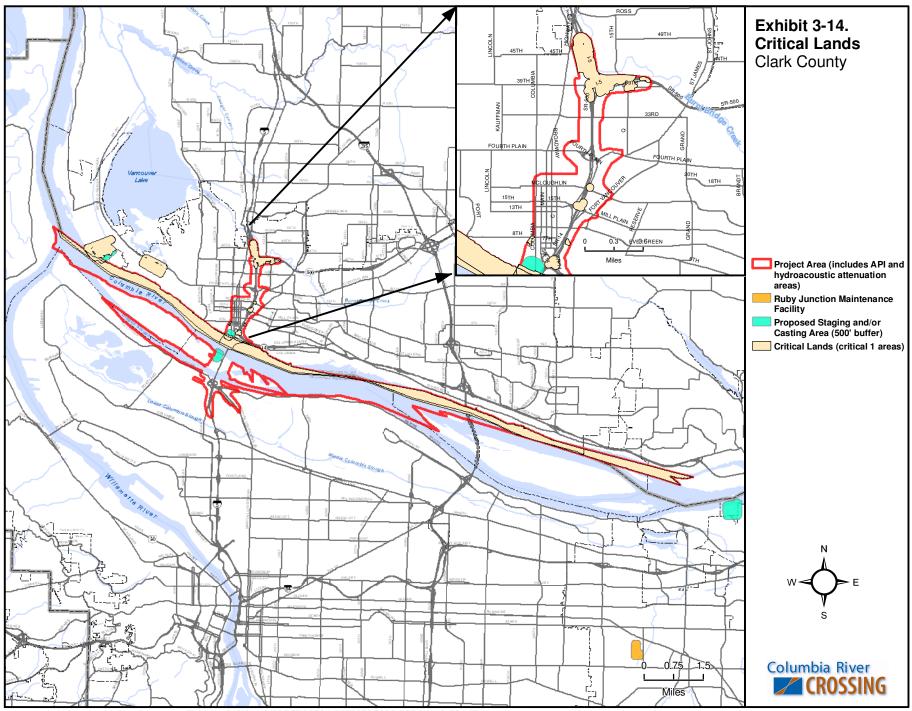
The City of Portland applies two environmental overlay zones, protection and conservation, to sites throughout the city to protect natural resources. The Environmental Conservation and Protection zones (or E-Zones) are defined in Title 33, Planning and Zoning, Chapter 33.430 of the Municipal Code (City of Portland 2007). These E-Zones are in place to limit development in "resource areas" that contain significant resources and functional values (values provided by the resources) and the transition areas that buffer them from surrounding pressures. The transition area is defined as the first 25 feet from an E-Zone boundary.

Environmental protection zones provide the highest level of protection for resource areas deemed highly valuable through a detailed inventory and economic, social, environmental, and energy (ESEE) analysis. Development is largely prevented in these areas. Conservation areas are also considered valuable, but can be protected while allowing "environmentally sensitive urban development."

The application of the environmental zones is limited to areas that have undergone a thorough inventory of resources and functional value, in addition to an ESEE analysis. Environmental zoning applies to all development and site disturbance activities. The Columbia River, North Portland Harbor, and Columbia Slough are zoned for conservation (Exhibit 3-14). No lands are designated as a preservation zone. See Exhibit 3-12 for a summary of acreage of this habitat type occurring in the project area.



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3.5.2.2 Metro

Metro provides regional planning initiatives to meet the statewide planning Goal 5: Natural Resources, Scenic and Historic Areas, and Open Spaces. Metro's Nature in Neighborhood initiative is intended to meet Goal 5 requirements. Metro has no enforcement mechanism within local jurisdictions (such as the City of Portland and City of Gresham); therefore, the CRC project has used these habitat definitions to help assess habitat types and importance, rather than as an analysis of regulatory requirements. Exhibit 3-15 summarizes the acreage of each of these habitat types within the project area.

Exhibit 3-15. Summary of Metro Habitat Protections in the Project Area^a

Metro Habitat Protections ³		Acres in Project Area
Riparian Habitat Class I		4992.9
Riparian Habitat Class II		85.8
Riparian Habitat Class III		17.6
Upland Wildlife Habitat Class A		5.7
Upland Wildlife Habitat Class B		0.1
Upland Wildlife Habitat Class C		4.7
	Total:	5106.9

a Includes staging and casting areas, noise attenuation distances, and Ruby Junction (see areas shown on Exhibit 3-16).

Metro adopted a methodology to inventory fish and wildlife habitat and conserve the most highly valued of this habitat. Metro established six classes of habitat inventory for "regionally significant habitat": Riparian Classes I, II, and III, and Upland Wildlife Classes A, B, and C. Highly ranked riparian and upland habitat are identified as "habitat conservation areas" in order to increase protection of these valuable areas from developmental pressures. Generally, these habitat conservation areas are selected based on two criteria: "habitat value or quality..., and urban development value." The regionally significant habitat classes are defined as follows on Metro's website (Metro 2007):

- Riparian class I is of the highest value and includes rivers, streams, wetlands, undeveloped floodplains, forested areas within 100 feet of streams or within 200 feet of streams in steep areas and unique, rare, or at-risk streamside habitats.
- Riparian class II is of moderate value and includes rivers, streams, areas within 50 feet of developed streams, areas with trees and other vegetation within 200 feet of streams, and portions of undeveloped floodplains. These areas provide fewer ecological values than class I areas but are still considered important for stream health.
- Riparian class III is of the lowest value and includes developed floodplains, grassy areas within 300 feet of streams, and small, forested areas that are farther away from streams but still influence them. Many Riparian class III areas are degraded due to development, but still provide some important ecological values and opportunities for restoration.

³ "Impact Areas" include non-habitat areas within 150 feet of stream and wetlands, or within 25 feet of remaining habitat areas. In December 2004, the Metro Council approved a habitat protection concept that integrates urban development priorities and habitat values. Per this approval, development is allowed within the Impact Areas, and they are therefore not included in the table above as an indicator of sensitive habitat.

- Upland wildlife class A is of the highest value and includes very large forested areas and rare or at-risk upland habitats that are farther away from streams, lakes, or wetlands.
- Upland wildlife class B is of moderate value and includes medium-sized and large forested areas that are not rare or at-risk habitats, and non-forested habitat areas that allow wildlife to access water or move from one habitat area to another.
- Upland wildlife class C is of the lowest value and includes smaller forested areas, as well as smaller non-forested areas somewhat near, but no more than 300 feet from, streams and rivers that allow wildlife to move from one area to another.

Exhibit 3-16 summarizes habitats protected by the City of Portland (E-Zones) and Metro.

Riparian class III habitat has been designated on both sides of I-5 south of the Columbia River. In addition, the south bank of the Columbia River is designated Riparian class II.

On the west side of I-5, a portion of the Vanport Wetlands complex is designated as Riparian class I and Upland class A habitat. A small area on the southwest edge of the Marine Drive interchange is designated Riparian class I; and the southern bank of the North Portland Harbor is designated as Riparian class III.

On the east side of I-5, a pond/wetland system between I-5 and Delta Park is designated Riparian class I and class II, and Upland class B. A closed slough system between Delta Park and the Columbia Slough and parallel to I-5, is designated as Upland class C, Riparian classes I, II, and III. A portion of this designated habitat extends to the east side of I-5 as well.

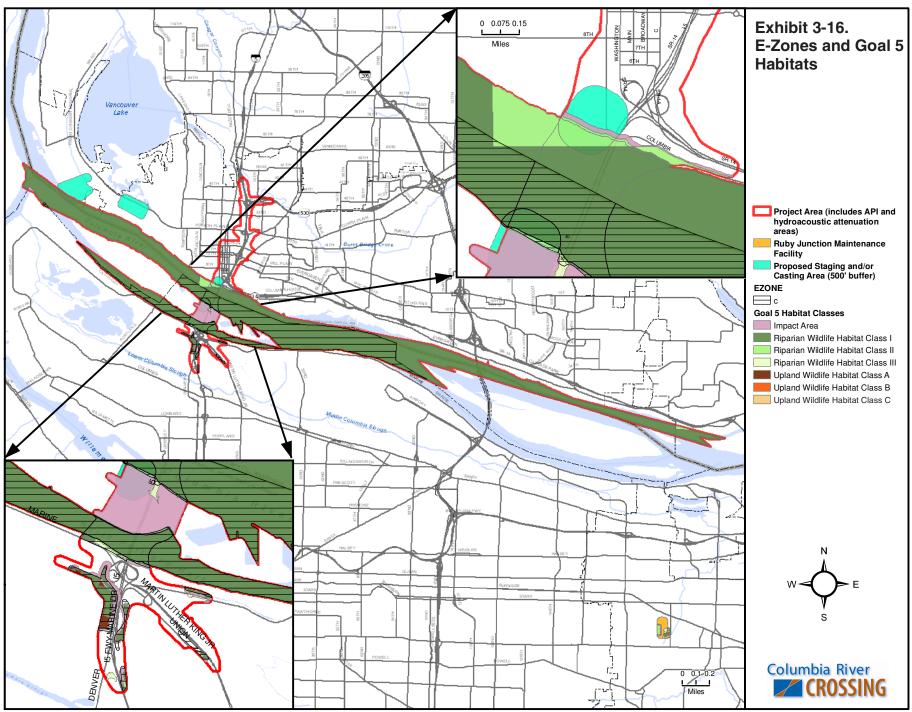
Many of these areas designated as Wildlife Habitat (Upland and Riparian Corridors) are also designated by Metro as habitat conservation areas. Habitat conservation areas are subject to performance standards and BMPs (Metro 2005).

Metro habitat conservation areas are rated as of high, moderate, or low importance. Within the project area, the north shore of the North Portland Harbor is mapped as a low conservation importance area. Various portions of the closed slough system paralleling the east side of I-5 are identified as low and moderate conservation priorities. An open pond/wetland system between I-5 and Delta Park is mapped as a high conservation priority. An area on the southwest side of the Marine Drive interchange and the Vanport Wetlands complex is mapped as a moderate conservation priority.

Metro maps a number of these habitat conservation areas as habitat areas of concern and classifies them as Riparian class I. Within the project area, the southwest side of the Vanport Wetlands complex is designated by Metro as a habitat area of concern.

3.5.2.3 City of Gresham

The City of Gresham inventoried wetlands, riparian areas, and upland areas in the fall of 1987. The findings of this survey are summarized by the Inventory of Significant Natural Resources and Open Spaces that was adopted by the City of Gresham as an appendix to the Community Development Plan (City of Gresham 2005, 2006). This survey was oriented primarily toward wildlife habitat values of lowland and upland natural areas within the City of Gresham. The resource areas included in the inventory are significant wildlife habitats and noteworthy scenic features that perform a variety of useful natural functions, including retention of soils, pollution control, groundwater recharge, and flood control. Forty-five sites having potential significance as natural resource areas were identified within the City of Gresham and include wetlands, riparian corridors, upland areas, and greenways (City of Gresham 2005).



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Among these sites, two sites along Fairview Creek were named as riparian resources: Fairview Creek at SE Burnside Street and Fairview Creek at SE Division Street (City of Gresham 2005). Both of these sites are within the 100-year floodplain of Fairview Creek and are in close proximity to the Ruby Junction Maintenance Facility. Three of the 14 parcels that would be added to the facility during expansion are located within the floodplain.

According to City of Gresham code, within the floodplain overlay district of Fairview Creek, proposed developments need to comply with the guidelines and recommendations of the Fairview Creek master storm drain plan and would need to be accompanied by documentation prepared by a registered civil engineer demonstrating that the development would not result in an increase in floodplain area on other properties, reduce natural flood storage volumes, or result in an increase in erosive velocity of the stream that may cause channel scouring or reduced slope stability downstream of the development (City of Gresham 2009).

3.6 Threatened, Endangered, and Proposed Species

"Listed" species refer to those with federal and/or state threatened, endangered, or proposed status. Data on listed species were obtained from USFWS, ORNHIC, WDNR-NHP, and WDFW-PHS. The bald eagle is a state-listed species in Oregon and Washington (Exhibit 3-17). See Section 3.6.1 for a discussion of peregrine falcons, which have been delisted federally and in Oregon, but retain Sensitive status in Washington. See Section 3.3.2 for a discussion on aquatic species listed as threatened, endangered, or proposed.

Exhibit 3-17. Listed Wildlife Species Known to Occur Within the Project Area

Species Common Name Species Scientific Name ^a	Federal Status ^b	OR State Status ^c	WA State Status ^d	Critical Habitat Present	Habitat Present in Project Area ^e	Habitat Type
Bald eagle Haliaeetus leucocephalus	Delisted	LT	LT	N/A	Yes	Open water; Westside riparian wetlands
Steller sea lion Eumetopias jubatus	LT	LT	LT	No	Yes	Open water

- a Source: ORNHIC (2003).
- b Federal status: LT = Listed Threatened, LE = Listed Endangered, P = Proposed, C = Candidate, SOC = Species of Concern, N/A = Not Applicable (ORNHIC 2003; USFWS 2003).
- c Oregon status: LT = Threatened, LE = Endangered, SC = Sensitive Critical, SV = Sensitive Vulnerable, SP = Sensitive Peripheral, SU = Sensitive Undetermined Status, N/A = Not Applicable (ORNHIC 2003; USFWS 2003).
- d Washington status: LT = Listed Threatened, LE = Listed Endangered, C = Candidate, SS = State Sensitive (WDFW 2008).
- e Source: Project Biologist Observations.

Bald eagles are associated with coastal environments, lakes, rivers, and marshes. They feed primarily on fish but also eat carrion, various water birds, and small mammals. Bald eagles typically nest in tall trees with strong branching structure near large water bodies. Nests are often constructed in the largest tree in a stand with an open view of the surrounding environment. Nest trees are usually near water and have large horizontal limbs. Snags and dead-topped live trees also provide perch and roost sites. Communal roost sites, typically located in treed areas protected from the wind and providing some thermal refugia, are used in the winter for resting. If suitable habitat is present, bald eagles may use urban environments for breeding, feeding, and roosting. In northwestern Oregon and southwestern Washington, the bald eagle breeding season lasts from

January 1 to August 31. Egg-laying takes place mid-February to April, hatching in late March to May, and fledging in late May to mid-August.

Based on a review of aerial and field photographs and topographic maps, viable bald eagle foraging and migration habitat exists within the project area. The Columbia River has sufficient fisheries resources to support bald eagles in the vicinity of the project area. No eagle nests or communal roosts were identified in the project area during the September 2005 field survey, nor during subsequent site visits. However, there are two known nesting sites on Hayden Island and one adjacent to the Columbia Slough, as well as three documented bald eagle nesting territories located within 1 mile of the project area near Vancouver Lake, Smith Lake, and the Columbia River (Isaacs and Anthony 2004; ORNHIC 2007). During limited field surveys in 2003, 2005, 2006, 2007, and 2008, no bald eagles were observed within the project area.

Steller sea lions are usually found in coastal waters near shore and in ocean waters over the continental shelf approximately 35 kilometers off shore, and seasonally up to several hundred kilometers off shore (NatureServe 2007). They feed opportunistically on fish (approximately 10-30 percent of which are salmon), lamprey, squid, and invertebrates (NOAA 2007). Steller sea lions use terrestrial rookeries and haulout locations such as beaches, rocks, jetties, reefs, floating docks, and other structures for breeding, pupping, and resting. They occur year-round at the mouth of the Columbia River, and, in the past, were known to occasionally enter rivers in pursuit of prey. In recent years, adult and subadult male Steller sea lions have regularly been observed at Bonneville Dam, where they prey primarily on white sturgeon and Chinook salmon that congregate below the dam. In 2002, the United States Army Corps of Engineers (USACE) began monitoring seasonal presence, abundance, and predation activities of marine mammals in the Bonneville Dam tailrace (Tackley et al. 2008). Steller sea lions have been documented every year since 2003; the lowest abundance was two Steller sea lions in 2004, and the highest was 53 in 2010 (Stansell et al. 2010).

Steller sea lions use the Columbia River for travel, foraging, and resting as they move between haulout sites and the dam. There are no known haulout sites within the project area. The nearest known haulout in the Columbia River is a rock formation (Phoca Rock) approximately 8 miles downstream of Bonneville Dam (approximately 32 miles upstream of the project area) (Tennis 2009 pers. comm.). Steller sea lions are also known to haul out on the south jetty at the mouth of the Columbia River, near Astoria, Oregon. There are no rookeries located in or near the project area. The nearest Steller sea lion rookery is on the southern Oregon coast at Orford Reef, approximately five miles northwest of Port Orford and more than 200 miles from the project area (NMFS 2008b).

3.6.1 Species of Interest

In addition to species protected by federal and state endangered species regulations, species of interest (SOI) (species which are defined as locally rare or with special habitat requirements) are associated with habitat types in the project area. These include migratory birds, marine mammals, certain terrestrial mammals (e.g., bats), and other species requiring special consideration for habitat and management, but which may not be protected under federal or state statutes. Migratory birds protected under the MBTA use habitat components (e.g., bridge structures, vegetation, riparian habitat) in the project area for nesting, roosting, foraging, and/or dispersing. Impacts to all migratory birds are considered in this report. Exhibit 3-18 lists examples of SOI that may occur in the project area. This list is not meant to be comprehensive but rather presents species groups that require special consideration in the course of the CRC project.

Exhibit 3-18. Examples of Species of Interest Associated with Habitat Types within the Project Area

	Federal Status ^a	OR State Status ^b	WA State Status ^c
Migratory Birds ^d			
Peregrine falcon (Falco peregrinus anatum)	Delisted	SV	S
Purple martin (Progne subis)	SOC	SC	С
Streaked horned lark (Eremophila alpestris strigata)	С	SC	LE
Osprey (Pandion haliaetus)	N/A	N/A	M
Barn owl <i>(Tyto alba)</i>	N/A	N/A	N/A
Belted kingfisher (Ceryle alcyon)	N/A	N/A	N/A
Cliff swallow (Petrochelidon pyrrhonota)	N/A	N/A	N/A
Barn swallow (Hirundo rustica)	N/A	N/A	N/A
Willow flycatcher (Empidonax traillii)	SOC	SU	N/A
Bullock's oriole (Icterus bullockii)	N/A	N/A	N/A
Yellow warbler (Dendroica petechia)	N/A	N/A	N/A
White-breasted nuthatch (Sitta carolinensis)	N/A	N/A	N/A
Great blue heron (Ardea herodias)	N/A	N/A	SM
Loons (<i>Gavia</i> spp.)	N/A	N/A	SS (Gavia immer)
Mergansers (<i>Mergus</i> spp.)	N/A	N/A	N/A
Geese (Branta spp.)	N/A	N/A	N/A
Grebes (Aechmophorus spp.)	N/A	N/A	N/A
Mammals			
Long-legged myotis (Myotis volans)	SOC	SU	M
Fringed myotis (Myotis thysanodes)	SOC	SV	M
Long-eared myotis (Myotis evotis)	SOC	SU	M
Townsend's big-eared bat (Corynorhinus townsendii)	SOC	SC	С
Silver-haired bat (Lasionycteris noctivagans)	SOC	SU	N/A
Harbor seal (<i>Phoca vitulina</i>)	N/A	N/A	M
California myotis (Myotis californicus)	N/A	N/A	N/A
Yuma myotis (Myotis yumanensis)	N/A	N/A	N/A
California sea lion (Zalophus californianus)	N/A ^e	N/A	N/A
Harbor seal (<i>Phoca vitulina</i>)	N/A ^e	N/A	N/A
Little brown myotis (Myotis lucifugus)	N/A	N/A	N/A
Big brown bat (Eptesicus fuscus)	N/A	N/A	N/A
Bushy-tailed woodrat (Neotoma cinerea)	N/A	N/A	N/A
Reptiles and Amphibians			
Western Pond turtle (<i>Emys marmorata</i>)	SOC	SC	LE
Painted turtles (Chrysemys picta)	N/A	SC	N/A
Northern red-legged frog (Rana aurora aurora)	SOC	SV/SU	N/A

a Federal status: C = Candidate, SOC = Species of Concern, N/A = Not Applicable (ORNHIC 2003; USFWS 2003).

b Oregon status: LT = Threatened, LE = Endangered, SC = Sensitive Critical, SV = Sensitive Vulnerable, SU = Sensitive Undetermined Status, N/A = Not Applicable (ORNHIC 2003; USFWS 2003).

c Washington status: LT = Listed Threatened, LE = Listed Endangered, C = Candidate, S = State Sensitive, M = State Monitor (WDFW 2008).

d All migratory birds are protected by the Migratory Bird Treaty Act.

e California sea lions and harbor seals are not federally listed; however, they are protected under the Marine Mammal Protection Act.

Peregrine falcon populations in Oregon and Washington include both resident and migratory populations. Peregrines adapt to a wide variety of nesting locations, including bridges. Their primary nesting locations are cliffs overlooking fairly open areas with ample food. Peregrines are known to feed on a wide variety of species, although birds are their primary food source. Rarely, peregrines feed on bats, squirrels, chipmunks, lizards, fish, and insects. Nests can be found near the coast, in marshes, in mountains, and in urban areas. Breeding occurs only if suitable nesting structures such as bridges, buildings, or cliffs are present (Johnson and O'Neil 2001). Adults remain close to the nest sites throughout the year. In the Portland area, courtship lasts from January to March, eggs are typically laid beginning in mid-March, and fledging occurs late May through late June or July (ODOT 2003).

Peregrine falcons are generally associated with open water, where they feed (Johnson and O'Neil 2001). The Columbia River and adjacent open areas provide sufficient resources to support peregrine falcons in and adjacent to the project area.

Monitoring conducted or funded by ODOT has documented peregrine falcons utilizing habitat in the project area every year since 2001 (Casey 2011).

Streaked horned larks occur adjacent to the project area. Habitat for this species includes bare ground or sparsely vegetated areas, even gravel roadsides. Streaked horned larks nest in grass seed fields, pastures, fallow fields and wetland mudflats. They may occur on any Columbia River beach, and on dredge deposition areas on islands or the mainland along the river.

Bridges within the project area were investigated for evidence of swallow or bat activity (roosting or nesting) in April 2007.,No signs of bat use were observed. Swallows commonly nest on bridges, and likely use the I-5 bridge for this purpose; two remnant mud structures were seen on the south side of the I-5 bridge. No occupied bird nests were found in the surveys. No birds protected under the MBTA were observed using any of the bridges for nesting within the project area.

Canada geese and swallows are known to nest on the concrete piers but are not known to nest on steel structure portions of the bridge.

3.6.2 Wildlife Passage

Due to the highly urbanized nature of the project area, suitable habitat for wildlife passage is fragmented and access to any habitat patches is restricted. I-5 and other arterial roads serve as passage barriers for SOI and urban wildlife. Underpasses, overpasses, and streams serve as potential corridors for crossing I-5. Due to extensive urbanization, the underpasses and overpasses are unsuitable and dangerous corridors for most terrestrial wildlife.

Species most likely to be moving through the project area and utilizing existing terrestrial and aquatic habitat, and which are therefore at risk of collisions with vehicles, are migratory birds (particularly waterfowl such as ducks and geese), and small mammals (e.g., raccoons, squirrels) (Hennings 2007 pers. comm.). A 2005 study of wildlife-vehicle collisions in northwestern Oregon, including the Portland area, did not identify any road kill "hotspots" in or near the project area (MBG 2005).

The Vanport Wetlands and Delta Park provide limited suitable habitat for small and mediumsized terrestrial species, and the habitats are fragmented by I-5. Throughout the remainder of the project area, wildlife corridors and passage opportunities are hindered by the density of urban structures and human disturbance. Passage along the banks of the Columbia River and the North Portland Harbor is possible, although the riparian habitat quality is low and riparian vegetation that could provide cover is sparse. The river bank under the bridges primarily consists of riprap, and is poor habitat for wildlife passage. Potential wildlife habitat and passage corridors exist in some portions of the Delta Park area on the Oregon side of the Columbia River. The river itself is considered a wildlife corridor for waterfowl and some mammals that travel in water, such as river otters and beavers (Hennings 2007 pers. comm.). Areas where terrestrial wildlife could travel under the highway structures between the east and west sides of I-5 include the Victory Boulevard/Whitaker Road area, and the Marine Drive interchange (Thompson 2007 pers. comm.). However, the abundance of roads, traffic, and development makes passage quality marginal at best.

3.7 Botanical Resources

3.7.1 Summary

Listed plant species, including threatened, endangered, proposed, and candidate species, are not known to occur in the project area (ORNHIC 2005; WDNR-NHP 2005). Field visits were conducted on September 1 and September 16, 2005, to survey for potential habitat in the project area. Field surveys for special-status plants (i.e., those not listed but with state designations such as sensitive or vulnerable) occurred between May and September 2006. The surveys used the intuitive controlled method (BLM 1998). No listed plants were found (Parametrix 2005, 2006).

Wapato (*Sagittaria latifolia*) and cattail (*Typha latifolia*), herbaceous wetland plants with important cultural significance as traditional food, craft, and medicinal sources for several Native American tribes, occur in wetland areas in the project area, including Schmeer Slough (a J-shaped slough that extends under I-5 and adjacent to North Whitaker Road and Schmeer Road).

3.7.1.1 Rare Plants

Listed species that have historically occurred within the region include Willamette daisy (Erigeron decumbens var. decumbens), Kincaid's lupine (Lupinus sulphureus ssp. kincaidii), Water howellia (Howellia aquatilis), Bradshaw's lomatium (Lomatium bradshawii), and Nelson's checker-mallow (Sidalcea nelsoniana) (USFWS 2006) (Exhibit 3-19). Willamette daisy and Kincaid's lupine occur in wet prairie, upland prairie, and oak/savannah habitats which were once widely distributed in western Oregon and Washington. Water howellia historically occurred in Multnomah County in small, vernal, freshwater wetlands or in former river oxbows; it is now thought to be extirpated in Oregon. This species occurs in limited distribution in Clark County, Pierce County, and Lincoln County in eastern Washington (WDNR-NHP 2005). Bradshaw's lomatium occurs in Clark County. Nelson's checker-mallow occurs in Oregon ash swales, meadows with wet depressions, or along streams. The species also grows in wetlands within remnant prairie grasslands. Bradshaw's lomatium primarily occurs in seasonally saturated or flooded prairies, adjacent to creeks and small rivers. Habitats associated with tall bugbane (Cimicifuga elata) and small-flowered trillium (Trillium parviflorum) were identified within the project area in Washington, although no instances of these species have been recorded there. Please refer to the Wetlands and Waters Technical Report for more detailed information on wetland plants in the project area.

Exhibit 3-19. Special-Status Plant Species Reported to Occur Within the Project Area

Species	Federal Status	OR Status	WA Status	Habitat Type	Suitable Habitat Present in Project Area ^a
Bristly sedge Carex comosa	N/A	N/A	Sensitive	Marshes, lake shores, wet meadows	No
Columbian watermeal Wolffia columbiana	N/A	N/A	Review Group 1	Freshwater lakes, ponds, slow-moving streams	No
Tall bugbane Cimicifuga elata	SC	С	Sensitive	Mixed coniferous-deciduous forest margins	No
Small-flowered trillium Trillium parviflorum	N/A	N/A	Sensitive	Moist, shady environments dominated by hardwoods	No

Source: ORNHIC 2005 and WDNR-NHP2005.

Habitat suitability for rare plants in the project area in Washington is extremely limited due to severe habitat fragmentation within an urban landscape, degradation by former and/or current land uses, and intense pressure from invasive plant species.

Within the Oregon portion of the project area, rare plant species are most likely to occur at the Vanport Wetlands, which are actively managed for wildlife habitat and wetland function by the Port of Portland. The Vanport Wetlands were not surveyed for rare plants because it was determined that no direct impacts to the site would occur.

3.7.1.2 Noxious Weeds

Small amounts of noxious weeds are found in the project area within most vegetated areas that are not regularly maintained. These include vegetated areas within Washington and Oregon DOT rights-of-way that are infrequently mowed and/or controlled with herbicide applications. Twelve noxious weeds listed by the ODA Noxious Weed Control Program were identified within the project area in Oregon. Fourteen noxious weeds identified by the Washington Department of Agriculture Washington Noxious Weed Control Board (WNWCB) were found within the project area in Washington (Exhibit 3-20). During the preliminary noxious weed survey, no Class A noxious weeds (i.e., those requiring eradication) were identified within the project area.

Exhibit 3-20. Noxious Weed Species Occurring within the Project Area

Botanical Name	Common Name	ODA Status	WNWCB Status
Agropyron repens	Quackgrass	В	N/A
Centaurea pratensis	Meadow knapweed	В	В
Cirsium arvense	Canada thistle	В	С
Cirsium vulgare	Bull thistle	В	С
Clematis vitalba	Old man's beard	В	С
Conium maculatum	Poison hemlock	В	С
Convolvulus arvensis	Field bindweed	В	С
Cytisus scoparius	Scotch broom	В	В
Daucus carota	Wild carrot	N/A	В
Geranium robertianum	Herb-Robert's	N/A	В
Hedera helix	English ivy	В	С

a Parametrix field surveys 2005 and 2006.

Botanical Name	Common Name	ODA Status	WNWCB Status
Hypericum perforatum	St. John's wort	В	С
Phalaris arundinacea	Reed canarygrass	N/A	С
Polygonum cuspidatum	Japanese knotweed	В	В
Rubus discolor	Himalayan blackberry	В	N/A
Verbascum thapsis	Common mullein	N/A	M

Notes:

ODA Key: A = Non-native species of economic importance with a limited distribution or not known to occur in the state; B = Non-native species of economic importance established only in some regions; T = Target A or B Designated weed for which a statewide management plan will be developed and implemented.

WNWCB Key: Class A = Non-native species with a limited distribution in the state – eradication required by state law; Class B = Established only in some regions – control required by state law in regions where the species is unrecorded or with limited distribution; Class C = Widely established in the state or of interest to agriculture – placed on the weed list so that local control is possible; M (Monitor) = Species being monitored for location, spread, and invasiveness.

N/A: Not Applicable indicates that the species does not have a listing status by either ODA or WNWCB.

3.8 Conclusions

3.8.1 Aquatic Resources

The existing I-5 bridges are located at RM 106 of the Columbia River. The Columbia River within the project area is highly altered by human disturbance. Urbanization extends up to the shoreline. There has been extensive removal of historic streamside forests and wetlands. Riparian areas have been further degraded by the construction of dikes and levees and the placement of streambank armoring. The existing bridge structures discharge highway runoff into the river. For several decades, industrial, residential, and upstream agricultural sources have contributed to significant water quality degradation in the river. The river also receives high levels of disturbance in the form of heavy barge traffic.

The 12 major dams located in the Columbia Basin are the dominant forces controlling river flow in the vicinity of the project area. Consequently, the Columbia River upstream of the project area is a highly managed waterway that in some sections (e.g., the reservoirs immediately above the major dams) resembles a series of slack-water lakes rather than its original free-flowing state, although the river is free-flowing in the immediate vicinity of the project area. The Columbia River is also tidally influenced by the Pacific Ocean, which affects flow and stage up to Bonneville Dam above the project area.

Due to the urbanization, industrial use, presence of existing structures, flow control, and channelization of the Columbia River within the vicinity of the project area, the riparian habitat quality within the project area is poor, providing little opportunity for large wood recruitment, nutrient cycling from litter fall, and general fish cover. Several listed fish species occur in the project area, primarily within the Columbia River. Listed fish species include Chinook salmon, steelhead trout, sockeye salmon, coho salmon, chum salmon, green sturgeon, and eulachon. Steller sea lions, California sea lions, and harbor seals also occur in the project area. Water quality is limited in the Columbia River by elevated temperatures, PCBs, PAHs, DDT metabolites (DDE), arsenic, and dissolved copper.

3.8.2 Terrestrial Resources

There are no federally listed terrestrial species that are likely to occur within the project area. One state-listed species, the bald eagle, may use the project area for foraging. Bald eagles are not

known to nest within the project area. The existing I-5 bridge provides potential habitat for bats and swallows. The peregrine falcon, an Oregon and Washington state sensitive species, is known to occur in the project area and would likely be affected by project activities.

The five habitat types identified in the project area (Johnson and O'Neil 2001) are found throughout the region. Priority habitats for this area include the Westside Riparian Wetlands and Herbaceous Wetlands. Metro habitat classification in the project area include riparian and upland wildlife habitat. These habitat types may support SOI such as pond turtles, migratory birds, and small mammals (e.g., bats).

3.8.3 Botanical Resources

No listed plants are documented within the project area. Suitable habitat for rare plants is extremely limited in the project area. The Vanport Wetlands contains the most functional habitat in the project area and is the most likely site for listed species to occur; direct impacts to this habitat are not expected to occur. Noxious weeds are present throughout the project area, although no Class A noxious weeds were detected.

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4. Long-term Effects

This section describes the long-term effects the LPA Option A ("LPA") full build would have to aquatic, terrestrial, and botanical resources. Unless stated otherwise, the LPA with highway phasing options would have the same impacts to aquatic, terrestrial, and botanical resources as the corresponding LPA full build options. Similarly, whether Option A or Option B would be built, the impacts to aquatic, terrestrial, and botanical resources are expected to be the same, except where noted.

The LPA with highway phasing option includes the same bridge pier design as the full LPA options, but because there would be 10.7 fewer acres of impervious surface than the full build, the highway phasing option has slightly fewer stormwater impacts to the Columbia Slough and to Burnt Bridge Creek. The effects analysis for all other long-term impacts remains the same under highway phasing. See sections 4.1.1.4 and 4.1.1.5 for discussions of differences in stormwater effects under the highway phasing option.

4.1 Effects to Aquatic Resources

4.1.1 Stormwater Effects to Water Quality and Water Quantity

Stormwater runoff from highways has elevated levels of contaminants. The project would replace and create impervious surfaces. However, improvements to stormwater treatment within the project CIA, including the I-5 and North Portland Harbor bridges, are anticipated to reduce stormwater pollutant loads discharged to Columbia Slough, Columbia River, North Portland Harbor, and Burnt Bridge Creek.

Besides an existing infiltration pond in the Burnt Bridge Creek drainage, existing water quality facilities would be replaced with stormwater treatment that would meet the project's stormwater management requirements. The majority of new stormwater facilities within the project corridor would provide enhanced treatment.

Much of the current stormwater runoff generated by the existing highway corridor is not treated in accordance with current stormwater treatment standards for new construction. At present, only 21 acres out of the 256 acres of existing impervious area receives treatment via infiltration. Post-project impervious surfaces within the project CIA would be treated in accordance with current stormwater treatment standards before being discharged to project area receiving waters. However, at this time, no options have been identified to treat runoff from about 7 acres of new and resurfaced I-5 impervious surface immediately north of Victory Boulevard within the Columbia Slough watershed and 1 acre comprising the eastbound lanes of SR 14 within the Columbia River watershed in Washington.

Exhibit 4-1 presents an overall summary of the anticipated impact of the project on impervious area and proposed treatment or infiltration. The project currently contains 256 acres of impervious area and would add a net 42 acres, resulting in a post-project net total of 298 acres. The addition of 42 acres may reduce natural infiltration rates and increase stormwater pollutants loads of suspended sediments, nutrients, PAHs, oils and grease, antifreeze from leaks, cadmium and zinc from tire wear, and copper from wear and tear from brake pads, bearings, metal plating, and engine parts. However, untreated impervious surface would be reduced by 228 acres.

The stormwater drainage areas used in these calculations do not include staging areas outside the project footprint or casting yards that might be required for fabricating bridge elements.

Exhibit 4-1. Summary of Changes in Impervious Area and Stormwater Treatment across the Entire Project Corridor

		Area (acr	es)	
	Infiltrated	Treated	Untreated	Total
Existing PGIS	20.5	0.0	218.6	239.1
Existing Non-PGIS	0.0	0.0	17.1	17.1
Existing CIA	20.5	0.0	235.7	256.2
Post-project PGIS				
Existing PGIS retained as-is	15.0	14.1	0.0	29.1
New, rebuilt, or resurfaced PGIS	91.8	137.8	8.1	237.7
Post-project Non-PGIS	4.7	26.2	0.0	30.9
Post-project CIA	111.5	178.1	8.1	297.7
Net change in CIA	91.0	178.1	-227.6	41.5

Traffic models projected to the year 2030 indicate that the project would substantially improve traffic congestion within the project corridor. Decreasing traffic congestion on the I-5 and North Portland Harbor bridges and associated roadways would decrease idling and brake pad wear and may consequently reduce the amount of copper and other traffic-related pollutants currently carried by corridor runoff. However, quantifying the effect of reduced traffic congestion on pollutant loads is not feasible.

Therefore, in comparison to the No-Build Alternative, the LPA would have an overall beneficial effect on stormwater generation and treatment in the long-term due to increased stormwater treatment and decreased traffic congestion.

4.1.1.1 General Effects of Stormwater on Fish

In general, addition of impervious surface to a watershed has the potential to affect fish by altering water quality in the receiving water bodies. Stormwater runoff flows over the roadway, picking up contaminants from impervious surfaces and delivering them to the roadside drainage system and eventually to surface water bodies (Pacific EcoRisk 2007). Sources of these contaminants include vehicles, atmospheric deposition, roadway maintenance, and pavement wear (Pacific EcoRisk 2007).

The addition of impervious surface increases the level of vehicle-generated pollutants deposited on the roadway and delivered to surface waters. Common pollutants present in stormwater runoff include total suspended solids, nutrients, oil and grease, other fluids associated with automobiles, PAHs, agricultural chemicals used in highway maintenance, total zinc, dissolved zinc, total copper, dissolved copper, and other metals (NMFS 2008c). These pollutants are known to be toxic to fish (Everhart et al. 1953; Sprague 1968; Hecht et al. 2007; Sandahl et al. 2007; Johnson et al. 2009; Scholtz et al. 2003) and have potential adverse effects on salmon and steelhead, even at ambient levels (Loge et al. 2006; Hecht et al. 2007; Johnson et al. 2007; Sandahl et al. 2007; Spromberg and Meador 2006, all cited in NMFS 2008c). These contaminants are persistent in the aquatic environment, traveling long distances in solution or adsorbed onto suspended sediments. Alternatively, they may also persist in streambed substrates, mobilizing during high-flow events (Anderson et al. 1996; Alpers et al. 2000a, 2000b, all cited in NMFS 2008c). Some of these

pollutants may also persist in the tissues of juvenile salmonids, resulting in long-term interference with important life functions such as olfaction, immune response, growth, smoltification, hormonal regulation, reproduction, cellular function, and physical development (Fresh et al. 2005, Hecht et al. 2007, LCREP 2007 all cited in NMFS 2008c). The addition of impervious surface may also increase the levels of contamination in surface waters, degrading water quality and causing further harm to fish.

The following sections provide more detail about the types of contaminants found in stormwater runoff and their likely effects on fish.

Contaminant Levels and Effects on Fish

There have been no comprehensive studies performed about the types and concentrations of pollutants found in stormwater runoff emanating from the project area. However, Herrera (2007) prepared a white paper on the types and concentrations of contaminants found in untreated runoff in western Washington, an area with climate and traffic volumes comparable to the project area. No such study exists in Oregon, and therefore, this study represents the most comprehensive review of the characteristics of stormwater runoff applicable to the CRC project area. The study reported that typical contaminants found in stormwater runoff included total suspended solids (TSS), metals, nutrients, and organic compounds. Additionally, stormwater runoff had levels of oxygen demand corresponding to detectable levels of these pollutants.

Geosyntec (2008) performed a comprehensive study of contaminant concentrations in treated stormwater runoff in western Washington. The results of both studies are presented in the subsections below in order to characterize the likely pollutant levels in stormwater runoff in the CRC project area and the risk that fish are exposed to toxic levels of contaminants in the CRC project area.

Total Suspended Solids

TSS has the potential to harm fish by causing gill tissue damage, physiological stress, altered behavior, and degradation of aquatic habitat (Pacific EcoRisk 2007). The level of effect generally depends on the characteristics of the particles, with hard angular particles causing more damage than softer, smoother ones. Given the short-term duration of most precipitation events, exposure of individual fish to such effects would be likely to be limited in space and time (Pacific EcoRisk 2007). However, chronically high levels of TSS may cause long-term degradation of habitat (such as spawning redds) or may reduce the productivity of the benthic communities that make up the food web of numerous fish species.

Herrera (2007) reported mean TSS concentration levels of 93 mg/L in untreated runoff in western Washington, with maximum concentrations of 900 mg/L. Stormwater treatment BMPs reduced TSS levels significantly such that post-treatment median concentration ranged from 6 to 20.5 mg/L (Geosyntec 2008).

There are several criteria for levels of TSS likely to harm aquatic organisms or habitats. Neither Oregon nor Washington offer numeric guidance for TSS. However, EPA guidance classifies impairment to aquatic habitat or organisms as follows:

- < 10 mg/L Impairment is improbable
- < 100 mg/L Potential impairment
- > 100 mg/L Impairment probable.

The National Academy of Sciences (NAS) (1973) offers the following:

- < 25 mg/L High level of protection to aquatic community
- 25–80 mg/L Moderate level of protection
- 80–400 mg/L Low level of protection
- > 400 mg/L Very low level of protection

In the absence of site-specific data about ambient turbidity levels in the receiving water body, the timing and duration of TSS concentrations, and the characteristics of the suspended particles, it is difficult to draw a clear line between TSS concentrations and harm to fish. However, the data show that stormwater treatment facilities significantly reduce TSS concentrations, and, in comparison to the NAS standard, potentially reduce to levels that offer a high level of protection to the aquatic community. In comparison to the EPA threshold, stormwater runoff treatment may reduce TSS concentrations to the low end of the potential impairment standard (Pacific EcoRisk 2007).

Section 5.2.9.4 provides a more detailed review of the effects of suspended sediment on fish.

Metals

The main sources of metals in stormwater runoff include friction in engine and suspension systems, attrition of brake pads and tires, and rust and corrosion of automobile body parts. Other sources include guardrail plating, vehicle emissions, impurities in de-icing compounds, and atmospheric deposition (Herrera 2007). Metals may occur as particulates or dissolved ions (Pacific EcoRisk 2007). Metals in highway runoff are often correlated with levels of suspended sediments because they either occur as particulates or are bound to the surfaces of other solids. Zinc, copper, and chromium show a particularly high correlation with TSS concentrations (Herrera 2007). In general, factors that affect levels of solids in the water column will also affect the levels of metals; however, due to the varied behavior of metals under different environmental conditions, this relationship is very complex (Pacific EcoRisk 2007).

Herrera (2007) reported the following metals in untreated stormwater runoff: antimony, arsenic, barium, cadmium, chromium, cobalt, copper, lead, mercury, molybdenum, nickel, vanadium, and zinc. About half of these (arsenic, antimony, barium, cobalt, molybdenum, nickel, and vanadium) occurred at levels well below any known thresholds for toxicity to aquatic organisms, and therefore, the authors deemed that these metals were not pollutants of concern for fish. Thus, only cadmium, chromium, copper, lead, mercury, and zinc will be addressed further in this discussion.

Cadmium: Herrera (2007) reported median concentrations of 1.2 μg/L in untreated stormwater runoff, with maximum concentrations of 2.80 μg/L. Treated stormwater runoff contained much lower concentrations, with median concentrations ranging from 0.05 to 0.20 μg/L (Geosyntec 2008). Median cadmium levels in treated stormwater were well below freshwater acute criteria. They were also below chronic water quality criteria and EPA Genus Mean Acute Values (GMAVs), that is, values specific to fish genera Oncorhynchus and Salvelinus. However, some of the upper 95th percentile values for treated stormwater exceeded freshwater acute and chronic criteria, indicating that, despite undergoing treatment, stormwater runoff may still contain cadmium at levels that could potentially harm fish (Pacific EcoRisk 2007).

Studies have indicated that chronic levels of cadmium at 0.5 µg/L for 30 days may have sublethal effects on bull trout, including interference with prey selection and prey capture efficiency (Riddell et al. 2005, cited in Pacific EcoRisk 2007). However, this concentration would not likely persist in highway runoff for such an extended period of time (Pacific EcoRisk 2007).

Chromium: Herrera (2007) reported median concentrations of 12.7 μ g/L of total chromium in untreated highway runoff, with maximum concentrations of 17.9 μ g/L. No data were presented for treated highway runoff (Geosyntec 2008). These values were well below the GMAV Cr (III) and Cr (IV) values for *Oncorhynchus* and *Salvelinus*, which ranged from 9,669 to 69,000 μ g/L. The values were also well below the chronic and acute freshwater criteria for Cr (III) (64.4 to 628.6 μ g/L), indicating that stormwater runoff does not contain Cr (III) at levels likely to harm salmonids (Pacific EcoRisk 2007).

Measured maximum values of total chromium did, however, exceed the freshwater acute (15 $\mu g/L$) and chronic criteria (10 $\mu g/L$) for Cr (IV). The measured median concentration is within the acute criterion, but exceeds the chronic criterion. This indicates that while typical chromium levels in untreated stormwater effluent may not cause direct injury or mortality to salmonids, there may be toxic effects on food chain organisms (Pacific EcoRisk 2007).

There were no direct data measuring chromium concentrations in treated runoff. However, it is presumed that levels in treated runoff would be much less than for untreated runoff. While it is reasonable to assume that chromium concentrations in treated runoff would be below levels likely to directly harm salmonids, it is uncertain as to whether concentrations are toxic to food chain organisms (Pacific EcoRisk 2007).

Copper: Herrera (2007) reported median concentrations of 5.18 μ g/L for dissolved copper and 24.4 μ g/L for total copper in untreated stormwater runoff in western Washington. Median concentrations of dissolved copper in treated effluent ranged from 4.4 to 10 μ g/L (Geosyntec 2008). Regardless of whether the samples originated from treated or untreated stormwater, concentrations were in exceedance of freshwater acute criteria, but were below GMAVs for salmon and bull trout (Pacific EcoRisk 2007).

Although dissolved copper concentrations in stormwater runoff may not typically occur at levels likely to cause lethal toxicity to salmonids, sub-lethal toxicity is of great concern. Salmonids may avoid waters with copper concentrations at 2.3 μ g/L (Sprague 1964). Dissolved copper is known to interfere with olfaction in fish, even at very low levels. Reduced olfactory ability interferes with important life functions, such as prey location, predator avoidance, mate recognition, contaminant avoidance, and migration. Baldwin et al. (2003) observed that an increase of 2.3 μ g/L above background levels reduced olfactory response in salmonids by 25 percent. Sandahl et al. (2007) observed 50 percent reduction in olfactory response and 40 percent reduction in predator avoidance when dissolved copper levels were 2.0 μ g/L above background levels of 0.3 μ g/L.

The above data indicate that stormwater runoff contains dissolved copper at levels that may cause sublethal effects in salmonids. However, it is important to note that site-specific conditions, such as the presence of dissolved organic carbon, can reduce the bioavailability of dissolved copper and mitigate for the negative effect on olfaction (Pacific EcoRisk 2007). Therefore, even though a given highway system may discharge dissolved copper at these levels, it is not possible to definitively conclude that this causes harm to fish in every setting (Pacific EcoRisk 2007).

Lead: Herrera (2007) reported median and maximum dissolved lead concentrations at 3.2 μg/L in untreated runoff. BMPs markedly reduced dissolved lead concentrations; median post-treatment concentrations ranged from 0.1 to 2.2 μg/L. Regardless of treatment, dissolved lead levels in runoff were well below acute criteria (16.3 μg/L), indicating that stormwater runoff does not contain dissolved lead at levels likely to kill fish or prey organisms. In some cases, median concentrations for treated runoff exceeded chronic freshwater criteria (0.64 μg/L). However, the authors note that exposure to chronic levels of dissolved lead is unlikely due to the short duration of most runoff events (Pacific EcoRisk 2007).

Lead is also under investigation as a potential endocrine disruptor in fish. Isidori et al. (2007, cited in Pacific EcoRisk 2007) found potential estrogen receptor sensitivity at lead concentrations as low as $0.0004~\mu g/L$. There are no data, however, that provide a direct evidence of actual endocrine disruption in fish at such low levels. The issue warrants more study (Pacific EcoRisk 2007).

Mercury: Herrera (2007) reported median concentrations of 0.02 μg/L for total mercury in untreated stormwater runoff in western Washington. There were no data for mercury concentrations typically found in treated stormwater (Pacific EcoRisk 2007). Total mercury concentrations were well below acute criteria and GMAVs for Hg (II) and were also below acute criteria for total mercury. These values indicate that mercury concentrations in stormwater runoff do not pose a risk to fish or their prey (Pacific EcoRisk 2007). Total mercury did, however, exceed chronic criteria, but Pacific EcoRisk (2007) concludes that chronic exposure to elevated levels of mercury is unlikely.

Organic mercury is of particular concern to fish due to its propensity to bioaccumulate in the aquatic environment. Pacific EcoRisk (2007) caution that it is impossible to extrapolate organic mercury levels or bioaccumulation rates from existing highway runoff sampling data. Nevertheless, the authors note that organic mercury is still an issue for fish, in particular where runoff flows into lentic systems that accumulate organic mercury.

Zinc: Herrera (2007) reported median dissolved zinc concentrations of 39 μ g/L in untreated stormwater (with maximum concentrations of 394 μ g/L). In the same study, median total zinc concentrations in untreated stormwater measured 116 μ g/L (with maximum concentrations of 394 μ g/L). Treated stormwater showed somewhat reduced levels of dissolved zinc, with median concentrations ranging from 7.5 to 41 μ g/L (Geosyntec 2008). All of the dissolved zinc levels, whether for treated or untreated stormwater, were well below GMAVs for salmon and steelhead (931.3 μ g/L) and bull trout (2,100 μ g/L). However, some dissolved zinc concentrations exceeded acute freshwater quality criteria (40 μ g/L) and chronic freshwater criteria (36.5 μ g/L), indicating that direct lethal effects to fish and their prey species may occur after exposure to stormwater runoff, even after it has undergone water quality treatment (Pacific EcoRisk 2007). As with dissolved copper, it is important to note that site-specific conditions may reduce bioavailability of dissolved zinc and mitigate for its toxicity in fish-bearing waters.

Dissolved zinc may also have sublethal effects on salmonids. Sprague (1968) reported that salmonids may avoid waters with zinc concentrations of 5.6 μ g/L above background levels of 3 to 13 μ g/L. Geosyntec (2008) reported that dissolved zinc concentrations in both treated and untreated stormwater exceeded these levels.

Nutrients

Nutrients are chemicals that promote growth in organisms. Nutrients are of concern to fish in that they may cause excessive algal growth in fish-bearing waters, which may in turn reduce dissolved oxygen available to fish or may outcompete food organisms for space in streambed substrate (Pacific EcoRisk 2007). Nutrient levels are not necessarily correlated with traffic levels and may be more closely tied to other land use practices (Pacific EcoRisk 2007). Chief sources of nutrients in highway runoff include atmospheric deposition, vehicle exhaust, and fertilizer applications on the adjacent right-of-way (Herrera 2007). The nutrients of highest concern include nitrogen (in the form of ammonia and nitrate/nitrite) and phosphorous (in the form of orthophosphate and total phosphorous).

Ammonia: Herrera (2007) reported that untreated runoff contained median ammonia concentrations of 1.84 μg/L, with maximum concentrations of 2.66 μg/L. Geosyntec (2008)

reported median ammonia concentrations in treated runoff at significantly lower levels, ranging from 0.03 to 0.08 μg/L. In surface waters, ammonia toxicity is highly variable, depending on ambient pH values; therefore, there is no one numeric acute toxicity criterion for ammonia. Acute toxicity is instead determined by using a complex numeric formula based on ambient pH. Using median highway runoff pH values (Herrera 2007), Pacific EcoRisk (2007) estimates acute toxicity for western Washington waters at 31.26 μg/L. In this case, ammonia found in both treated and untreated runoff is well below the estimated acute toxicity standards, indicating that ammonia levels in highway runoff do not occur at levels likely to kill salmonids.

Stormwater runoff may contain ammonia at levels that could cause sublethal effects to fish. Wicks et al. (2002, as cited in Pacific EcoRisk 2007) found that ammonia at concentrations of 0.02 to 0.08 µg/L may reduce the ability of coho to maintain their highest levels of swimming speed, potentially interfering with upstream migration.

Nitrate/Nitrite: Herrera (2007) reported that untreated runoff contained median nitrate/nitrite concentrations of 1.54 μ g/L, with maximum concentrations of 2.99 μ g/L. In the Geosyntec (2008) study, median concentrations of nitrate/nitrite in treated stormwater ranged from 0.20 to 0.70 μ g/L. Both treated and untreated stormwater runoff has concentrations well below the 96-hour acute toxicity standard of nitrate to salmonids (ranging from 994 to 2342 mg/L). Additionally, levels were well below the 96-hour acute toxicity standard for nitrite (ranging from 110 to 1,700 mg/L). These data indicate that stormwater runoff is not a significant source of nitrate/nitrite in surface water bodies, at least not at levels that are likely to harm fish.

Phosphorus: Herrera (2007) reported that untreated runoff contained median orthophosphate concentrations of 0.10 mg/L, with maximum concentrations of 0.42 mg/L. The same study reported median total phosphorus levels of 0.19 mg/L, with maximum concentrations of 0.57 mg/L. The Geosyntec (2008) study noted that treated stormwater runoff contained median concentrations of 0.04 to 0.26 mg/L. There are no toxicity-based water quality criteria for phosphorus; however, a Pacific EcoRisk (2007) review of the scientific literature concluded that 96-hour exposure to 90 to 1,875 mg/L of di-ammonium phosphate may cause acute harm to certain species of fish (including coho, Chinook, and trout). Given that these standards far exceed levels typically found in both treated and untreated runoff, stormwater does not appear to be a significant source of phosphorus to surface water bodies.

Petroleum Hydrocarbons

This category of pollutants includes vehicle emissions from fuels, such as oil and grease, total petroleum hydrocarbons (TPH), and PAHs. Sources of PAHs include asphalt sealing, vehicle emissions, oils, and atmospheric deposition (Herrera 2007). These contaminants correlate closely with traffic volumes. Additionally, these contaminants have a high affinity for particulates, and therefore they are highly correlated with concentrations of suspended solids. PAHs in streambed sediments have been shown to cause adverse impacts to benthic invertebrates, with potential implications to the prey base of fish (Pacific EcoRisk 2007).

Petroleum hydrocarbons include a large subset of compounds, generally occurring as mixtures of many different chemicals. Accordingly, petroleum hydrocarbons are evaluated in broad groupings such as oil and grease, total PAHs (the sum of numerous individual PAHs), and TPH (the sum of individual petroleum hydrocarbons) (Pacific EcoRisk 2007).

Pacific EcoRisk (2007) examined the Herrera (2007) data regarding PAH concentrations in untreated stormwater runoff and concluded that concentrations of individual PAHs were well below freshwater acute values. This indicates that PAHs from stormwater runoff do not occur at levels that are toxic to fish or their prey base, even when the runoff is untreated. (No data were

presented for treated runoff). For total PAH, the study concluded that median concentrations were well below freshwater acute values, but maximum concentrations were high enough to warrant concern and continued monitoring.

Other studies demonstrate that PAH may cause toxicity in fish embryo-larval life stages (Incardona et al. 2004; Incardona et al. 2005; Incardona et al. 2006, all cited in Pacific EcoRisk 2007); however, no study presents the concentration levels at which this toxicity may occur. Pacific EcoRisk (2007) posits that this type of toxicity may occur at lower levels than the acute toxicity criteria presented above, and therefore this issue warrants further study.

PCBs

PCB use has been banned in the United States since the 1970s (Herrera 2007). However, these compounds are highly persistent, and PCB residues still occur throughout the aquatic environment. PCBs are of particular concern for their propensity to bioaccumulate in fish (Yonge et al. 2002, as cited in Herrera 2007). Sources include atmospheric deposition, pesticides, and herbicides. Few data are available for PCBs concentrations in stormwater runoff. However, they have not been detected in stormwater runoff in western Washington (Zawlocki 1981 as cited in Herrera 2007). Pacific EcoRisk (2007) posits that PCBs are not believed to be a contaminant of concern in highway runoff in western Washington.

Oxygen Demand

Herrera (2007) reported that biological oxygen demand (BOD) median concentrations in untreated runoff were 40.3 mg/L, with maximum concentrations of 71.0 mg/L. For chemical oxygen demand (COD), median concentrations in untreated runoff were 106 mg/L, with maximum levels of 1,377 mg/L.

The State of Washington water quality standards mandate that if a stream has an ambient DO below the water quality criteria, then anthropogenic oxygen demand cannot lower the dissolved oxygen levels by more than 0.2 mg/L. Additionally, the State of Washington offers dissolved oxygen levels necessary for sustaining various salmonid life stages in freshwater, ranging from 6.5 to 9.5 mg/L. Site-specific conditions, such as water flow, turbulence, and ambient temperature, influence the degree to which stormwater runoff with high BOD or COD would result in reduced dissolved oxygen levels in a given surface water body. It is likely that mixing and turbulence in a stream would mitigate the effect of stormwater discharge with high oxygen demand, such that effects would be limited in spatial extent and duration. Nevertheless, Pacific EcoRisk (2007) posits that levels of BOD and COD found in stormwater runoff have the potential to reduce dissolved oxygen in surface water bodies, particularly in warm or lentic water bodies, although it is not possible to predict to what extent. Addition of impervious surface increases the level of vehicle-generated pollutants deposited on the roadway and delivered to surface waters. These pollutants include total suspended solids, nutrients, oil and grease, other fluids associated with automobiles, PAHs, total zinc, dissolved zinc, total copper, dissolved copper, and other metals. These pollutants are known to be toxic to fish (Everhart et al. 1953; Sprague 1968; Hecht et al. 2007; Sandahl et al. 2007; Johnson et al. 2009).

Dissolved copper is of particular concern because it interferes with navigation in fish. Additionally, exposure to dissolved copper, even at extremely low concentrations, causes disruption of the chemical cues that allow juvenile salmonids to avoid predation. These behavioral effects may occur after even brief exposures to low concentrations of dissolved copper; various studies have documented effects at concentrations of 2 μ g/L (Baldwin et al. 2003;

Sandahl et al. 2007) to 5 μ g/L (Hecht et al. 2007). Depending on the exposure concentration and dose period, effects can persist for several weeks.

Factors Affecting Toxicity of Pollutants in Stormwater Runoff

Although stormwater runoff certainly contains contaminants that are known to be toxic to fish, it is difficult to predict what specific concentration levels are likely to cause harm. Water quality criteria are nearly always based on laboratory studies that used purified water to avoid confounding influences from other waterborne contaminants. Accordingly, these results may not reflect site-specific field conditions. Ambient water quality conditions may influence the bioavailability of contaminants, either increasing or decreasing the ability of the contaminant to enter fish tissues. A contaminant concentration that is toxic in one setting may not be toxic in another, depending on the site-specific factors that determine the bioavailability of the contaminant. Similarly, toxicity levels in actual water bodies may be much less than that encountered in a laboratory setting (Pacific EcoRisk 2007).

Suspended solids may bind to chemical contaminants in the water column, reducing their bioavailability to fish. Suspended clay particles have a high capacity for binding, with particular affinity for metals and polar organics (Li et al. 2004, Roberts et al. 2007; Sheng et al. 2002; all cited in Pacific EcoRisk 2007). Thus, presence of clay in the water column may reduce the toxicity of contaminants to fish. On the other hand, silica-based particles (such as sand) have little affinity for such contaminants, and therefore their presence in the water column is not likely to reduce toxicity of chemicals in the water column (Cary et al. 1987, cited in Pacific EcoRisk 2007).

Dissolved organic carbon may have a similar effect, binding to both metals and organics and reducing the potential toxicity of both to aquatic organisms (Newman and Jagoe 1994, cited in Pacific EcoRisk 2007).

Water hardness (particularly concentrations of calcium and magnesium) has an antagonistic relationship with metals, potentially hindering with the uptake of metals into gill tissue (Hollis et al. 2000, cited in Pacific EcoRisk 2007). Interestingly, water hardness does not appear to significantly limit the uptake of copper into fish olfactory tissues (McIntyre et al. 2007, cited in Pacific EcoRisk 2007). On the other hand, water hardness my increase the bioavailability of some PAHs and PCBs (Akkanen and Kukkonen 2001, cited in Pacific EcoRisk 2007).

The pH of water may affect the ionic charge of waterborne contaminants. In general, conditions that promote the ionic form of a contaminant will reduce the contaminant's bioavailability and its toxicity to fish.

Water Quantity

New impervious surface also may also alter water quantity in the receiving water body. In general, addition of impervious surface to a watershed increases the amount of runoff entering surface waters. This may cause changes in stream dynamics, including higher peak flow, reduced peak-flow duration, and more rapid fluctuations in the stream hydrograph. These changes may in turn lead to scour, potentially resulting in impacts to water quality and degradation of stream bed habitat. Increasing the amount of impervious surface also decreases infiltration to groundwater, potentially reducing base flows in streams and decreasing the amount of water available during summer months.

4.1.1.2 General Effects to Aquatic Resources in the CRC Project Area

The project would install numerous stormwater treatment facilities to provide flow control where required and to sequester pollutants before runoff enters any surface water body. It is important to note that even treated stormwater contains some level of pollutants. Most treatment facilities are not 100 percent efficient, and although they greatly reduce pollutant levels, they do not completely eliminate discharges of pollutants to receiving water bodies. Flow-through facilities, in particular, would discharge pollutants during most events. Certain kinds of infiltration facilities have outfalls that discharge untreated stormwater to surface water bodies during events that exceed their design storm.

Treatment would comply with current WSDOT, ODOT, Ecology, and DEQ standards, as well as standards for the cities of Portland and Vancouver (for portions of the project along city-managed roads), before being discharged to project area water bodies. This may result in a net benefit to water quality and water quantity in the project area water bodies during the majority of storm events. Flow control would be provided for runoff discharged to Burnt Bridge and Fairview Creeks; however, flow control is not required for discharges to Columbia Slough, North Portland Harbor or Columbia River.

During events that exceed the design storm for each jurisdiction, stormwater will likely overwhelm treatment facilities, resulting in a release of undertreated or untreated stormwater into project area water bodies. Following these events, fish may be exposed to elevated levels of contaminants. However, the elevated contaminant levels would likely be concentrated around stormwater facility outfalls, would only occur infrequently following large storm events, would be diluted due to the storm event, and would occur only within the first few hours after a storm event when the greatest quantities of contaminants are mobilized from impervious surface into receiving waters (Lee et al. 2004).

The following sections outline the effects to fish species as they occur in each of the project area receiving water bodies.

4.1.1.3 Stormwater Impacts to the Columbia River and North Portland Harbor

Exhibit 4-2 summarizes the treatment scenario for the impervious area within the project corridor that drains to the Columbia River South watershed in Oregon. The project would create approximately 52.8 acres of new, rebuilt, and resurfaced PGIS for LPA Option A and 53.2 acres for Option B. Runoff from 2.2 acres of the existing North Portland Harbor Bridge and 7.6 acres of non-PGIS would be treated prior to being released to North Portland Harbor or the Columbia River. Currently, there are no water quality facilities for runoff from the project footprint in this watershed.

Exhibit 4-2. Summary of Changes to Impervious Area and Stormwater Treatment – Columbia River South (Oregon) Watershed

		Area (acr	es)	
	Infiltrated	Treated	Untreated	Total
Existing PGIS	0.0	0.0	59.4	59.4
Existing Non-PGIS	0.0	0.0	3.0	3.0
Existing CIA	0.0	0.0	62.4	62.4
Post-project PGIS				
Existing PGIS retained as-is	0.0	2.2	0.0	2.2
New, rebuilt, or resurfaced PGIS	0.0	52.8	0.0	52.8

		Area (acr	res)	
	Infiltrated	Treated	Untreated	Total
Post-project Non-PGIS	0.0	7.6	0.0	7.6
Post-project CIA	0.0	62.6	0.0	62.6
Net change in CIA	0.0	62.6	-62.4	0.2

Exhibit 4-3 summarizes the treatment scenario for impervious area that drains to the Columbia River North watershed in Washington. The CIA in this watershed would be increased by approximately 21.1 acres, most of which may be attributed to the reconfigured interchanges and increased number and length of merge lanes for I-5. The project would create approximately 97.8 acres of new and rebuilt PGIS and 13.3 acres of new and rebuilt non-PGIS. In addition, 15.0 acres of existing PGIS, mostly on I-5, would be resurfaced. Water quality facilities are proposed for approximately 134.7 acres of PGIS and 18.3 acres of non-PGIS. In contrast, runoff from only 3.0 acres of PGIS is currently treated.

Exhibit 4-3. Summary of Changes to Impervious Area and Stormwater Treatment – Columbia River North (Washington) Watershed

	Area (acres)			
	Infiltrated	Treated	Untreated	Total
Existing PGIS	3.0	0.0	117.7	120.7
Existing Non-PGIS	0.0	0.0	12.2	12.2
Existing CIA	3.0	0.0	129.9	132.9
Post-project PGIS				
Existing PGIS retained as-is	13.1	9.8	0.0	22.9
New, rebuilt, or resurfaced PGIS	71.3	40.5	1.0	112.8
Post-project Non-PGIS	4.0	14.3	0.0	18.3
Post-project CIA	88.4	64.6	1.0	154.0
Net change in CIA	85.4	64.6	-128.9	21.1

It is difficult to quantify exactly to what extent the treatment scenario would affect water quality in the Columbia River and North Portland Harbor. However, given that the LPA would decrease untreated impervious area by 191 acres in the Columbia River watershed, pollutant loads are anticipated to decrease and the exposure level of fish to these pollutants would likely be lower than current conditions.

Only during events exceeding the design storm would the project likely discharge untreated runoff into the receiving water bodies, potentially resulting in exposure of fish to waterborne pollutants. These watersheds fall under the jurisdiction of the City of Portland, ODOT, and Ecology. For the City of Portland, the design storm is 90 percent of the average annual runoff volume, meaning that, on average, 10 percent of the annual runoff volume would discharge untreated into the receiving water bodies. For ODOT, the design storm is 85 percent of the average annual discharge, meaning that approximately 15 percent of the annual runoff would discharge untreated. In Washington, the design storm is 91 percent of the average annual runoff volume, meaning that 9 percent of the average annual runoff volume would discharge untreated.

Exhibit 4-4 outlines the number of times that a precipitation event typically exceeds the design storms used in areas that drain to the Columbia River and North Portland Harbor. It also

illustrates the percent chance that such events will occur in a given month. Events that exceed the design storm are very likely to occur from September through February, but are also possible during other months. Exceedances are unlikely in July and August. In any case, given the large volume of water in the Columbia River and North Portland Harbor, dilution levels are expected to be very high, and pollutant levels would likely dissipate to background levels within a short distance of the outfalls.

Exhibit 4-4. Frequency and Probability of Design-Storm Event Exceedance for a Given Month (Columbia River and North Portland Harbor)

	City o	f Portland	Ec	ology	C	DOT
Month	No. Events	Probability of Exceedance	No. Events	Probability of Exceedance	No. Events	Probability of Exceedance
Jan	12	14%	19	23%	30	36%
Feb	9	11%	13	16%	22	27%
Mar	1	1%	4	5%	10	12%
Apr	1	1%	1	1%	1	1%
May	0	0%	2	2%	2	2%
Jun	3	4%	4	5%	6	7%
Jul	0	0%	0	0%	0	0%
Aug	0	0%	1	1%	2	2%
Sep	4	5%	7	8%	9	11%
Oct	4	5%	8	10%	11	13%
Nov	18	22%	25	30%	44	53%
Dec	24	29%	44	53%	60	72%

Traffic models projected to 2030 predict that the project would substantially decrease overall traffic congestion on the new bridges and the roadways that contribute runoff to the Columbia River and North Portland Harbor. Idling and brake pad wear, which contribute to the amount of oil, grease, copper, and other pollutants released, would be expected to decrease with congestion relief, as would the amount of pollutants transported to the Columbia River and North Portland Harbor. This may further decrease exposure of fish to pollutants.

Numerous fish species are present in the Columbia River and North Portland Harbor. The following are a subset of species that may be exposed to water quality effects:

- Adult and juvenile Lower Columbia River (LCR) coho; Columbia River (CR) chum;
 Snake River (SR) sockeye; Lower Columbia River, Middle Columbia River (MCR),
 Upper Columbia River (UCR), and Snake River steelhead; and Lower Columbia River,
 Upper Columbia River (Spring-run), Snake River (Fall-run), Snake River (Spring/Summer-run) Chinook salmon;
- Adult and subadult bull trout;
- Adult and subadult green sturgeon;
- All life stages of white sturgeon;
- All life stages of eulachon;
- Adult and juvenile lamprey; and
- Native resident fish, such as sculpins, suckers, threespine sticklebacks, starry flounder, peamouth, and chiselmouth.

These species could be exposed to increased levels of pollutants during the overlap of: 1) when the species are present in the project area and, 2) any event that exceeds the design storm of the treatment facilities (Exhibit 4-4). However, exposure would likely be less than it is currently due to the high level of treatment provided.

USFWS and NMFS have both determined that the Columbia River and North Portland Harbor are "flow-control exempt" water bodies. This means that impervious surface draining to these water bodies does not require flow control facilities. Increases in impervious surface in these watersheds would have no measurable effect on flow.

4.1.1.4 Stormwater Impacts to Columbia Slough

Exhibit 4-5 summarizes the treatment scenario for impervious area that drains to the Columbia Slough watershed. Stormwater outfalls in this watershed discharge directly to Walker Slough and Schmeer Slough. From there, flows are pumped over a levee into the Columbia Slough.

The project would treat or infiltrate 44.5 acres of new and rebuilt PGIS, 2.1 acres of existing PGIS that would be retained, and 4.3 acres of non-PGIS for a net total of 50.9 acres of treated or infiltrated impervious surface. There does not appear to be adequate space between I-5 and Walker Slough to retrofit the existing stormwater conveyance system to treat runoff from approximately 3.7 acres of resurfaced and 3.4 acres of new and rebuilt I-5 PGIS. Therefore, 7.1 acres of impervious area would not receive treatment according to the current design.

Exhibit 4-5. Summary of Changes to Impervious Area and Stormwater Treatment – Columbia Slough Watershed

	Area (acres)			
	Infiltrated	Treated	Untreated	Total
Existing PGIS	3.0	0.0	39.8	42.8
Existing Non-PGIS	0.0	0.0	1.6	1.6
Existing CIA	3.0	0.0	41.4	44.4
Post-project PGIS				
Existing PGIS retained as-is	0.0	2.1	0.0	2.1
New, rebuilt, or resurfaced PGIS	0.0	44.5	7.1	51.6
Post-project Non-PGIS	0.0	4.3	0.0	4.3
Post-project CIA	0.0	50.9	7.1	58.0
Net change in CIA	-3.0	50.9	-34.3	13.6

It is difficult to quantify exactly how the treatment scenario would affect water quality in the Columbia Slough. However, given that the project would treat nearly 400 percent of the net new impervious area in this watershed, it is likely that the stormwater treatment would decrease pollutant loads entering the Columbia Slough, resulting in a net benefit to the environmental baseline during the majority of storm events (i.e., events that do not exceed the design storm). However, there may be a slight increase in dissolved copper pollutant loads due to the addition of impervious area and the lack of stormwater treatment provided for 7.1 untreated PGIS acres. The CRC design team will continue to explore options to provide stormwater treatment for the 7.1 acres of untreated impervious surface. During most events, fish would continue to be exposed to pollutants, but due to increased stormwater treatment they would likely be exposed to lower pollutant levels than current conditions (with the exception of dissolved copper).

Only during events that exceed the design storm would untreated stormwater be discharged into Walker Slough and Schmeer Slough. Exhibit 4-4 depicts the predicted frequency and probability that untreated runoff would enter these sloughs (note the City of Portland and ODOT frequencies). Such events are very likely to occur from September to March, but are also possible during the other months of the year. These events are very unlikely in July and August.

Upon entering Walker and Schmeer Sloughs, stormwater runoff would become diluted at the outfalls. The water would then travel through several thousand feet of vegetated open conveyance, where it would be further diluted in the water column before discharging to Columbia Slough. The diluted runoff would discharge into the Columbia Slough only during periods when the pump is running. (The pump schedule is unknown. This analysis assumes that the pump is continually running in order to provide a worst-case scenario). Because discharge to Walker and Schmeer Sloughs is likely to occur only during larger events (that is, events that exceed the design storm), untreated runoff is likely to become highly diluted by the increased volume of water. Given the high levels of dilution and the large distance between the nearest outfall and the Columbia Slough, it is expected that dilution would reduce pollutant concentrations so they are similar to background levels before this runoff enters fish-bearing waters. Therefore, exposure to fish in Columbia Slough is unlikely.

Traffic models projected to 2030 predict that the project would substantially decrease overall traffic congestion in the areas that drain to the Columbia Slough. Idling and brake pad wear, which contribute to the amount of oil, grease, copper, and other pollutants that are released, are expected to decrease with congestion relief, as would the amount of pollutants transported to the Columbia Slough. This may have a net benefit for fish species using this waterway.

With the exception of bull trout, all of the salmonids addressed by this analysis could potentially use the Columbia Slough for rearing and migration. Of these ESUs/DPSs, the following are likely to be present, based on numerous documented detections: Lower Columbia River Chinook, Upper Willamette River Chinook, Lower Columbia River steelhead, Upper Willamette River steelhead, and Lower Columbia River coho. Other ESUs/DPSs are not documented but are presumed present, given that recent studies have documented up-river (i.e., Middle and Upper Columbia River) ESUs using the Slough and its adjacent floodplain wetlands (Teel et al. 2009). Because the Columbia Slough portion of the project area is accessible to fish, their presence in this area cannot be discounted.

There are no precise data on the times of year that listed salmonids use Columbia Slough. However, they are likely only present from fall through spring, and may to be exposed to water quality effects at any time during this period when there are events that exceed the design storm (Exhibit 4-4). However, as described earlier, exposure is likely to be minimal due to the high level of stormwater treatment and the high levels of in-stream dilution. Exposure during the summer is possible but not likely, because events that exceed the design storm are relatively rare in summer and because water temperatures often exceed levels in which juvenile salmonids can survive (DEQ 2007).

Green sturgeon and eulachon are not known to occur in the Columbia Slough. These species are not likely to be exposed to stormwater effects in the Columbia Slough. However, a recent study showed sturgeon—likely white sturgeon—to be present in the lower Columbia Slough (Van Dyke et al. 2009). Therefore, it is possible that white sturgeon would be exposed to stormwater effects in the Columbia Slough.

Native resident fish, such as sculpins, threespine sticklebacks, and suckers, occur in the Columbia Slough and would be exposed to stormwater effects. There are no precise data on the times of year that these species use the Columbia Slough, although they may be present year-round. These

species may be exposed to water quality effects at any time when there are events that exceed the design storm (Exhibit 4-4). However, exposure is likely to be minimal due to the high level of stormwater treatment and the high levels of in-stream dilution.

No data are available on Pacific lamprey distribution, abundance, timing, or habitat use in the Columbia Slough. However, there is a minimal chance that Pacific lampreys would be exposed to degraded water quality in the Columbia Slough. Stormwater outfalls discharge directly into water bodies that do not contain listed fish, and by association, are unlikely to contain lampreys. Stormwater discharging into these water bodies would travel through several thousand linear feet of a vegetated open conveyance system before entering the Columbia Slough. Given the distance between stormwater outfalls and the nearest locations where fish and lamprey are present, and given the high levels of dilution likely to occur, pollutants would likely dissipate to ambient levels before discharging to fish bearing waters.

Addition of impervious surface to this stormwater drainage area would have no effect on flows in the Columbia Slough. The Columbia Slough is a flow control-exempt water body, meaning that addition of impervious surface in this area is not expected to degrade the flow regime in the Slough, and therefore, the stormwater treatment facilities in this drainage area do not require flow control. Discharges to the Slough are regulated by a Multnomah County Drainage District pump system designed to handle up to the 100-year event. Because the pumps regulate flows between the outfalls and Columbia Slough, additional runoff from these areas would not affect flows in the Slough during the large majority of events, and the inclusion of flow control in treatment facilities would be redundant. Additionally, the tidal influence in Columbia Slough is likely to overwhelm any water quantity impacts occurring during high tides.

LPA with Highway Phasing

The highway phasing option would result in lower levels of pollutants, including dissolved copper, entering the Columbia Slough. Stormwater impacts with highway phasing would be slightly lower because 1) the flyover ramp from eastbound Marine Drive and northbound I-5 would not be initially constructed; and 2) the ramp would be terminated north of Victory Boulevard. These changes would reduce the PGIS within the Columbia Slough watershed by about 5.5 acres (from 53.5 acres to 48.0 acres).

4.1.1.5 Stormwater Impacts to Burnt Bridge Creek

Exhibit 4-6 summarizes the treatment scenario for facilities that drain to the Burnt Bridge Creek watershed. At present, nearly all of the impervious area in this watershed is treated. The project would increase the total impervious area in the watershed by 6.6 acres and would infiltrate 23.1 acres of impervious surface. According to Ecology standards, discharge to Burnt Bridge Creek between 50 percent of the 2-year event and the 50-year event must be reduced to the predevelopment (forested) condition.

Exhibit 4-6. Summary of Changes to Impervious Area and Stormwater Treatment – Burnt Bridge Creek Watershed

		Area (acr	es)	
	Infiltrated	Treated	Untreated	Total
Existing PGIS	14.5	0.0	1.7	16.2
Existing Non-PGIS	0.0	0.0	0.3	0.3
Existing CIA	14.5	0.0	2.0	16.5
Post-project PGIS				

	Area (acres)			
	Infiltrated	Treated	Untreated	Total
Existing PGIS retained as-is	1.9	0.0	0.0	1.9
New, rebuilt, or resurfaced PGIS	20.5	0.0	0.0	20.5
Post-project Non-PGIS	0.7	0.0	0.0	0.7
Post-project CIA	23.1	0.0	0.0	23.1
Net change in CIA	8.6	0.0	-2.0	6.6

It is difficult to quantify whether the enhanced proportion of infiltration would outweigh the impacts associated with the net new impervious area. However, given that the project would provide treatment or infiltration for more than 300 percent of the net new impervious area in this watershed, it is possible that the improved treatment scenario would result in a net benefit to the environmental baseline and to fish in Burnt Bridge Creek during events that do not exceed the design storm. In any case, the project is not likely to significantly degrade conditions in the creek during events less than the design storm.

During events that exceed the design storm, however, untreated runoff would certainly enter Burnt Bridge Creek. Exhibit 4-7 depicts the estimated frequency and probability of events that would exceed the design storm. These types of events are most likely to occur from November through February, but may also occasionally occur during the rest of the year. Discharge during May, July, and August is highly unlikely. However, given the high level of infiltration in this drainage area, actual discharge of untreated stormwater is expected to occur less often than predicted in Exhibit 4-7. Additionally, pollutants would likely be diluted due to the large volume of water that typically is present during these events. Although fish may be exposed to untreated stormwater during events that exceed the design storm, exposure would likely be less than it is currently due to the high level of treatment proposed.

Exhibit 4-7. Frequency and Probability of Design-storm Event Exceedance – Burnt Bridge Creek

	91% Desig	gn Volume
Month	No. Events	Probability of Exceedance
Jan	12	14%
Feb	9	11%
Mar	1	1%
Apr	1	1%
May	0	0%
Jun	3	4%
Jul	0	0%
Aug	0	0%
Sep	4	5%
Oct	4	5%
Nov	18	22%
Dec	24	29%

During events that exceed the design storm, stormwater runoff may also degrade the flow regime in Burnt Bridge Creek. However, due to the high levels of infiltration proposed, impacts are expected to be slight.

All freshwater life stages of coho, Chinook, and steelhead are potentially present in the creek (Weinheimer 2007 pers. comm.). Therefore, runoff may affect all life stages, as well as spawning, migration, foraging, and rearing habitat. The abundance of these species is thought to be very low in Burnt Bridge Creek (PSMFC 2003). Therefore, it is expected that very few individuals would be exposed to these effects. Steelhead and coho have been detected in Burnt Bridge Creek in proximity to stormwater outfalls, and exposure of these species to stormwater effects is likely. Chinook have been detected in Burnt Bridge Creek within 1 mile of the project-area stormwater outfalls. However, because abundance of Chinook is very low and there is a partial passage barrier between the location of the detection and the nearest project-area outfall, the likelihood of exposure is discountable.

Lower Columbia River coho, Chinook, and steelhead could be exposed to stormwater runoff during events that exceed the design storm. Exposure is likely from fall through spring, when design storm-exceeding events most frequently occur and when these species have been documented in the stream. Due to the limited data on fish presence, there are no precise dates for when these species occur in Burnt Bridge Creek. There are only two known stream surveys in Burnt Bridge Creek, conducted in November/December 2002 and April 2003 (PSMFC 2003). The results of the surveys indicate that these species are at least present from November through April. They presumably occur there at all times of year except during the warmest summer months.

During summer, exposure is possible, but less likely. Given the lack of data, we cannot discount the possibility that fish occur there during the summer. However, the Washington 303(d) list has documented water temperatures that exceed the range tolerated by salmonids during some summers (Ecology 2009b). Therefore, these species may not be present in Burnt Bridge Creek in the summer, at least not during some years. Additionally, events exceeding the design storm are less likely in the summer, further reducing the likelihood for exposure.

Lamprey ammocoetes have been documented in Burnt Bridge Creek (PBS 2003); however, no comprehensive data are available on Pacific lamprey abundance, timing, or habitat use in Burnt Bridge Creek. Lampreys may be exposed to degraded water quality and flow regime during periods when lamprey are present and when there is an event that exceeds the design storm.

Native resident fish, such as dace, threespine stickleback, redside shiners, suckers, and sculpin, are present in Burnt Bridge Creek (PBS 2003), and may be exposed to degraded water quality and flow regime when there is an event that exceeds the design storm.

Other salmonid ESUs/DPSs, eulachon, and green sturgeon are not present in Burnt Bridge Creek and would not be affected by stormwater runoff in this stream.

LPA with Highway Phasing

Stormwater impacts to Burnt Bridge Creek would be lower with highway phasing because there would be no improvements to I-5 itself north of 39th Street. Phasing of this highway construction would result in a reduction in impervious area of approximately 5.2 acres, all of which is in the Burnt Bridge Creek watershed.

4.1.1.6 Ruby Junction

The expansion of the Ruby Junction maintenance facility would result in a slight decrease of impervious area (0.5 acre). Since the City of Gresham's requirements for stormwater treatment and flow control must be met for this portion of the project, runoff from all impervious areas within the expansion area would be infiltrated to reduce pollutants of concern. The infiltration

techniques would comply with the City of Gresham stormwater management requirements and would protect and/or improve the quality and quantity of existing groundwater flows. There would be no discharge to any surface water body at any time. During events that exceed the design storm, stormwater would pond in a nearby field adjacent to the treatment facility. Because there would be no discharge to any surface water body, this element of the project would have no effect on fish or on the water quality of Fairview Creek.

4.1.1.7 Summary of Stormwater Effects to Fish

The project would provide a high level of treatment for a large proportion of the project CIA, installing treatment not just for new PGIS but also for 228 acres of impervious area that is currently untreated. Project-wide, there would be treatment for about seven times the area of net new impervious area. While the project would not completely eliminate effects to water quality and flow, the high level of treatment would be expected to provide an overall benefit to current conditions. Effects to individual species are summarized below.

Bull trout adults and subadults could potentially be exposed to degraded water quality in the Columbia River and North Portland Harbor. However, given the very low abundance of bull trout and high levels of dilution in these water bodies, the likelihood of exposure is insignificant.

Green sturgeon adults and subadults could also be exposed to degraded water quality in the Columbia River and North Portland Harbor. However, given the high levels of dilution, exposure is expected to be insignificant. Due to the rarity of green sturgeon in the areas subjected to diminished water quality, the likelihood of exposure is discountable.

Stormwater effects to fish species are as follows:

- In the Columbia River and North Portland Harbor, listed salmon and steelhead, bull trout, green sturgeon, white sturgeon, eulachon, lamprey, and other native resident fish (e.g., sculpins, threespine sticklebacks, suckers, and dace) may be exposed to degraded water quality within a short distance of the outfalls during periods when fish are present and when there is an event that exceeds the design storm (Exhibit 4-4). Exposure would be minimal due to the high dilution capacity of these large water bodies. During events that do not exceed the design storm, the project is expected to discharge runoff that has less pollutant content than the pre-project condition due to the high level of stormwater treatment relative to the net new PGIS. While it is inconclusive whether this constitutes a benefit to these fish, the high level of treatment makes it improbable that the runoff would degrade current levels of water quality or cause higher levels of exposure during these events.
- In the Columbia Slough, there is a minimal chance that listed salmonids, white sturgeon, lamprey, and othert native resident fish (e.g., sculpins, threespine sticklebacks, and suckers) would be exposed to degraded water quality. Stormwater outfalls discharge directly into water bodies that do not contain sensitive fish species and travel through several thousand linear feet of a vegetated open conveyance system before entering the Columbia Slough. Given the distance between stormwater outfalls and the nearest locations where sensitive fish species are present, and given the high levels of dilution likely to occur, pollutants would likely dissipate to ambient levels before discharging to fish-bearing waters.
- In Burnt Bridge Creek, LCR coho, steelhead, Chinook, lamprey, and othert native resident fish (e.g., sculpins, threespine sticklebacks, suckers, and dace) may be exposed to degraded water quality and flow regime during periods when fish are present (fall through spring) and when there is an event that exceeds the design storm (Exhibit 4-7).

Due to the low abundance of these species in Burnt Bridge Creek, few individuals would be exposed to these effects. Steelhead and coho would be likely to experience exposure to these effects, as they have been detected in proximity to stormwater outfalls associated with this project. For Chinook, exposure would be insignificant, as they have been detected more than a mile from the nearest outfall and downstream of a partial passage barrier.

4.1.2 Effects to Aquatic Habitat

Potential project impacts to aquatic habitat would be manifested in effects to both shallow- and deep-water habitats. These effects are discussed below.

4.1.2.1 Shallow-water Habitat

The project would have permanent impacts on shallow-water habitat in the Columbia River and North Portland Harbor, including the addition of in-water and overwater bridge elements and the removal of existing in-water and overwater structures. For analysis purposes, the project defined shallow water as that water 20 feet or less in depth (approximate depth from observed lowest water [0'CRD]).

This section outlines the role of shallow-water habitat in the life history of fish and provides an analysis of the project's likely effects on fish in shallow-water habitat in the CRC project area.

Fish Distribution in Shallow-Water Habitat

Shallow water is of particular importance in the life history of fish for migration, feeding, holding, rearing, and predator avoidance (Simenstad et al. 1982; Spence et al. 1996; Everhart et al. 1953). Lower Columbia River Chinook and Columbia River chum migrate as subyearlings and are particularly dependent on nearshore, shallow-water areas during outmigration (Levy and Northcote 1982, Myers and Horton 1982, Simenstad et al. 1982, and Levings et al. 1986 as cited in Bottom et al. 2005). Typically, these fish are less than less than 50 to 60 millimeters fork length and primarily use water that is less than 1 meter deep (Bottom et al. 2005). Numerous studies have documented smaller fish (subyearling Chinook) utilizing nearshore habitats (Johnsen and Sims 1973; Dawley et al. 1986; McCabe et al. 1986; Ledgerwood et al. 1991, as cited in Carter et al. 2009), frequently at depths of 3 meters or less (Carlson et al. 2001, as cited in Carter et al. 2009). However, Lower Columbia River Chinook and Columbia River chum can and do occupy other parts of the channel (Bottom et al. 2005). While these fish are highly dependent on shallow water and are most likely to occur there, they do not occur exclusively in the nearshore and may potentially be present across the entire cross-section of the channel (Bottom et al. 2005).

Other juvenile salmonids outmigrate after they reach the yearling stage or older. These species include all of the salmonid runs addressed by this analysis except for chum (note that Lower Columbia River Chinook may emigrate as either subyearlings or as yearlings). In general, cross-sectional distribution of these larger juveniles in the stream channel appears to be correlated with size. Fish measuring 60 to 100 millimeters fork length use deeper water, such as shoals and distributary channels. Fish greater than 100 millimeters in length are found in both deep and shallow-water habitats, indicating that these individuals do not show preferential use of a particular water depth (Bottom et al. 2005), although they may seek out these areas for resting or as flow refugia during high-velocity events. Fish that migrate as yearlings or older tend to move quickly and occupy deeper-water habitats, but it is well documented that all use the nearshore to some extent during their outmigration (Bottom et al. 2005; NMFS 2006; Celedonia et al. 2008; Southard et al. 2006; Friesen 2005; Carter et al. 2009). These juveniles may alternate active

migration in deeper water interspersed with periods of holding and resting in shallow water and/or low-velocity areas (Bottom et al. 2005; Celedonia et al. 2008). Thus, while these older juveniles are less dependent on the nearshore than their subyearling migrant counterparts, they are likely to be present across the entire cross-section of the channel (Bottom et al. 2005; Southard et al. 2006).

Rearing juveniles are largely dependent on shallow-water habitats (Bottom et al. 2005; Southard et al. 2006; NMFS 2006). Listed ESUs that rear in the project area include Lower Columbia River Chinook, Upper Columbia River spring-run Chinook, Upper Willamette River Chinook, Columbia River chum, Lower Columbia River coho, and Lower Columbia River steelhead.

Adult salmonids generally migrate at mid-channel, but may occupy depths of 1 to 50 feet (NMFS 2006). While they may occur in shallow-water habitat, they are not particularly dependent on it, although they may seek out these areas for resting or as flow refugia during upstream migration (Bottom et al. 2005).

Similar to the salmonids discussed above, bull trout subadults and adults may use shallow-water habitat for migration and holding. Bull trout are thought to occur in extremely low numbers in this portion of the project area; therefore, effects to shallow-water habitat and bull trout are expected be insignificant.

Adult green sturgeon and subadult and juvenile white sturgeon may use shallow-water habitat for feeding and migration, although they tend to be less dependent on shallow-water areas and are often found holding in fairly deep holes. Green sturgeon are thought to occur in extremely low numbers in this portion of the project area; therefore, effects to shallow-water habitat and green sturgeon are expected to be insignificant.

Eulachon use shallow-water habitat during all life stages: feeding, spawning, and migration (Langness 2009 pers. comm.).

White sturgeon utilize shallow-water habitat during periods of high activity (i.e., summer months), and deep water during the winter (Brannon and Sutter 1992). In the Columbia River, adult white sturgeon have been observed at a mean water depth of 36 feet (11m) (Counihan et al. 1999), although they are also known to utilize habitat in the Columbia River of less than 23 ft (7 m) in depth (Parsley et al. 1993). Adult and subadult white sturgeon are primarily benthic feeders, taking prey such as crabs, clams, and shrimp, and are likely to use shallow water for foraging (Moyle 2002). In the lower Columbia River, most spawning was observed at depths of 19 feet (6 m) (Parsley et al. 1993). Juvenile white sturgeon prefer deep-water habitat and are often observed in the deepest part of the channel; however, they have been observed in water as shallow as 6 feet (Parsley et al. 1993).

Shallow-water habitat is used by lamprey for spawning, rearing, and migration (Ocker et al. 2001).

Native resident fish such as sculpins, threespine sticklebacks, dace, and suckers spend the majority of their life cycle in shallow-water habitat, utilizing emergent vegetation for cover, spawning, and foraging. Because their life history requirements include the use of emergent vegetation and other types of cover (e.g., rocks and overhanging trees), their distribution even within shallow-water areas is relatively limited to depths of only a few feet where emergent vegetation is present. These fish species typically forage on prey items (e.g., benthic invertebrates, algae, and detritus) that also depend on emergent vegetation, or at least are present at depths at which primary productivity is high. These species may migrate locally among habitat

areas in the project area in response to seasonal flows, water temperatures, life stage, and temporal cycles (e.g., moving between various depths according to time of day).

There would be no long-term impacts to shallow-water habitat for California or Steller sea lions.

Physical Loss of Shallow-water Habitat

Several new bridge piers would be located in water of 20 feet depth or less (Exhibit 4-8). The inwater portions of the new structures would result in the permanent physical loss of approximately 250 square feet of shallow-water habitat at pier complex 7 (Pier 7) in the Columbia River. Demolition of the existing Columbia River structures would permanently restore about 6,000 square feet of shallow-water habitat. Additional shallow-water habitat would be restored by removal of a portion of the large overwater structure at the Quay, although the area cannot be quantified at this stage in the design. Overall, there would be a net permanent gain of at least 5,945 square feet of shallow-water habitat in the Columbia River (Exhibit 4-9). In North Portland Harbor, there would be a permanent net loss of about 2,435 square feet of shallow-water habitat associated with the new in-water bridge bents (Exhibit 4-10). Note that all North Portland Harbor impacts are in shallow water.

Physical loss of shallow-water habitat is of particular concern for rearing or subyearling migrant salmonids. In general, in-water structures that completely block the nearshore may force these juveniles to swim into deeper-water habitats to circumvent them. Deep-water areas generally represent lower quality habitat because predation rates may be higher there. Numerous studies show that predators such as walleye and northern pikeminnow occur in deep-water habitat for at least part of the year (Johnson 1969; Ager 1976; Paragamian 1989; Wahl 1995; Pribyl et al. 2004). In the case of the CRC project, in-water portions of the structures would not pose a complete blockage to nearshore movement anywhere in the project area. Although these structures would cover potential rearing and nearshore migration areas, the habitat is not rare and is not of particularly high quality. These juveniles would still be able to use the abundant shallow-water habitat available for miles in either direction. Neither the permanent nor the temporary structures would force these juveniles into deeper water, and therefore pose no added risk of predation. Additionally, northern pikeminnow and walleye tend to avoid high-velocity areas during the spring juvenile salmonid outmigration (NMFS 2000; Gray and Rondorf 1986; Pribyl et al. 2004). The high velocities present in deep-water portions of the CRC project area may limit the potential for actual predation in deep-water areas.

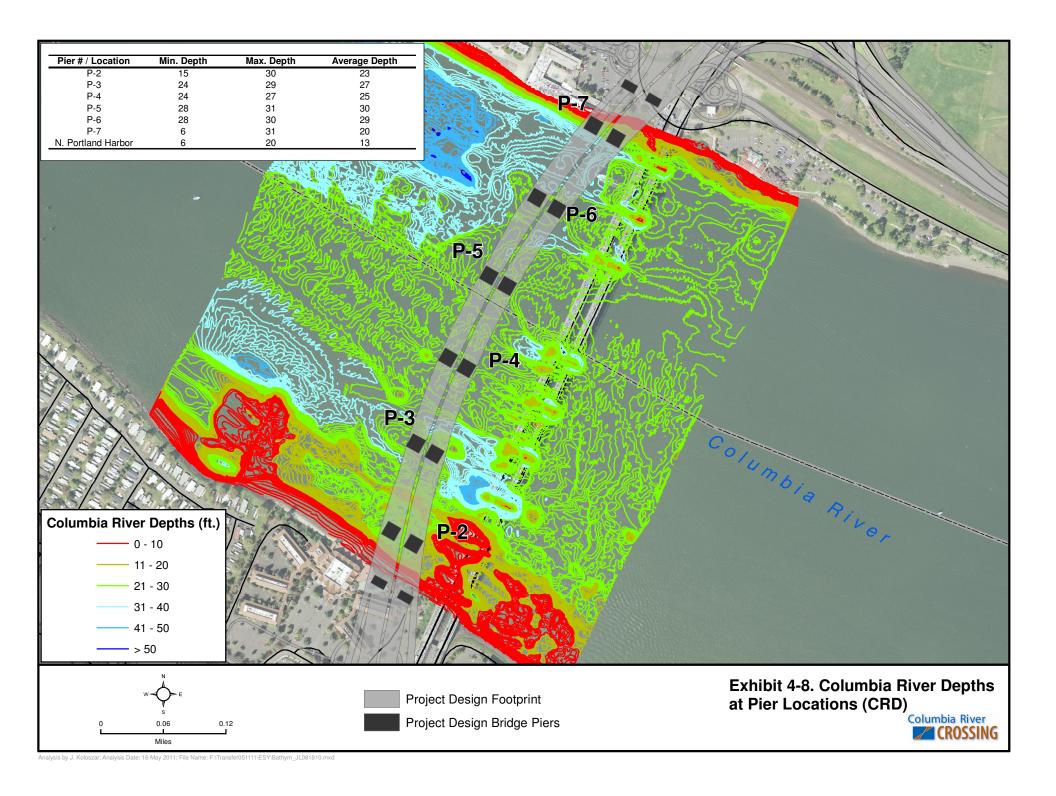


Exhibit 4-9. Physical Impacts to Shallow-water Habitat in the Columbia River

	Columbia River	
Structure	Area	Time in Water
Permanent		
New Bridge Shafts (2 Drilled Shafts @ Pier 7)	236 sq. ft.	Permanent
Existing bridges piers to be removed (existing Pier 10, 11)	- 6,181 sq. ft.	Permanent
Total Permanent Impact	- 5,945 sq. ft. ^a	

a This total does not include square footage of existing piers to be removed at the Red Lion at the Quay because that figure cannot be quantified at this stage of design; therefore, this total likely underestimates the amount of shallowwater habitat that will be restored via removal of in-water structure.

Exhibit 4-10. Physical Impacts to Shallow-water Habitat in North Portland Harbor

		North Portland Harbor	
Structure		Area	Time in Water
Permanent			
New Bridge Piers (31 columns)		2,435 sq. ft.	Permanent
Total	Permanent Impact	2,435 sq. ft.	

Increase in Overwater Coverage

The project would place several overwater structures in shallow water in the Columbia River and North Portland Harbor; however, permanent overwater structures likely to have effects on fish include only the shaft caps on the Columbia River bridges. Exhibit 4-11 quantifies the area and duration of project-related overwater structures in the project area.

Effects of overwater coverage on fish and fish habitat are discussed in Section 5.3.2.

Exhibit 4-11. Overwater Coverage in Shallow-water Habitat in the Columbia River

	Columbia River	
Structure Type	Area	Duration in Water
Shaft Caps (P7 – Half of SB)	1,688 sq. ft.	Permanent
Pier at Red Lion at the Quay to be Removed	-18,965 sq. ft.	Permanent
Total Permanent Impact	-17,277 sq. ft.	

4.1.2.2 Deep-water Habitat

Deep-water habitat occurs only in the Columbia River portion of the project area. Aquatic SOI have mixed use of this deep-water habitat. Typically, subyearling migrant salmonids are restricted to shallow-water habitat in the upper estuary (including the project area) (Carter et al. 2009); however, the possibility exists that some will occasionally stray into the surface layer of deeper waters (Bottom et al. 2005). Larger juvenile salmonid migrants commonly use deep-water portions of the navigation channel in high numbers during outmigration, taking advantage of higher velocities there (Carter et al. 2009).

Adult salmonids do not show any specific preference for deep-water habitat over shallow-water habitat (Bottom et al. 2005). While they generally migrate at mid-channel, they may be found at depths of 1 to 50 feet (NMFS 2006). They commonly use deep-water portions of the project area for foraging and hold in low-velocity areas of deep-water habitat (such as behind bridge piers).

Eulachon adults and juveniles are known to forage at depths of greater than 50 feet and are likely to be present in deep-water portions of the project area (Hay and McCarter 2000).

Adult and subadult green sturgeon use waters at a depth of 30 feet or less and also could be present in deep-water portions of the project area (73 FR 52084).

Data on Steller sea lion, California sea lion, and harbor seal use of the water column within the Columbia River during transiting are not available. However, these species utilize marine habitat of significantly greater depths of 20 feet, and it is reasonable to assume that they would utilize deep water habitat of the Columbia River for transiting.

As discussed in the previous section on shallow-water habitat, white sturgeon are known to utilize both shallow and deep water habitat in the Columbia River. Adult white sturgeon have been observed in waters approximately 7 to 98 feet (2 meters to 30 meters) in depth (Counihan et al. 1999) and are likely to use deep water habitat for foraging, resting, breeding, and spawning (Moyle 2002). White sturgeon may spawn in the lower Columbia River in depths of up to 75 feet (23 m) (Parsley et al. 1993). Juvenile white sturgeon in the Columbia River system have been documented in median water depths of 52 to 62 feet (16 to 19 m) (Parsley et al. 1993). White sturgeon may be more likely to use deep-water habitat during the winter months, often congregating in deep holes in the Columbia River (Brannon et al. 1992).

Other native fish resident to the Columbia River that may be present in deep water habitat include northern pikeminnow (*Ptychocheilus oregonensis*).

The project would have permanent impacts on deep-water habitat in the Columbia River, including physical gain of habitat area and an increase in overwater coverage.

Impacts to deep-water habitat would affect the following species and life stages:

- Feeding, holding and migration habitat for juveniles and holding and migration habitat
 for adults of the following ESUs/DPSs: LCR coho; CR chum; SR sockeye; LCR, MCR,
 UCR, and SR steelhead; and LCR, UCR spring-run, SR fall-run, and SR spring/summerrun Chinook.
- Rearing habitat for juvenile Chinook (LCR, UCR spring-run, and UWR), LCR coho, LCR steelhead, and CR chum.
- Adult and subadult bull trout migration and holding habitat. (Because of the extremely
 low numbers of bull trout in this portion of the project area, risk of exposure to this effect
 is discountable).
- Adult and subadult green sturgeon feeding and migration habitat. (Because of the
 extremely low numbers of green sturgeon in this portion of the project area, risk of
 exposure to this effect is discountable).
- All life stages of white sturgeon.
- Adult and larval eulachon spawning and migration habitat.
- Lamprey rearing and migration habitat.

Physical Gain of Deep-water Habitat

Exhibit 4-12 summarizes physical impacts to deep-water habitat in the Columbia River.

Exhibit 4-12. Physical Impacts to Deep-water Habitat in the Columbia River

	Columbia River	
Impact	Area	Time in Water
Permanent		
New bridge drilled shafts (P2 - P7)	6,361 sq. ft.	permanent
Existing bridges piers to be removed (existing Piers 2 – 9)	- 21,633 sq. ft.	permanent
Total Permanent Impact	-15,272 sq. ft.	

The in-water portions of the new structures would result in the permanent physical loss of approximately 6,300 square feet of deep-water habitat at pier complexes 2 through 7 in the Columbia River. Demolition of the existing Columbia River piers would permanently restore about 21,000 square feet of deep-water habitat. Overall, there would be a net permanent gain of about 15,000 square feet of deep-water habitat in the Columbia River.

Increase in Overwater Coverage

The project would place several overwater structures in deep-water portions of the Columbia River. The only permanent new overwater structures likely to have effects on fish would be the shaft caps on the Columbia River bridges. Exhibit 4-13 quantifies the area and duration of project-related overwater structures in deep-water portions of the project area.

Exhibit 4-13. Overwater Coverage in Deep-water Habitat in the Columbia River

Туре	Area	Duration in Water (days)
Permanent		
Shaft caps (P3 – P6)	56,813 sq. ft.	Permanent
Total permanent impact	56,813 sq. ft.	

Overwater coverage may create dense shade that may potentially attract predators. The sharp dark-light interface found underneath overwater structures may also cause visual disorientation to juvenile fish, which may in turn result in delayed migration and increased vulnerability to predators. Of the juvenile fish that use the project area, rearing juveniles and subyearling-migrant salmonids are highly dependent on shallow-water habitat and therefore are less vulnerable to these effects in deep water. However, as these individuals are not restricted to the nearshore (Bottom et al. 2005), they may stray or be carried into deeper water with the current, and there is a small chance of exposure to these effects. Larger juveniles of the yearling age class or older commonly use deep-water habitat during migration, and therefore are likely to be exposed to these effects.

The existing and proposed bridge spans in the Columbia River are more than 30 feet above the water surface and are therefore not likely to create dense shade on the water surface. For this reason, shade cast by these structures is unlikely to affect fish.

The shaft caps of the proposed Columbia River structures are at the water line and could create a net gain of permanent new dense shade (approximately 57,000 sq. ft.) in deep water.

The permanent structures would not create a swath of dense shade completely spanning deep-water habitat. Therefore, even if these structures were to create a shadow line that juvenile salmonids avoid crossing during daylight hours, juveniles could simply circumvent the shadow, resulting in no measurable delay to migration. Nighttime migration would be unaffected. Larval eulachon do not have volitional movement and are therefore not subject to visual disorientation or migration delays. Both adult and juvenile lamprey migrate primarily at night and are unlikely to be affected by shade.

The increase in the shade footprint increases the amount of suitable habitat for predators and therefore could presumably increase the number of predators in this portion of the project area. This could potentially cause a temporary and/or permanent increase in predation rates on juvenile salmonids, although it is not possible to quantify the extent of this effect.

Although it is impossible to quantify the extent to which increased shade may affect predation rates or cause visual disorientation in juveniles, it is possible to estimate the physical extent and duration of the effect. This effect would occur both when the structures are present in the water and during the timing of juvenile fish presence in this portion of the project area.

4.1.3 Hydraulic Shadowing

The modeling for the Columbia River bridges uses an earlier design with three sets of bridge piers with up to 12 drilled shafts each. The proposed design now consists of only two sets of piers, with only nine drilled shafts per pier. At present, the design team has not yet revised the hydraulics analysis for the two-pier structure. In lieu of this information, we will continue to use data from the three-pier hydraulics analysis. Because the three-pier scenario would result in a larger hydraulic shadow, it is assumed that this is an overestimate of the effect of hydraulic shadowing.

The in-water piers of the new structure would permanently increase hydraulic shadowing in North Portland Harbor and the Columbia River. Exhibit 4-14 shows the current hydraulic footprint of the existing structures in the Columbia River for the 100-year event, given preliminary construction design data. In the Columbia River, the hydraulic shadow extends 200 to 1,100 feet downstream of the piers, with velocities in the shadow ranging from 0 to 3 feet per second (fps). The hydraulic footprint was not modeled for the existing North Portland Harbor structures (DEA 2006).

Error! Reference source not found. Exhibit 4-15 and Exhibit 4-16 show the predicted post-project hydraulic footprint for a 100-year flow event in the Columbia River and North Portland Harbor. In the Columbia River, the hydraulic shadow of the completed structures is expected to increase significantly compared to that of the existing structures, extending up to 1,600 feet downstream of each pier, with velocities in the shadow ranging from 0 to 3 fps.

Although the hydraulic shadow was not modeled at the existing North Portland Harbor structures, it is expected to increase in length because of the increase in the number of shafts and the width of the structures. The hydraulic shadow of the completed North Portland Harbor structures is expected to extend up to approximately 400 feet downstream of each pier, with velocities in the shadow ranging from 0 to 2 fps.

4.1.3.1 Effects of Hydraulic Shadowing on Fish

Hydraulic shadowing may affect fish by creating low-velocity eddies that have the potential to increase predation, interfere with movement patterns, and alter sediment transport.

Predation

In general, hydraulic shadowing has the potential to harm prey fish by creating low-velocity areas or eddies that enhance the foraging success of predaceous fish and birds. Juvenile salmonids, all life stages of eulachon, juvenile lamprey, and all life stages of other native resident fish are vulnerable to predation. Subvearling salmonids are particularly vulnerable (Pribyl et al. 2004). Yearling salmon move quickly and migrate when they are of a size that reduces vulnerability to predators. In contrast, subyearling salmon are slower and are of a size that increases their vulnerability to predation (Gray and Rondorf 1986). Likewise, juvenile eulachon do not have volitional mobility and consequently cannot evade predation easily. Additionally, subyearling salmonids and other juvenile fish are highly dependent on low-velocity areas for rearing and resting. This overlaps with the preferred habitat type of northern pikeminnow, smallmouth bass, largemouth bass, and walleye (Pribyl et al. 2004), which are chief predators of juvenile salmon and other native fish in the lower Columbia River (Gray and Rondorf 1986). Predation on juvenile salmonids by fish generally occurs at velocities of 4 fps or less (NMFS 2008d). Predation rates on native resident fish (e.g., dace, threespine stickleback, redside shiners, suckers, and sculpin) are not available for the project area; however, these species are also taken as prey by larger fish.

Northern pikeminnow is the major predator of emigrating juvenile salmonids in the Lower Columbia (Poe et al. 1994; NMFS 2000). Northern pikeminnow are associated with pilings and other in-water structures during most of the year (Pribyl et al. 2004; Petersen and Poe 1993). Northern pikeminnow select slower-velocity areas, generally avoiding velocities greater than 2.3 fps (NMFS 2000). Petersen and Poe (1993) reported northern pikeminnow congregating at overwater structures, such as back eddies behind pilings. Consumption rates are especially high in areas where juvenile salmonids congregate.

The literature is not in complete agreement about northern pike minnow consumption rates of juvenile salmonids in the Lower Columbia basin. Buchanan et al. (1981, as cited in NMFS 2000) reported that only 2 percent of northern pikeminnow found in free-flowing sections of the Willamette River contained salmonids in their diets. In a free-flowing reach of the lower Columbia River, Thompson (1959, as cited in NMFS 2000) found that only 7.5 percent of northern pikeminnow contained salmonids in their diets. However, in a survey of the lower Columbia River from Bonneville Dam (RKm 235) to Jones Beach (RKm 71-77), Petersen and Poe (1993) found that catches of northern pikeminnow and the number of salmonid prey per pikeminnow were higher in free-flowing sections of the river than in impounded areas in John Day Reservoir. At a sampling site in Vancouver, the spring diet of northern pikeminnow was comprised of 70 percent fish, 92 percent of which were salmonid smolts. In summer, the diet was 25 percent fish, 84 percent of which were salmonid smolts (Petersen and Poe 1993). The study estimated that the average predation rate in spring at the Vancouver site was 1.3 smolts per pikeminnow. In summer, the predation rate in the same location was 1.7 smolts per pikeminnow. Of the non-salmonid fish prey, approximately half were sculpins (*Cottus* spp.); evidence of predation on lamprey was noted in less than 1 percent of sampled pikeminnow. Zimmerman (1999) found that daily consumption of juvenile salmonids in unimpounded portions of the Columbia River were about 0.8 prey per northern pikeminnow in the spring and 1.6 in the summer.

Mean maximum length of salmon consumed was 167 mm, although northern pikeminnow consumed both steelhead and Chinook measuring more than 200 millimeters in length. Of the salmonid smolts consumed, the large majority were juvenile Chinook (64 percent of all fish consumed), but they also ate steelhead (2 percent of fish consumed), and "unidentified salmonids" (26 percent of fish consumed). In another study, NMFS (2000b) estimates that the ratio of northern pikeminnow to the number of salmon smolts consumed between Bonneville Dam to the mouth to the Columbia River is 0.09 smolts per day. Northern pikeminnow are especially abundant in free-flowing reaches of the lower Columbia River. In a 2-year predator sampling study of the lower Columbia from Bonneville Dam to RKm 70, northern pikeminnow comprised over 90 percent of the predaceous fish species encountered (Poe et al. 1994). Other predators (smallmouth bass and largemouth bass) were few in the study area. Other important prey species identified in these studies included redside shiners (*Richardonius balteatus*) and threespine sticklebacks (*Gasterosteus aculeatus*) (Gray and Rondorf 1986).

Smallmouth bass are known to exhibit strong cover-seeking behavior and typically seek out pools or deep areas behind rocks where the current is slack (Edwards et al. 1983; Pflug and Pauley 1984; Probst et al. 1984, as cited in Pribyl et al. 2004). They also associate with in-water structures such as pilings and riprap (Pribyl et al. 2004). In the Columbia River basin, smallmouth bass prey heavily on juvenile salmonids (Gray and Rondorf 1986). While Zimmerman (1999) found that the mean maximum length of smolts consumed was 119 mm, they may also ingest very large prey (up to 240 mm) (NMFS 2000). Subyearling salmonids are at highest risk, not only because their shallow-water habitat overlaps with the preferred habitat of smallmouth bass in summer, but also because they are the ideal forage size for this species (Gray and Rondorf 1986). Rearing subyearling Chinook are particularly vulnerable (Poe at al. 1994; NMFS 2000). Zimmerman (1999) estimates that consumption rates exceeded 1.0 juvenile salmonids per smallmouth bass in both impounded and unimpounded reaches of the Columbia River. All of the prey items were either Chinook (12 percent of all fish consumed) or "unidentified salmonids" (3 percent of all fish consumed). No steelhead were detected. Other important prey species identified in these studies included sculpins and suckers (Catostomus spp.) (Gray and Rondorf 1986, Zimmerman 1999).

Largemouth bass prefer low-velocity areas, such as backwaters, when in riverine environments (Wheeler and Allen 2003; Wydoski and Whitney 2003). Additionally, when located in high-velocity river channels they are associated with in-water structures (Pribyl et al. 2004). Largemouth bass are present in the Columbia system, but because their numbers are relatively low, they do not have the potential to significantly affect the abundance of juvenile salmonids (Gray and Rondorf 1986).

Walleye are present in the lower Columbia River, but there is disagreement about the impact of this species on the abundance of juvenile salmonids in this area (Gray and Rondorf 1986). Walleye are frequently associated with pilings, as they avoid strong current. During their spring spawning period, walleye may prey preferentially on smaller juvenile salmonids (less than 100 mm) where both overlap in shallow-water habitat (Gray and Rondorf 1986). At other times of the year, walleye may be spatially segregated from juvenile salmonids, occurring more frequently offshore in deep water (Pribyl et al. 2004). In a sampling study, Poe et al. (1994) found that walleye abundance was low in the Columbia River from Bonneville Dam to RKm 70, comprising only 2 percent of all piscivorous fish captured. Zimmerman (1999) also detected very few walleye in the same area and found that 12.5 percent of the walleye diet was Chinook, with no other salmonids species detected. In the lower Columbia River, NMFS (2002) research underscores this point, noting that non-salmonid fish dominated the walleye diet.

It is not possible to quantify the number of individuals potentially exposed to increased predation. However, given that there is a net increase in the extent of suitable predator habitat, it is probable that the project would result in some level of increased predation on juvenile salmonids, eulachon, possibly juvenile lamprey, and other native resident fish in the Columbia River and North Portland Harbor.

There are no specific data regarding the impact of hydraulic shadowing on predation rates of eulachon (reports do not specify prey items at the species level); however, because both adult and larval eulachon are well within the size range (less than roughly 150 mm) consumed by common predators in the Columbia River, it cannot be discounted that hydraulic shadowing could also increase predation on adult and larval eulachon in the same manner as for juvenile salmonids.

The change in hydraulic footprint is not expected to increase predation on adult salmon and steelhead, adult and subadult bull trout, or adult and subadult green sturgeon, as predation on fish of these size classes is rare (Zimmerman 1999). Additionally, because of the extremely low numbers of bull trout and green sturgeon in this portion of the project area, risk of exposure to this effect is discountable.

Outmigration of Juvenile Salmonids

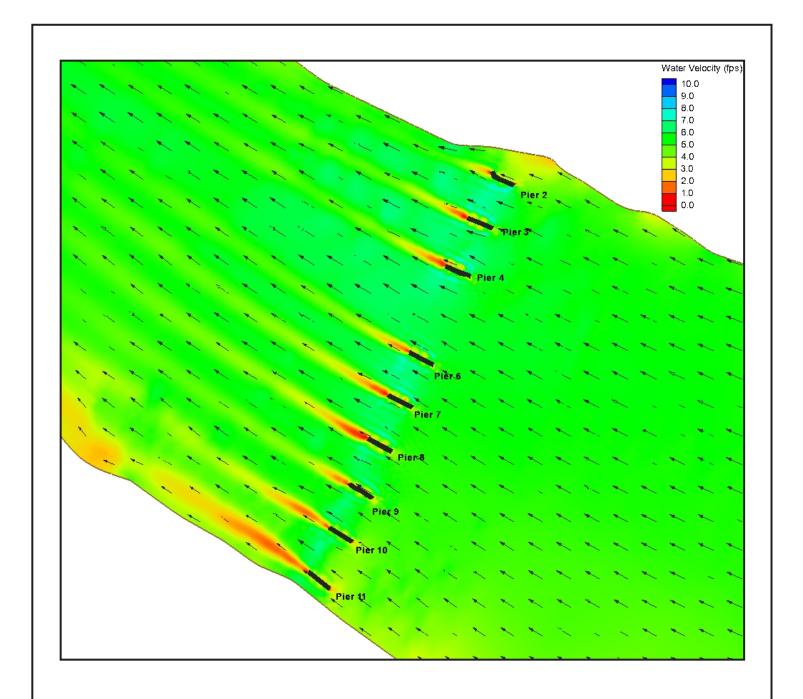
In general, hydraulic shadowing and resulting low-velocity areas have the potential to delay outmigration for smolts. Increased travel time exposes smolts to a variety of mortality vectors, including predation, disease, poor water quality, and thermal stress. Migration delays may also deplete energy reserves and disrupt arrival times in the lower estuary. The latter may cause salmonids to arrive in the estuary when predation levels are high and/or prey species are limited (NMFS 2008e). In the case of this project, effects to outmigration are expected to be slight. Although the size of the hydraulic shadow would increase, the range of velocities found in the hydraulic shadow is comparable to that which fish would encounter in the natural environment. Therefore, none of the juvenile fish present in the CRC project area are likely to become trapped or significantly delayed by the hydraulic shadow. Additionally, none are likely to be directed towards or away from shallow-water habitat because the structures neither pose a complete physical blockage to the shallow-water habitat, produce water velocities low enough to trap fish, nor produce velocities high enough to direct fish into deeper water. The effects of hydraulic shadowing on juvenile migration would be insignificant.

Velocity Refugia

Increased hydraulic shadowing may also benefit salmonids by creating larger velocity refugia for both adults and juveniles during periods or in reaches of high flow. A Bonneville Power Administration study showed that upstream passage through reaches with long, relatively uninterrupted stretches of high-velocity flow requires high levels of bio-energetic expenditure, similar to that of ascending a waterfall. Without resting areas, migrating adults use larger amounts of energy, posing risks for spawning success (Brown and Geist 2002). Velocity refugia allow fish to rest and replenish energy reserves. The CRC project area and vicinity consist of long relatively uninterrupted stretches of high-velocity flow. Presumably, the increased size of the hydraulic shadows would increase the area of flow refugia over the preproject condition. The extent to which this increase may benefit listed fish is impossible to quantify, but given that the increase in flow refugia is small relative to the large size of the Columbia River and North Portland Harbor, the effect is probably slight and therefore insignificant.

Sediment Transport

The hydraulic effect of the new bridges may alter sediment transport in the Columbia River and North Portland Harbor. Between bridge piers, water velocities are likely to increase, resulting in increased sediment transport. In lower-velocity areas behind the piers, sediment is likely to accumulate. Several new piers are located immediately adjacent to the shoreline (in the Columbia River: pier complexes 2 and 7; in North Portland Harbor, the six new nearshore bridge bents). Low-velocity areas behind these piers would likely accumulate sediment; therefore, the new bridge piers are not anticipated to cause shoreline erosion.



Source: CRC Hydraulic and Scour Parameters Report 2008

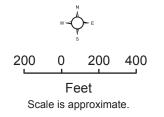
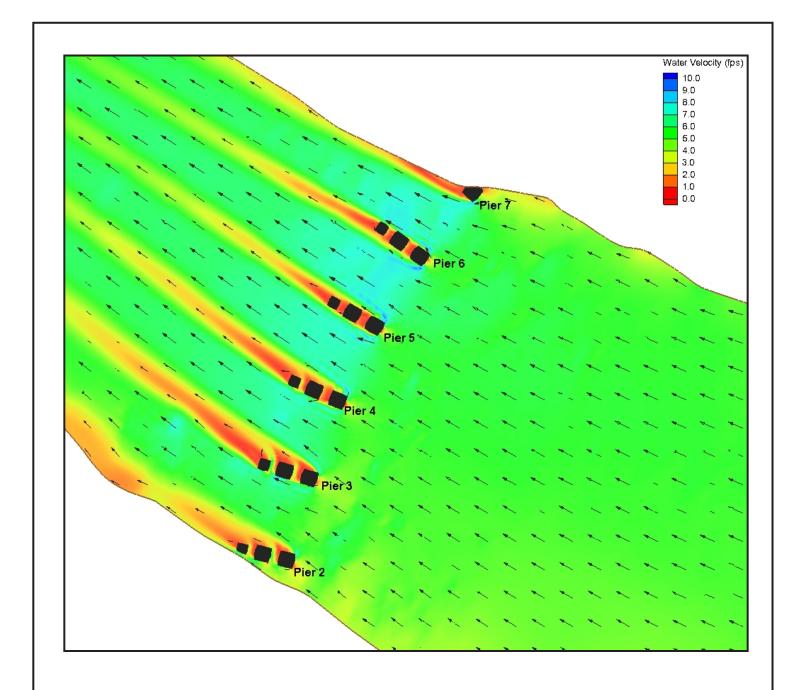


Exhibit 4-14. Velocity Vector Plot of Existing Structures in the Columbia River for 100-Year Flood





Source: CRC Hydraulic and Scour Parameters Report 2008

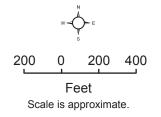
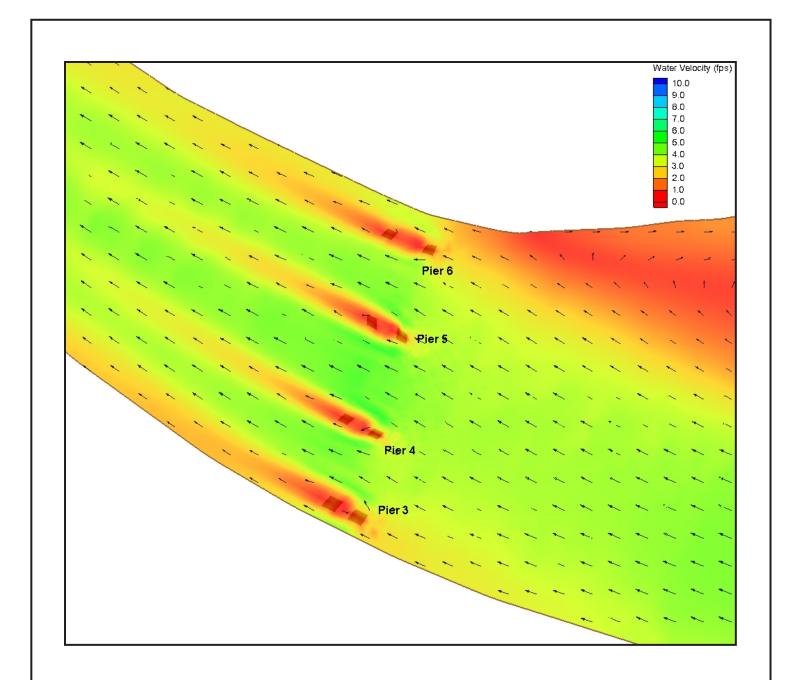


Figure 4-15. Velocity Vector Plot of Proposed Structures in the Columbia River for 100-Year Flood





Source: CRC Hydraulic and Scour Parameters Report 2008

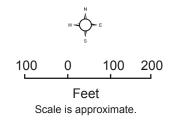


Exhibit 4-16. Velocity Vector Plot of Proposed Structures in the North Portland Harbor for 100-Year Flood



4.2 Effects to Terrestrial Resources

4.2.1 Terrestrial Habitat

Construction activities associated with the LPA would impact terrestrial habitats. Exhibit 4-17 compares project impacts to the various terrestrial habitat types between the No Build alternative and the LPA. Comparison is made of acres of habitat within right-of-way to represent the most consistent project footprint possible between these alternatives. For purposes of this analysis, impacts include construction cut/fill activities, paved surface, area that may be accessed in the right-of-way for maintenance, and other ground-disturbing and potentially habitat-disturbing activities.

Exhibit 4-17. Terrestrial Habitat Impacts

_	LI	No-Build	
	LPA Option A	LPA Option B	
Washington Priority Habitats	37.0	36.9	29.5
Vancouver Critical Areas	121.0 (119.4)	120.7 (119.2)	108.8
Metro Goal 5	48.5 (48.6)	47.0 (46.9)	25.8
City of Portland E-Zones	42.6 (42.4)	41.3 (40.9)	27.9
Totals	249.1 (247.4)	245.9 (243.9)	192.0

a Text in parentheses indicates impacts if the LPA Option A or B is constructed with highway phasing.

4.2.2 Riparian Habitat

In North Portland Harbor and the Columbia River, effects to riparian habitat would be negligible, as there is very little functioning riparian vegetation in the project area. The project would revegetate disturbed shoreline areas, resulting in a net benefit to riparian habitat in the long term. It has not yet been determined exactly where replanting would take place. However, it is anticipated that replanting would occur on or adjacent to the current sites of the trees where practicable. In any case, the number, type, and size of the replanted trees would be selected to comply with standards outlined in the City of Portland and City of Vancouver tree ordinances.

In Oregon, the project would remove three deciduous trees, all with trunks less than 1 foot in diameter, from the riparian zone on the south bank of the Columbia River. The project would also remove two deciduous ornamental trees from the riparian zone adjacent to North Portland Harbor. These trees are located in a landscaped setting and have trunks of approximately 1 foot in diameter. In Washington, 10 trees with trunks less than 1 foot in diameter would be removed from the riparian zone on the north shore of the Columbia River.

In general, removal of trees from riparian areas results in a reduction of shade in the water column and a concurrent increase in water temperature. However, in the case of the CRC project, only approximately 15 trees would be removed from the Columbia River/North Portland Harbor riparian area. This represents an extremely small amount of shaded water (less than 10,000 square feet, patchily distributed among at least three locations) relative to the thousands of acres of unshaded water located immediately adjacent to the area from which trees would be removed. Because of the small size of the shaded area relative to the large volume of water and because of the high current velocity in these water bodies, it is unlikely that these 15 riparian trees create enough shade to measurably decrease water temperatures in the water column. Thus, the loss of these trees is expected to cause only negligible effects to water temperature, if any.

Additionally, removal of trees from riparian areas may reduce the potential for large woody debris recruitment in a watershed over the long term. However, given the large size of the lower Columbia system and the thousands of remaining riparian trees in this area, removal of 15 trees would not measurably decrease the potential for long-term large woody debris recruitment in the project area or in the lower Columbia system overall.

There would be no excavation, vegetation clearing, or removal of trees from the Columbia Slough riparian area. Therefore, the project would have no effect on Columbia Slough riparian habitat.

Exhibit 4-18 illustrates the acreage and locations of PHS riparian buffer at Burnt Bridge Creek that are likely to be impacted under the LPA.

4.2.2.1 LPA with Highway Phasing

If the project improvements at SR 500 are deferred under the LPA with highway phasing option, riparian impacts near Burnt Bridge Creek would be also be deferred.

4.2.3 Threatened, Endangered, and Proposed Species

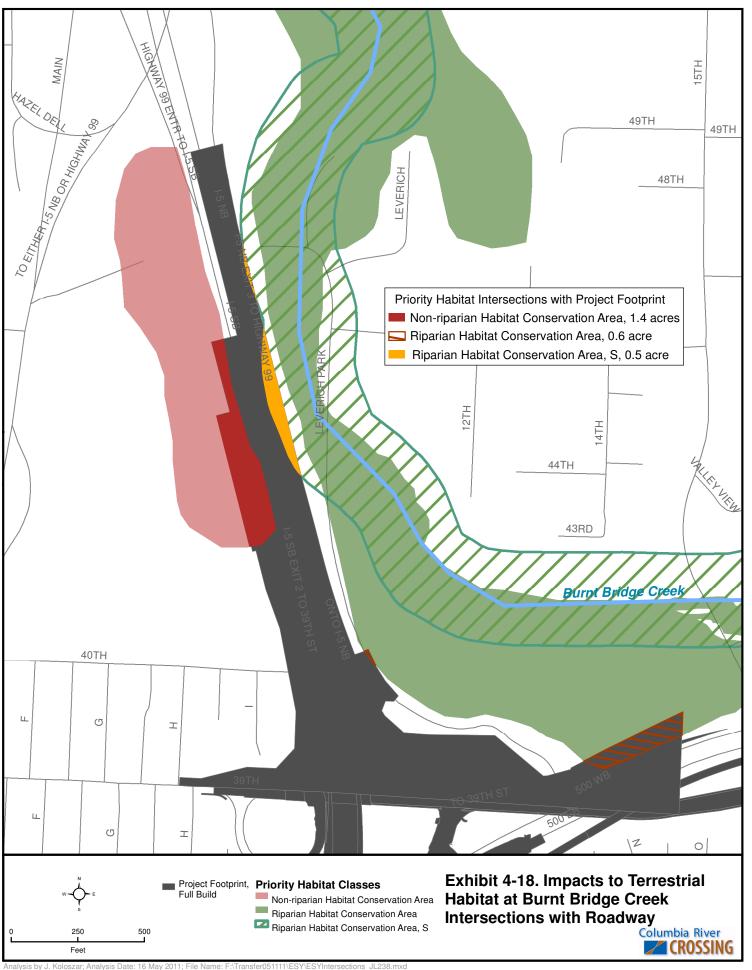
No long-term effects to terrestrial threatened, endangered, proposed, or candidate species would be expected under the LPA.

4.2.4 Species of Interest

The LPA would affect terrestrial resources by removing structures used by migratory birds and potentially by bats. Removal of these structures is a concern because life stages such as feeding and breeding may be affected. New habitat elements such as nest boxes could be included in the new structure to offset removal of habitat elements associated with the existing bridge. However, the LPA is not anticipated to have adverse long-term impacts to most terrestrial resources.

Peregrine falcons would be affected because the existing bridge, which the falcons have been documented using since 2001, would be removed. Removal of the habitat structure on the existing bridge would appreciably disrupt peregrine breeding, foraging, and roosting activity. Peregrines using the existing bridge would be forced to find alternative structures in the area, or would vacate the area in the long term.

Long-term effects to migratory birds could include altered habitat for nesting and roosting if the new bridge design provided less structure suitable for these species (e.g., the new structure would not include steel girders that birds currently use).



4.2.5 Wildlife Passage

Wildlife passage may be hindered compared to existing conditions. The existing shoreline provides minimal passage habitat in the form of open riprap and concrete. Piers for the new bridges would likely impact one or both shores of the Columbia River, creating an obstruction to movement along the shoreline. Options for improving wildlife passage are limited; however, habitat connections would be upgraded where feasible, particularly in riparian areas.

4.3 Effects to Botanical Resources

The LPA is not anticipated to have long-term impacts to botanical resources. Effects to vegetation will be addressed through mitigation measures discussed in Section 6.

4.4 Indirect Effects

Changes in auto and transit use, biking and walking, as well as changes in land use, are anticipated to occur under the LPA and No-Build alternatives. Potential positive and negative impacts to species and habitats could occur from land use development and resulting changes to trip patterns, including impacts to water quality and water quantity. In addition, development may result in changes to riparian and nearshore areas, including changes in vegetation and overwater structures. Species may be affected through the addition of impervious surface (particularly PGIS), and a decrease in riparian and aquatic habitat.

The LPA and local plans are expected to promote redevelopment adjacent to or near proposed light rail stations in downtown Vancouver and on Hayden Island. With redevelopment of existing infrastructure, applicable land use codes would be implemented, in particular the need to upgrade to existing stormwater treatment regulations. Because these sites are located in already highly developed areas, habitat for terrestrial species is extremely limited to non-existent at these sites; however, stormwater runoff could indirectly positively or negatively impact habitat associated with fish species. Development and redevelopment, including removal or renovation of existing in-water structures and near-shore development, would comply with the relevant laws, regulations, policies, and codes in force at the time of the action. These regulatory approvals range from tree and street tree removal, to stormwater treatment, to environmental zone and critical areas protections, to more complicated processes for larger developments.

With the integration of local and state land use requirements, negative impacts from development and redevelopment would be limited. Local regulations require the avoidance or minimization of impacts to protected resources. These resources include shorelines, wetlands, stream banks, and their buffers, resources that are often most important to juvenile salmonids and their habitat. With implementation of regulations such as environmental zones, the Shoreline Management Act (SMA), and CAO, impacts to existing resources would be negligible.

Interstate 5 Columbia River Crossing Ecosystems Technical Report for the Final Environmental Impact Statement

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5. Short-term Effects

This section describes the short-term effects the LPA Option A ("LPA") full build would have to aquatic, terrestrial, and botanical resources. Unless stated otherwise, the LPA with highway phasing options would have the same impacts to aquatic, terrestrial, and botanical resources as the corresponding LPA full build options. Similarly, the short-term impacts to aquatic, terrestrial, and botanical resources are expected to be the same, whether Option A or Option B is built, except where noted (see section 5.3.2.1).

5.1 Introduction

Unavoidable impacts to ecosystem resources, particularly the Columbia River and North Portland Harbor, are likely to occur. Modifications to migratory bird habitats are likely to occur as existing vegetation, as well as the bridge structures themselves, are expected to be removed. This would be a short-term effect to migratory bird habitat because vegetation replanting, and new bridge construction, would occur.

Temporary effects to aquatic habitat and aquatic species are anticipated from in-water work. In-water work may include removing existing bridge piers, constructing new piers, and installing and removing temporary in-water work structures. In-water work is likely to include coffer dams, barges, drilling equipment, impact pile drivers, vibratory pile drivers, and other construction vehicles in and near the water. In-water work would likely cause localized increases in underwater noise, turbidity, artificial lighting, avian predation, hydraulic shadowing, and shading. Construction activities may cause injury or death to aquatic species. Exhibit 5-1 shows the anticipated sequencing of in-water structures for the construction in the Columbia River.

Temporary effects to terrestrial species are anticipated from construction noise and impacts to vegetation. Construction activity causing noise disturbance could result in reduced nesting success for migratory birds. Trees, shrubs, and other vegetation serving as cover, nesting, roosting, and perching habitat may be removed during construction. Such vegetation removal could also impact terrestrial wildlife using such habitat structure for cover, feeding, breeding, and dispersal. Appendix B provides site-specific hydroacoustic data from a test pile installation project completed in February 2011.

5.2 Effects to Aquatic Resources

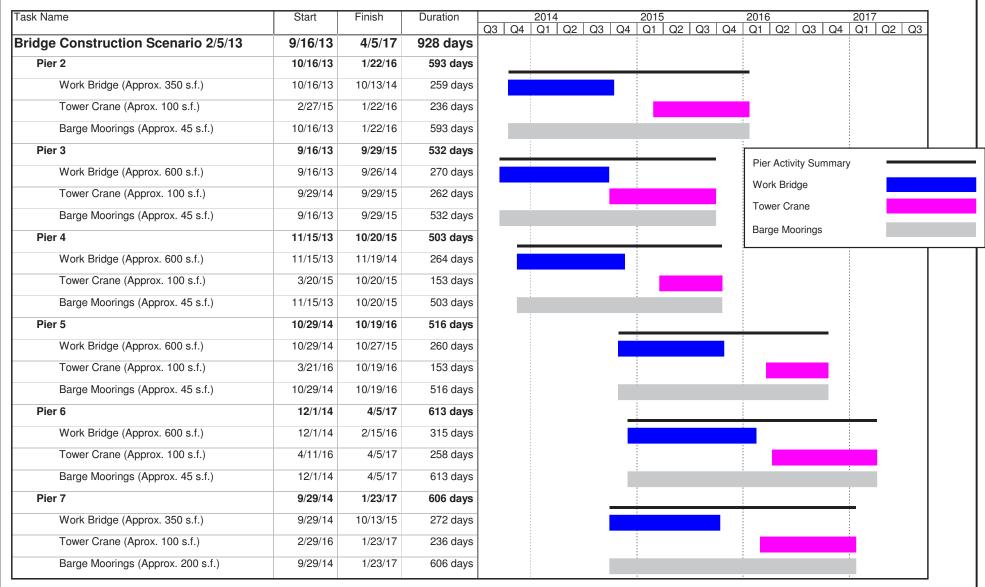
5.2.1 Acoustic Impacts

Direct injury, mortality, or behavioral disturbance to fish species may result from sound levels produced by impact pile driving, vibratory pile driving, and other in-water construction techniques used for the installation of temporary and permanent in-water structures in the Columbia River and North Portland Harbor. Effects associated with pile driving may include physical injury (particularly to air-filled spaces such as swim bladders), auditory tissue damage, temporary or permanent hearing loss, behavioral effects, and immediate and delayed mortality. The amount of energy and the resulting sound pressure from impact pile driving depend on the size and type of pile, type of hammer, energy of the hammer, depth of the water column, and substrate. Impacts to individual fish depend on sound pressure levels, fish species, fish size, fish condition, fish movement, and depth of the water column (Popper et al. 2006). Use of bubble curtains or other noise attenuation devices during impact pile driving may reduce the level of noise impacts to fish (Caltrans 2009).

Sound, measured in dB, is a relative measure and is referenced in the context of underwater sound pressure to 1 micropascal (μ Pa) ("dB re: 1 μ Pa"). One pascal is the pressure resulting from a force of 1 newton exerted over an area of 1 square meter. For purposes of this analysis, underwater sound is referenced in units of decibels re: 1 μ Pa when referring to sound pressure levels (SPLs) or 1 μ Pa²-second when referring to sound exposure levels (SELs), and will be denoted as dB.

Root-mean-square (RMS) is the quadratic mean sound pressure over the duration of an impulse. This measurement is often used in the context of discussing behavioral effects to fish, in part because behavioral effects, which often result from auditory cues, and effects on hearing may be better expressed through averaged units rather than by peak pressures. When discussing the effects of explosions on animals, authors often use impulse as the acoustic parameter, as in Yelverton et al. (1973) discussed below. Positive impulse is the integral of pressure over time, from arrival of the leading edge of the pulse until the pressure becomes negative. Impulse is measured in pascal-seconds (Pa-s). As sound propagates away from a source, several factors change its amplitude. These factors include the spreading of the sound over a wider area (spreading loss), losses to friction (absorption), scattering and reflections from objects in the sound's path, and interference with one or more reflections of the sound off the surface of the streambed (in the case of underwater sound).

The sum of all propagation and loss effects on a signal is referred to as the transmission loss. A major component of transmission loss is spreading loss. From a point source in a uniform medium (water or air), sound spreads outward in spherical waves. Sound transmission in shallow water is highly variable and site specific. Refraction can result in either reduced or enhanced sound transmission in shallow water (Richardson et al. 1995). Ambient noise is the background noise. In water, sources of ambient noise include wind, waves, organisms, shipping traffic, and rain.



Conceptual Schedule Only, March 2010

Note: This is a proposed schedule, so activity start and finish dates are likely to change.

Exhibit 5-1. Sequencing of In-Water Structures for Construction in the Columbia River



5.2.1.1 Hydroacoustic Effects to Fish from Impact Pile Driving

Hydroacoustic injury and disturbance thresholds and guidance for fish species have been identified by NMFS and USFWS for impulse noises, such as impact pile driving (Exhibit 5-2) (Popper et al. 2006; Southall et al. 2007; NMFS 2009). Some of the thresholds are dependent on whether the fish are greater than or equal to 2 grams (g) in size. Fish potentially occurring in the project area include adult salmonids, adult and subadult green sturgeon, adult eulachon migrating upriver, larval eulachon, steelhead kelts migrating downriver, outmigrating juvenile salmonids, and other native resident fish (e.g., sculpins, threespine sticklebacks, suckers, dace, and shiners). All of these species fall into the greater-than-2-g size class, except for juvenile chum, larval eulachon, and some larval and juvenile resident fish.

Exhibit 5-2 lists the injury thresholds and disturbance guidance for noise impacts to fish.

Other native and non-native fishes occur in the project area. These species include lamprey species, white sturgeon, cutthroat trout, and others. Most of the fish in the project area have swim bladders or other air-filled cavities in their bodies. Lampreys do not have swimbladders and it is therefore difficult to determine the extent of this impact. Fish species without swimbladders are thought to be at lower risk from underwater sound than fishes with swimbladders (Stadler, pers. comm. 2010, Hastings and Popper 2005, Coker and Hollis 1952, Gaspin 1975, Baxter et al. 1982, Goertner 1994). No thresholds for disturbance or injury have been established for such fish (Stadler pers. comm. 2010). Therefore, hydroacoustic impacts to lamprey should not be discounted, but they cannot be quantified or analyzed with any level of certainty.

Exhibit 5-2. Hydroacoustic Injury Thresholds and Disturbance Guidance for Fish^a

Underwater Sound Criteria (dB measured at 10 meters from source)					
Size Class	Injury Threshold	Disturbance Guidance			
Fish ≥ 2 grams	206 dB _{peak} ; 187 SEL _{cum}	150 dB _{RMS}			
Fish < 2 grams	206 dB _{peak} ; 183 SEL _{cum}	150 dB _{RMS}			

Where cumulative SEL (SEL_{cum}) is calculated as: SEL(_{cum}) = SEL(single strike at ~10 meters from the pile) + 10 log * (# strikes).

Impact pile driving would occur during installation of temporary in water work structures in the Columbia River and North Portland Harbor as described in Section 1 (Description of the LPA). Temporary piles used in these structures are expected to fall into two size classes: 18 to 24 inches and 36 to 48 inches in diameter.

Approximately 1,500 temporary steel piles would be installed and removed during the multi-year construction of the Columbia River and North Portland Harbor bridges. The need for piles would be staged over the in-water construction and demolition periods so that between 100 and 400 piles may be in the water at any given time.

Temporary structures that are not load-bearing, such as mooring piles and cofferdams, would be installed with a vibratory driver only. Drilled shaft casings may also be vibrated into position. These vibratory driving activities are proposed to occur year-round and without the use of an attenuation device.

Structures requiring load bearing piles include temporary work bridges, work platforms, tower cranes, and oscillator support platforms. These piles would be installed first with a vibratory driver to refusal and then proofed with an impact hammer.

Each pier complex of the Columbia River bridges would require approximately 132 load-bearing piles for support of work platforms/bridges and an additional eight load-bearing piles for a tower crane several months later, for a total of approximately 840 impact driven piles. An average of six temporary, load-bearing piles could be installed per day using one or two impact drivers. The project is anticipating that temporary piles for each of the six work bridges/work platforms would be installed in one 22-day period. Temporary piles for each of the six tower cranes would be installed in one day. Impact pile driving in the Columbia River would occur on approximately 138 days over the approximately 4-year construction period.

Each of the 31 oscillator support platforms in North Portland Harbor would require four load-bearing piles (124 total piles). In addition, the nine temporary work bridges would each require approximately 25 load-bearing piles (225 total piles). There would be a total of approximately 349 impact-driven piles in North Portland Harbor. Only one impact driver would operate at a given time in North Portland Harbor. Impact pile driving in North Portland Harbor would occur on approximately 134 days over the approximately 4-year construction period.

In-water noise attenuation measures would be employed during impact driving activities for the majority of pile strikes. The CRC project assumes that an at-source noise reduction of approximately 10 dB is achievable through use of a noise attenuation device. Unattenuated pile driving may occur as part of the hydroacoustic monitoring program for this project or incidentally during attenuation equipment failures. In the Columbia River, unattenuated pile driving may occur for up to 7.5 minutes per week. In North Portland Harbor, unattenuated pile driving may occur on average for up to 5 minutes per week.

Based on NMFS models, calculation of distances to injury thresholds and disturbance guidance is related to noise from a single pile strike. For accumulated SEL, the variables include: single-strike dB SEL, the number of pile strikes over a time period, the time period, the distance from pile, and fish movement.

During construction of the Columbia River bridges, up to two impact pile drivers may operate simultaneously in close proximity to one another. The operation of two pile drivers is not anticipated to produce noise levels greater than that of a single pile driver. Pile strikes from both drivers would need to be synchronous (within 0.0 and approximately 0.1 seconds apart) in order to produce higher noise levels than a single pile driver operating alone. Because it is highly unlikely that two pile drivers would operate with exactly synchronous pile strikes, the CRC team assumes that two pile drivers would not generate noise at levels greater than that of a single pile driver.

For construction of the mainstem Columbia River bridges, an average of 300 impact blows per pile are estimated to be needed. Project designers estimate that up to 1,800 attenuated pile strikes would occur per day of pile driving. For construction of the North Portland Harbor bridges, a total of 1,800 attenuated pile strikes per day of driving were also assumed. The actual number of pile strikes would vary depending on the type of hammer, the hammer energy and substrate composition. However, these pile strikes would not be spread evenly throughout the work day. It is likely that day-to-day pile driving activities would vary. This hour-to-hour and day-to-day variation, coupled with timing of fish runs and fish speed through the area, creates a complex scenario for analyzing effects.

To accommodate this complex scenario of pile sizes, initial sound levels, pile strike numbers, timing and duration of pile driving, etc., the CRC team developed an analytical tool to determine the extent to which fish are exposed to potentially injurious accumulated sound levels within the project area. The CRC project has called this extent of exposure the "exposure factor." The exposure factor uses the variables for calculating the accumulated SEL through the moving fish

model (size of pile [initial sound levels], daily pile strikes, timing and duration of pile strikes, fish speed, and fish mass) and combines that with variables, such as days of pile driving within a week, to estimate the potential exposure to fish that are within or pass through the project area. Different combinations of any of these elements (such as pile strikes, duration or timing of pile strikes, and initial sound levels) would yield different exposure factors. During construction, the contractor would calculate the weekly, maximum yearly, average yearly, and total project exposure factor to ensure that they do not exceed levels specified in Section 6 of this document.

Exposure factors were calculated for impact pile driving activities in both the Columbia River and North Portland Harbor.

The Services have accepted the use of a revised moving fish model based on this project's specific conditions to determine exposure factors and to quantify effects to listed fish. This model uses the mass and the measured or assumed rate of travel for juvenile and adult fish through the project area. Juvenile chum and larval eulachon were assumed to be under 2 g in mass and travel with the current at 0.6 m/s. Other juvenile fish were assumed to be over 2 g in mass and travel a little faster than the current of 0.8 m/s. All adult fish were assumed to be over 2 g in mass and travel at 0.1 m/s through the project area.

It is important to correctly assume the rate of travel and mass for the moving fish model. The faster a fish moves through an area, the less time it has to become exposed to accumulated levels of potentially injurious sound energy. The effect of speed on the area of effect is more noticeable at higher fish movement speeds (nearing 1.0 m/s), whereas the area of effect for fish moving 0.1 m/s are substantially the same as the area of effect calculated using the stationary fish model. For example, an attenuated 36- to 48-inch-diameter pile struck 300 times would result in a pile-driving time of approximately 7.5 minutes. A fish (over 2 g) moving at a speed of 0.8 m/s would travel approximately 360 meters in a 7.5-minute period. If that fish passed within approximately 47 meters of the driven pile, it could receive enough sound energy for injury to occur. If the fish were traveling at only 0.6 m/s, then it could experience enough sound energy for injury to occur within approximately 58 meters from the pile. If the fish were traveling at 0.1 m/s or was stationary, then it could experience enough sound energy for injury to occur within approximately 83 meters from the pile. If the fish passed inside of the threshold distance for its given speed, injury would be more likely.

In order to analyze potential impacts to listed fish, the CRC project team calculated the proportion of a listed fish run that may be impacted within the Columbia River and North Portland Harbor through potential injury due to increased sound pressure levels from the impact driving of temporary piles. Calculating exposures to fish requires multiplying the proportion of a fish run likely present in the project area in a given week by the weekly exposure factor for that week. The CRC project used 13 full Columbia River Bridge construction scenarios to estimate potential and maximum exposure factors.

Due to the numerous variables in determining exposure factors, the CRC team used representative numbers of pile strikes, such as those in Exhibit 5-3 and Exhibit 5-4, to estimate exposure factors for the project. The numbers in Exhibit 5-3 and Exhibit 5-4 are also used to illustrate the extent of underwater noise exceeding the injury thresholds and disturbance guidance.

Exhibit 5-3. Pile-Strike Summary for Columbia River Bridge Construction

Pile Size	Strikes per Day	Days per Week ^a	Strike Interval ^b
Without Attenuation Device			
Single pile driver: 18- to 24-inch pile	150	1	1.5

Pile Size	Strikes per Day	Days per Week ^a	Strike Interval ^b
Single pile driver: 36- to 48-inch pile	150	1	1.5
With Attenuation Device			
Single pile driver: 18- to 24-inch pile	400	5	1.5
Single pile driver: 36- to 48-inch pile	800	5	1.5
Two pile drivers: each with 18- to 24-inch pile	200	5	0.75
Two pile drivers: one 18- to 24-inch pile and one 36- to 48-inch pile, or two 36- to 48-inch piles	400	5	0.75

a Days per week during active driving only.

Exhibit 5-4. Pile-Strike Summary for North Portland Harbor Bridge Construction

Pile Size	Strikes per Day	Days per Week ^a	Strike Interval ^b
Without Attenuation Device			
Single pile driver: 18- to 24-inch pile	75	1	1.5
Single pile driver: 36- to 48-inch pile	75	1	1.5
With Attenuation Device			
Single pile driver: 18- to 24-inch pile	900	3 to 5	1.5
Single pile driver: 36- to 48-inch pile	900	2	1.5

a Days per week during active driving only.

Estimated Extent, Timing, and Duration of Effect

Exhibit 5-5, Exhibit 5-6, Exhibit 5-12, and Exhibit 5-13 summarize the distances within which noise exceeds the injury thresholds and disturbance guidance in the Columbia River and North Portland Harbor during impact pile driving. These distances are presented for impact pile driving occurring both with and without the use of an attenuation device for comparison. Note that the upstream extent of pile-driving noise may differ from the downstream extent. These values indicate the distance at which noise encounters a landform (such as an island or streambank) that completely blocks the spread of in-5-7water noise. The calculations assume that the noise attenuation device would achieve 10 dB of noise reduction at the source.

Exhibit 5-5, Exhibit 5-7, and Exhibit 5-8 show the distances within which noise exceeds peak injury thresholds.

Exhibit 5-5. Distances at Which Underwater Noise Exceeds 206 dB Peak Injury Threshold Levels for Peak Noise in the Columbia River and North Portland Harbor

	Distance	ce (m)
Pile Size	Without Attenuation Device	With Attenuation Device
18- to 24-inch pile	25	5
36- to 48-inch pile	34	7

Exhibit 5-9, Exhibit 5-10, and Exhibit 5-11 show the distances within which noise is estimated to exceed the 187 dB SEL injury thresholds for fish over 2 g and moving at 0.1 m/s for a single pile

b Measured in seconds between strikes.

b Measured in seconds between strikes.

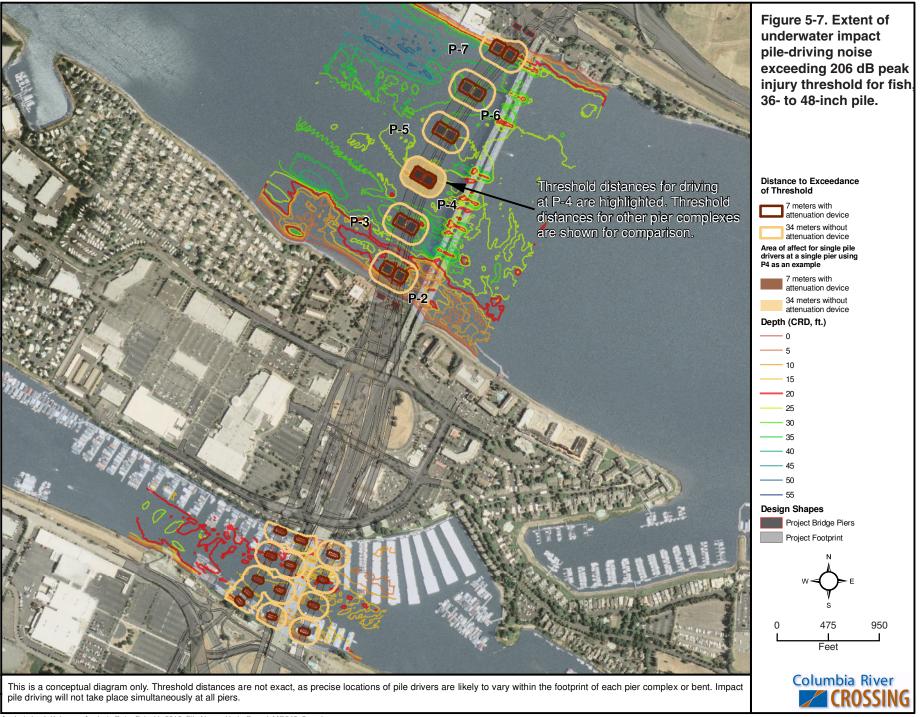
driver and for two pile drivers operating simultaneously, as calculated using the moving fish model.

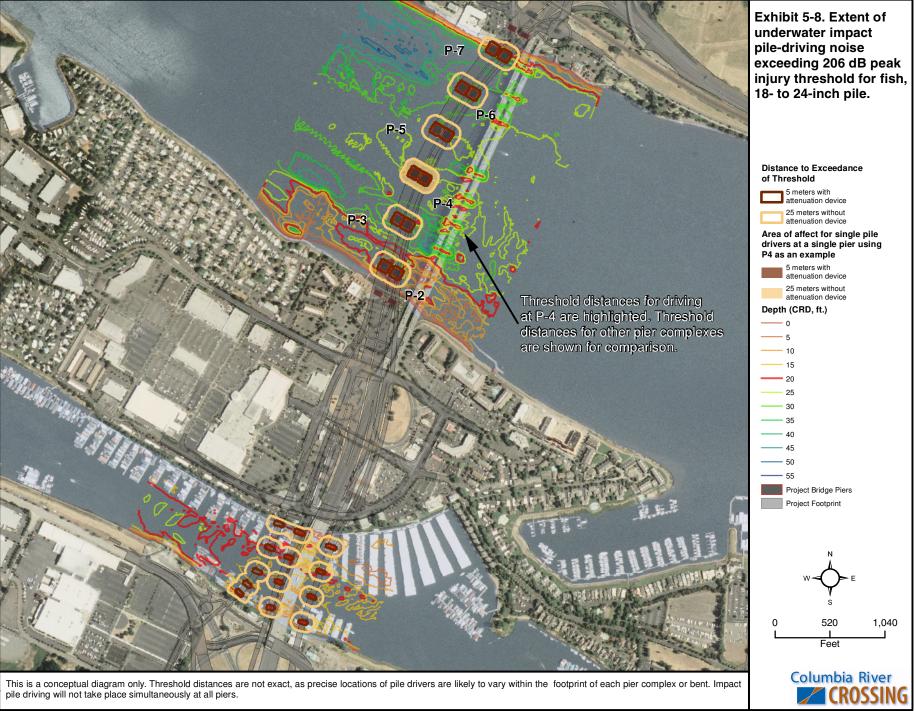
Exhibit 5-6. Distances at Which Underwater Noise Exceeds 187 dB SEL Injury Threshold for Adult Fish Under 2 g at 0.1 m/s in the Columbia River and North Portland Harbor

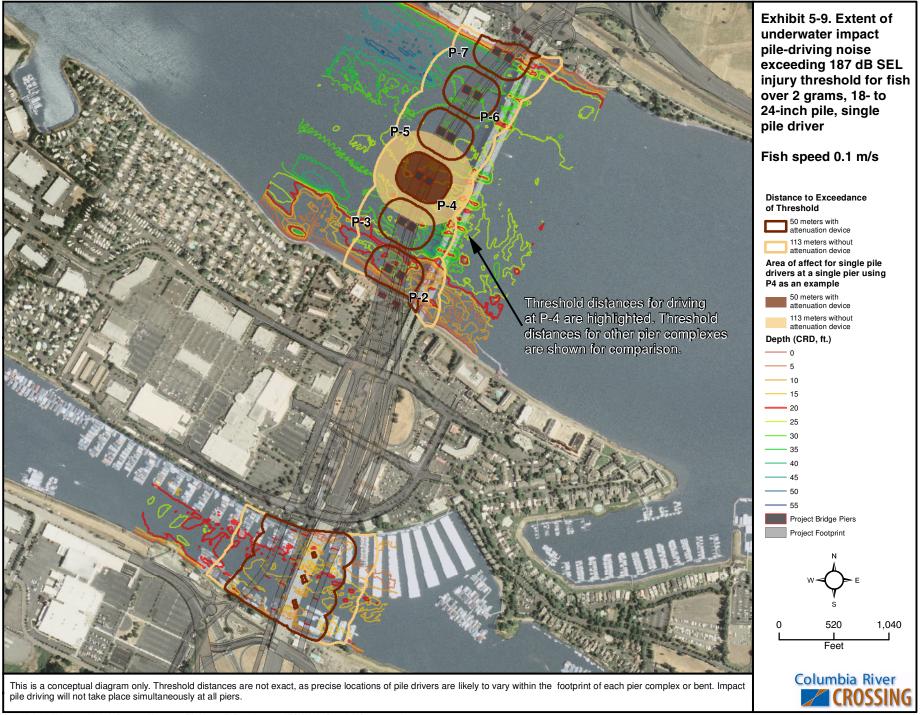
	Distance (m)				
Pile Size	Without Attenuation Device	With Attenuation Device			
18- to 24-inch pile	113	50			
36- to 48-inch pile	243	156			
Two 18- to 24-inch piles	N/A	59 ^a			
Two 36- to 48-inch piles OR One 18- to 24-inch and one 36- to 48-inch pile	N/A	130 ^a			

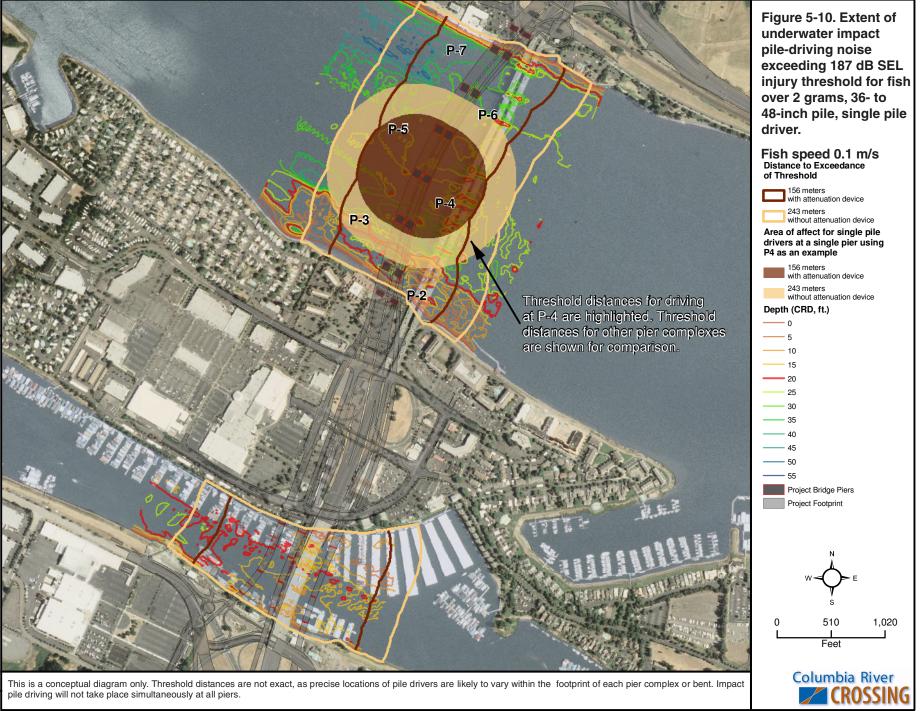
Note: Includes adult salmon, steelhead, and eulachon.

a Applies to Columbia River only.5-8









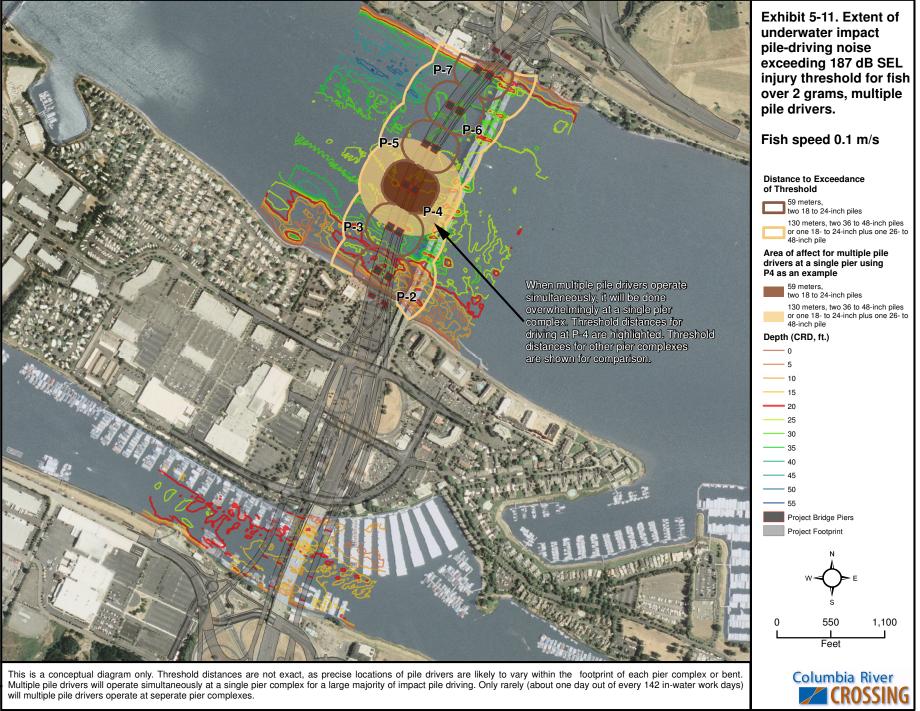


Exhibit 5-12, Exhibit 5-17, Exhibit 5-18, and Exhibit 5-19 show the distances within which noise is estimated to exceed the 187 dB SEL injury thresholds for fish over 2 g and moving 0.8 m/s for a single pile driver and for two pile drivers operating simultaneously.

Exhibit 5-12. Distances at which Underwater Noise Exceeds 187 dB SEL Injury Threshold for Moving Fish Over 2 g at 0.8 m/s in the Columbia River and North Portland Harbor

	Distance (m)			
Pile Size	Without Attenuation Device	With Attenuation Device		
18- to 24-inch pile	102	9		
36- to 48-inch pile	237	67		
Two 18- to 24-inch piles	N/A	48 ^a		
Two 36- to 48-inch piles OR One 18- to 24-inch and one 36- to 48-inch pile	N/A	111 ^a		

Note: Includes juvenile salmonids except for chum.

Exhibit 5-13, Exhibit 5-14, Exhibit 5-15, and Exhibit 5-16 present the results of calculations showing distances within which noise is estimated to exceed the 183 dB SEL injury thresholds for fish under 2 g and moving at 0.6 m/s for a single pile driver and for two pile drivers operating simultaneously.

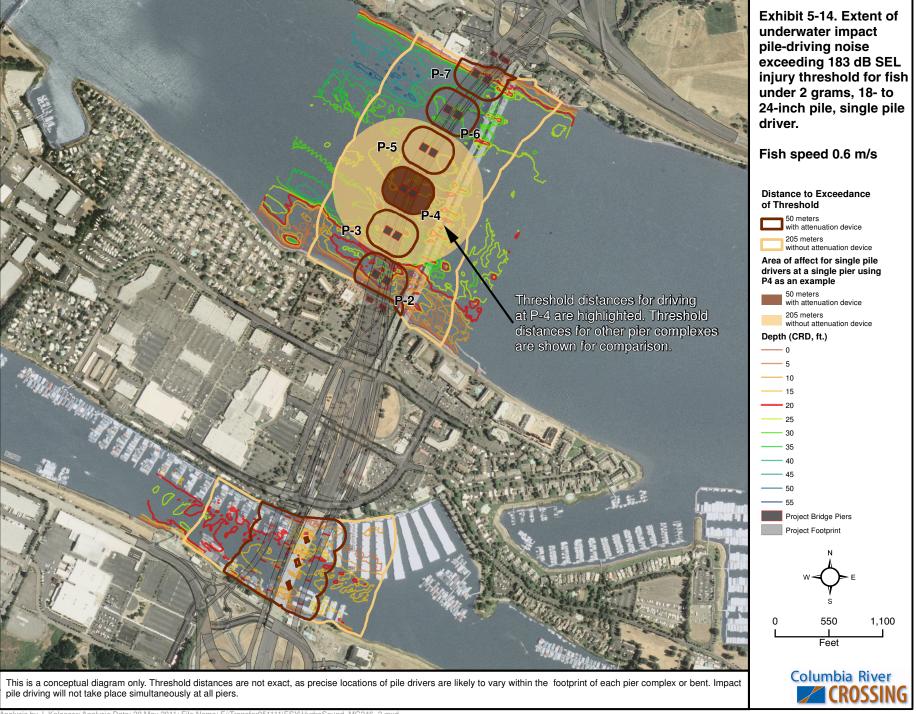
Exhibit 5-13. Distances within which Underwater Noise Exceeds 183 dB SEL Injury Threshold for Moving Fish Under 2 g at 0.6 m/s in the Columbia River and North Portland Harbor

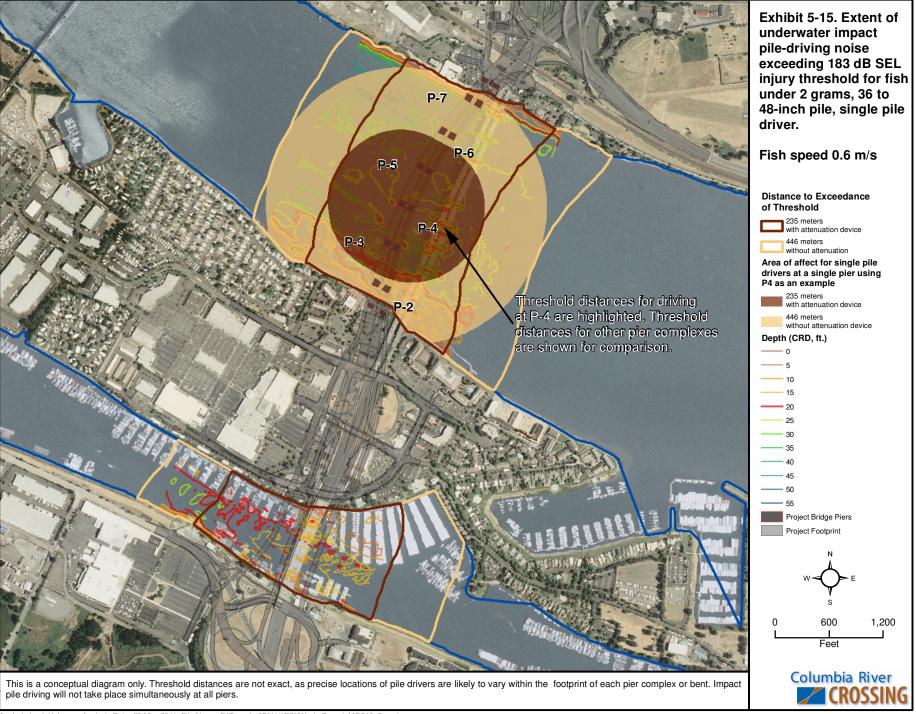
	Distance (m)				
Pile Size	Without Attenuation Device	With Attenuation Device			
18- to 24-inch pile	200	50			
36- to 48-inch pile	446	235			
Two 18- to 24-inch piles	N/A	79 ^a			
Two 36- to 48-inch piles OR One 18- to 24-inch and one 36- to 48-inch pile	N/A	209ª			

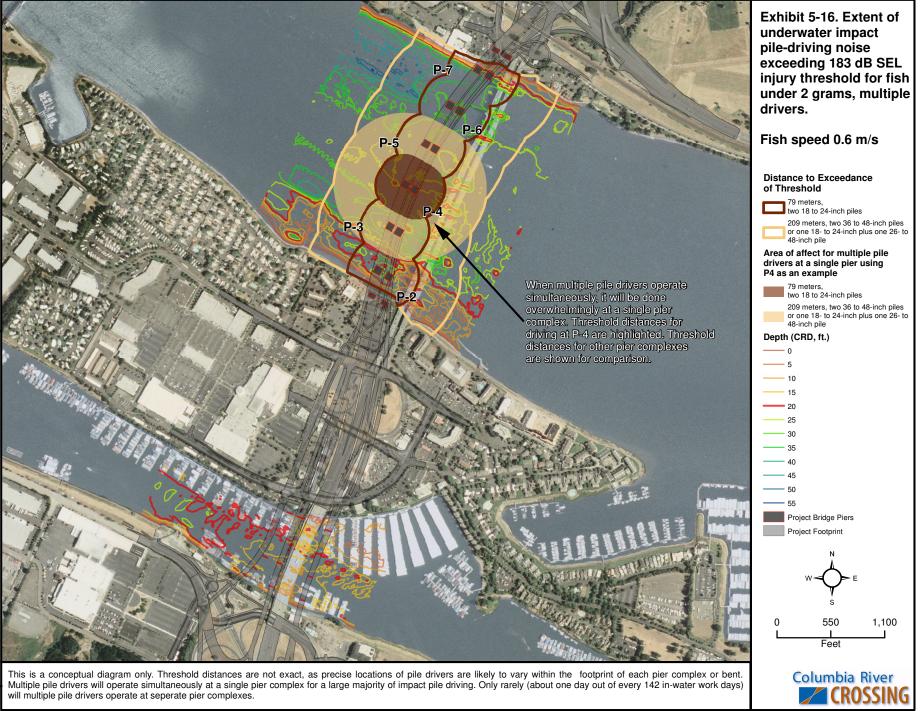
Note: Includes juvenile chum and larval eulachon.

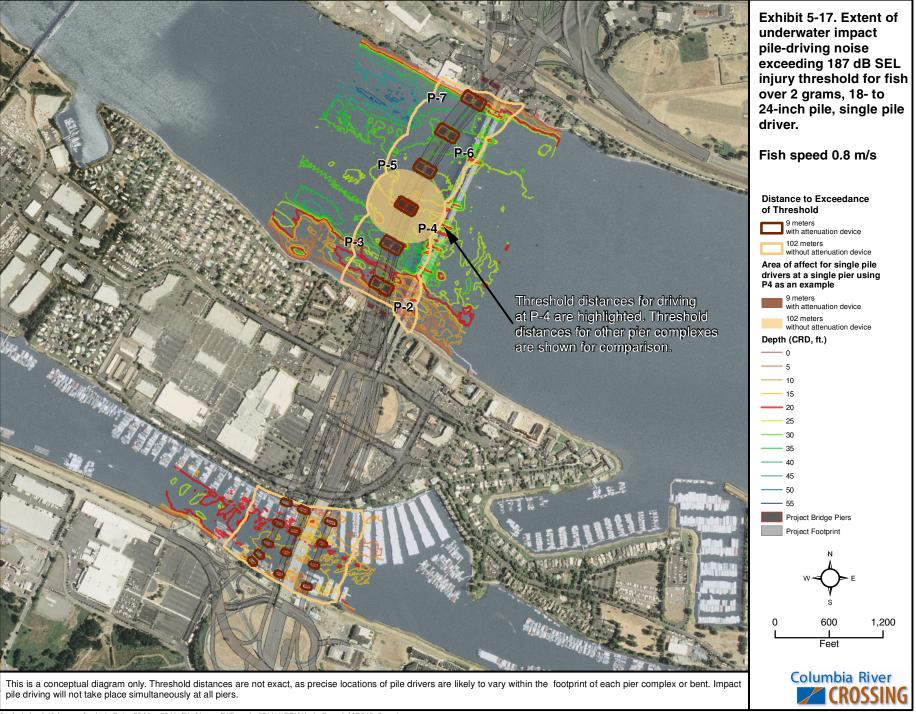
a Applies to Columbia River only.

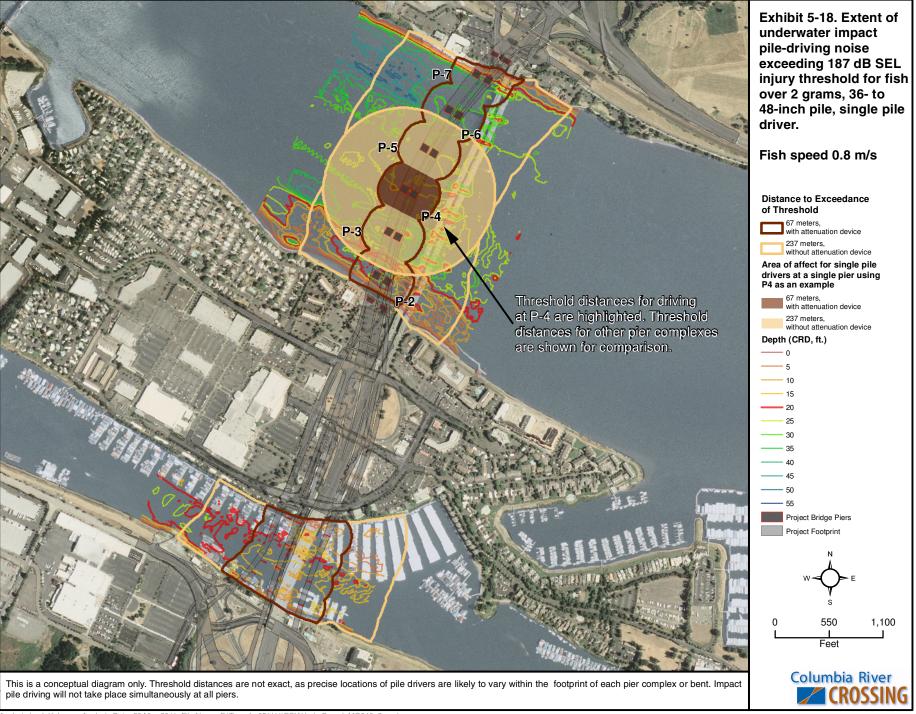
a Applies to Columbia River only.











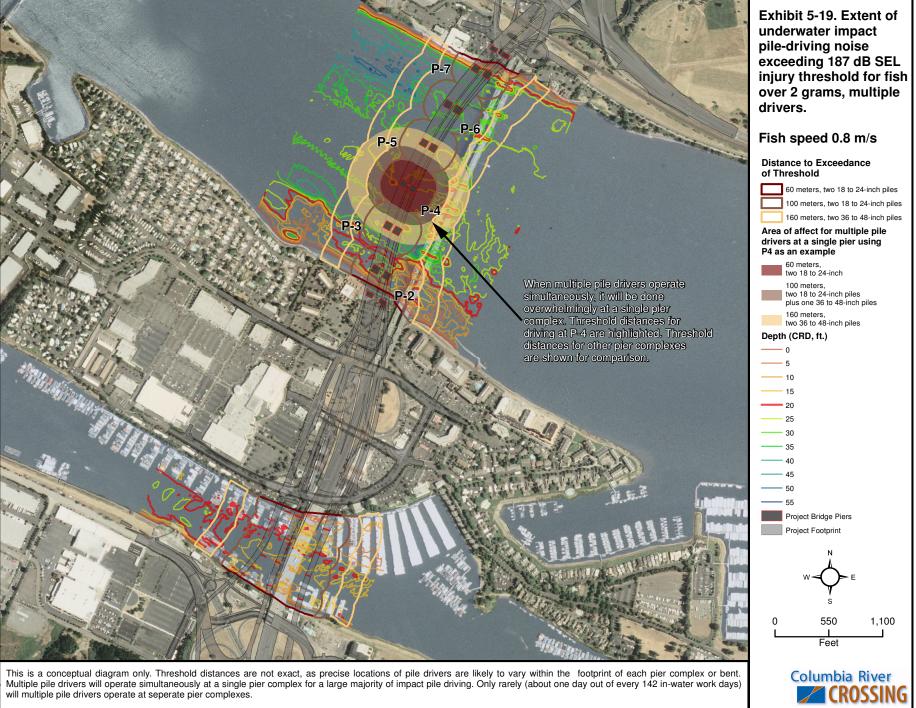
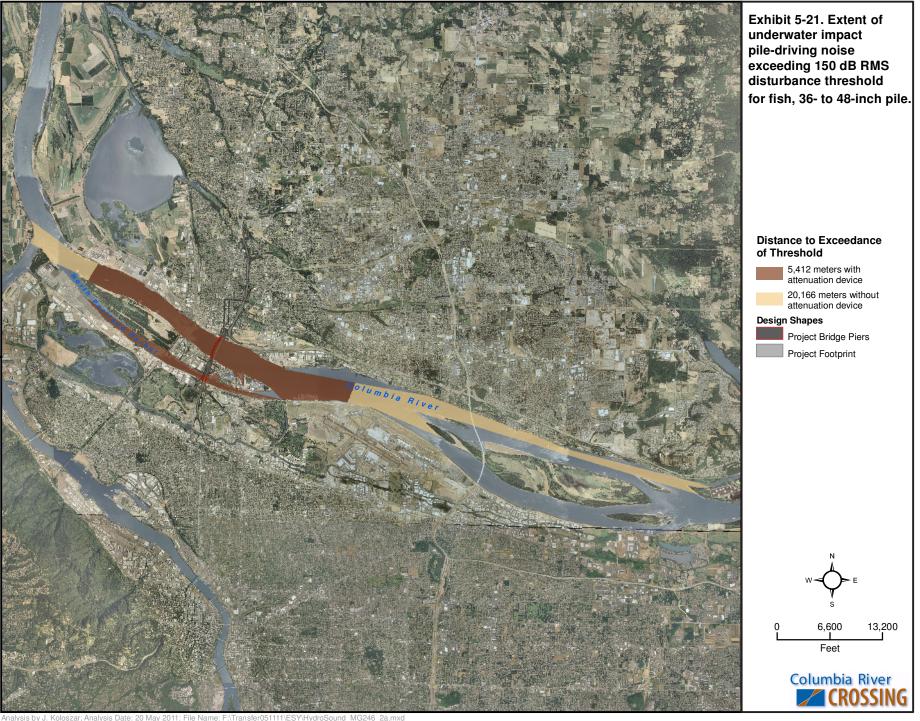
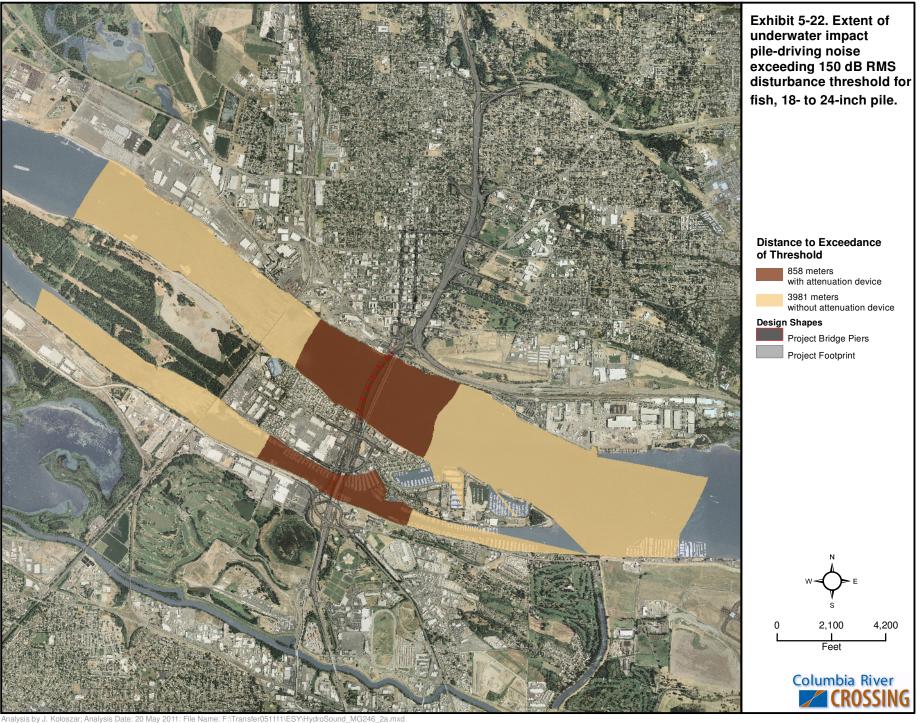


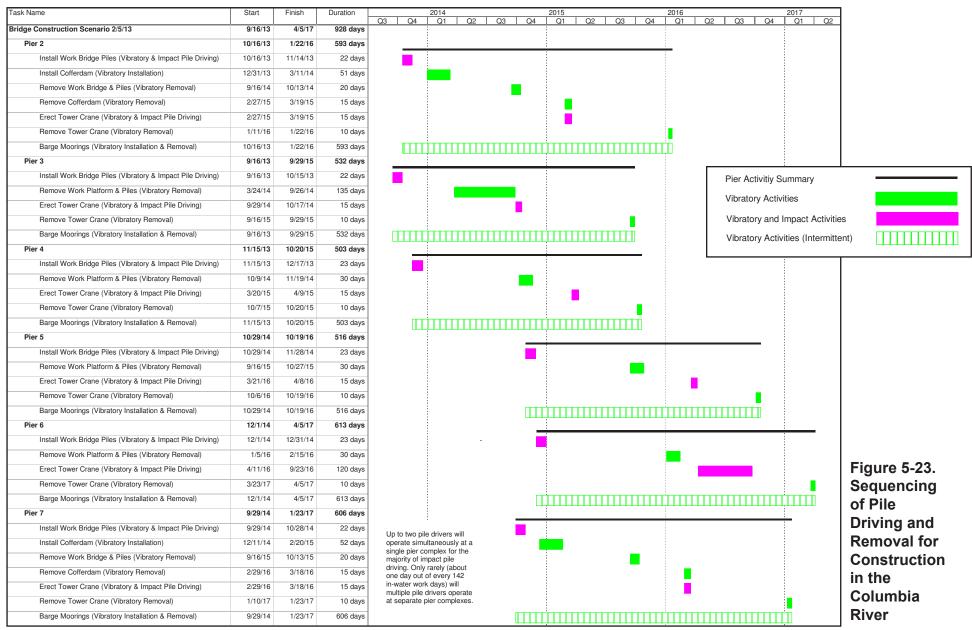
Exhibit 5-20, Exhibit 5-21, and Exhibit 5-22 show the distances within which noise is estimated to exceed the 150 dB RMS disturbance guidance.

Exhibit 5-20. Distances at Which Underwater Noise Exceeds 150 dB RMS Disturbance Guidance in the Columbia River and North Portland Harbor

	Colum	bia River	North Portland Harbor		
Impact Pile Driving	Distance Upstream (m)	Distance Downstream (m)	Distance Upstream (m)	Distance Downstream (m)	
Without Attenuation Device					
18- to 24-inch pile	3,981	3,981	3,058	3,981	
36- to 48-inch pile	20,166	8,851	3,058	5,632	
With Attenuation Device					
18- to 24-inch pile	858	858	858	858	
36- to 48-inch pile	5,412	5,412	3,058	5,412	



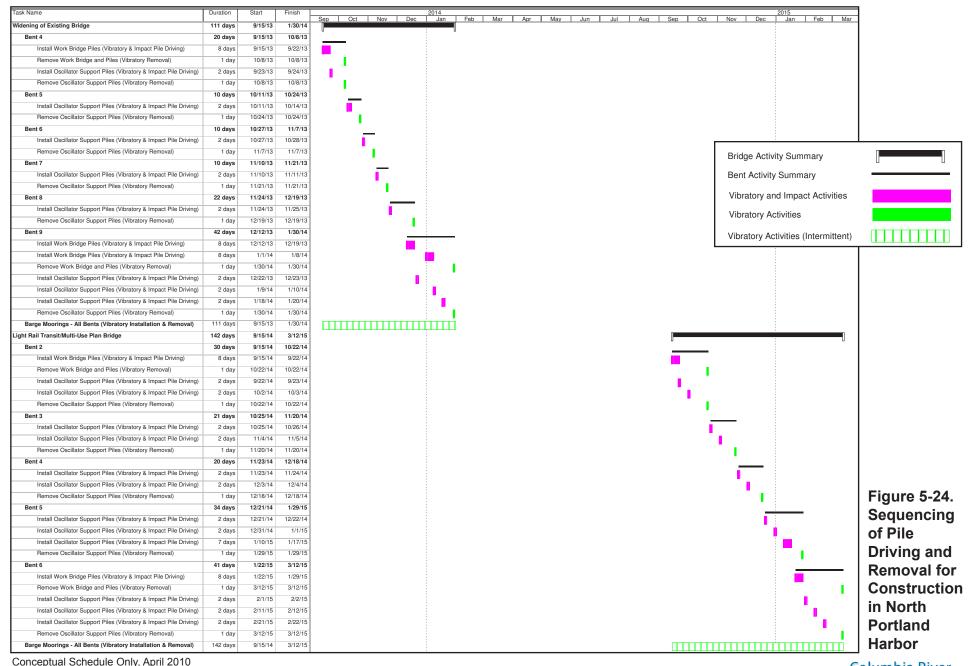




Conceptual Schedule Only, April 2010

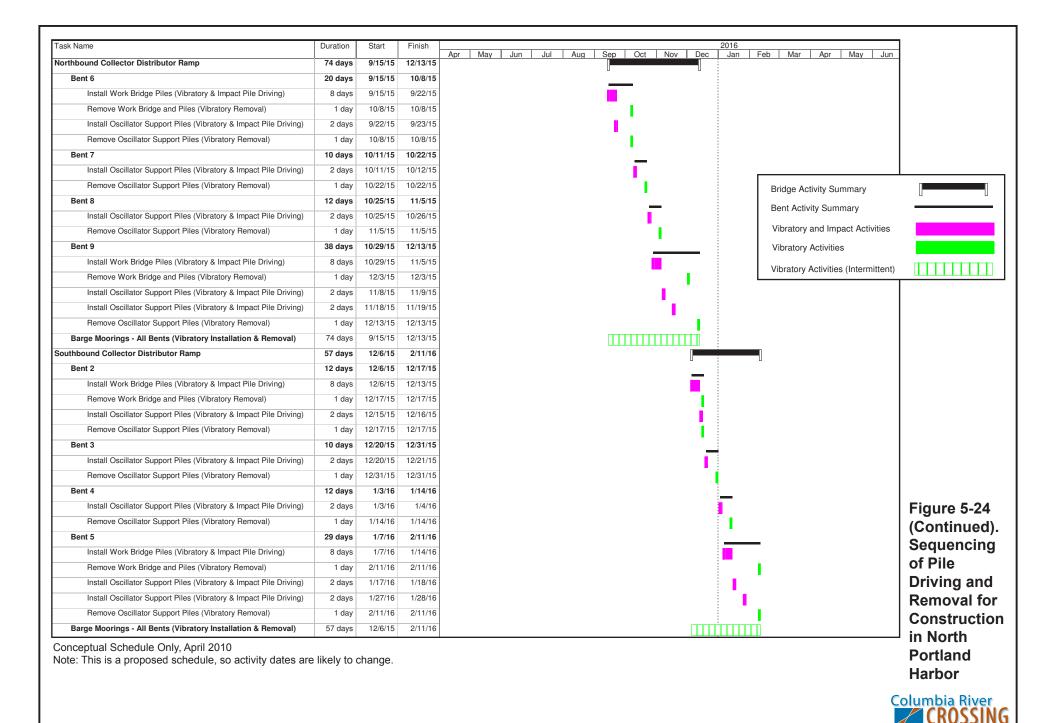
Note: This is a proposed schedule, so activity dates are likely to change.

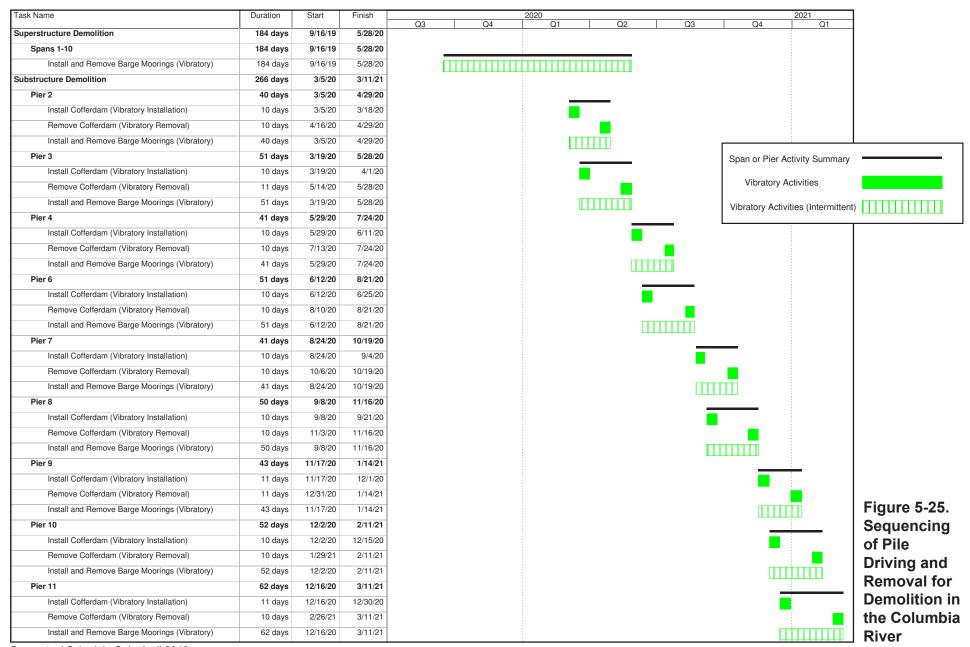




Note: This is a proposed schedule, so activity dates are likely to change.







Conceptual Schedule Only, April 2010

Note: This is a proposed schedule, so activity dates are likely to change.



Note that in most instances, use of an attenuation device decreases the area of effect appreciably. For example, when comparing scenarios in which a single pile driver is operating:

- The radius of the 206 dB peak injury zone decreases by about 80 percent.
- In the Columbia River, the radius of the disturbance zone decreases by about 80 percent for smaller piles and by 40 to 70 percent for larger piles, depending on the direction (upstream or downstream).
- In North Portland Harbor, radius of the disturbance zone decreases for smaller piles by about 75 percent. For the larger piles, use of a noise attenuation device does not shrink the disturbance zone because noise encounters landforms at fairly short distances from the source (3,058 meters upstream and 5,412 meters downstream).
- Similar reductions in distances to accumulated SEL threshold levels would occur with attenuation devices, but details are not presented here due to the numerous variables associated with calculating accumulated SEL in the moving fish model.

Exhibit 5-26 and Exhibit 5-27 summarize these results, showing the duration of impact and the areas in which noise levels would exceed the injury thresholds and disturbance guidance.

Exhibit 5-26. Exposure of Fish to Threshold/Guidance Levels of Underwater Noise in the Columbia River^a

		Without Attenuation Device ^a			With Attenuation Device (assumes 10 dB of attenuation)		
Size Class	Threshold or Guidance	Distance (m)	Duration	Number of Days	Distance (m)	Duration	Number of Days
≥ 2 grams	Injury: 206 dB Peak	25 - 34			5 - 7		
9	Injury ^b : 187 SEL _{cum} 0.1 m/sec 0.8 m/sec	113 - 243 102 - 237	7.5 min/		50 - 156 9 - 111	0.66 hr /	
	Disturbance: 150 dB RMS Upstream	3,981 - 20,166	week	38	858 - 5,412	day	138
	Downstream	3,981 - 8,851			858 - 5,412		
< 2	Injury: 206 dB Peak	25 - 34			5 - 7		
grams	Injury ^b : 183 SEL _{cum}	205 - 446			50 - 235		
	Disturbance: 150 dB RMS Upstream	3,981 - 20,166	7.5 min/ week	38	858 - 5,412	0.66 hr / day	138
	Downstream	3,981 - 8,851			858 - 5,412		

As part of the hydroacoustic monitoring program and to account for equipment failure, impact pile driving is assumed to occur for up to 7.5 minutes, one day per week.

b Accumulated SEL (injury) threshold distances are based on the construction scenario presented in Exhibit 5-3.

c Distances show extent of calculated values or where noise stops at landforms.

Exhibit 5-27. Exposure of Fish to Threshold/Guidance Levels of Underwater Noise in North Portland Harbor

		Without Attenuation Device		With Attenuation Device (assumes 10 dB of attenuation)			
Size Class	Threshold	Distance (m)	Duration	Number of Days	Distance (m)	Duration	Number of Days
≥ 2	Injury: 206 dB Peak	25 -34			5 - 7		
grams	Injury: 187 SEL _{cum} 0.1 m/sec 0.8 m/sec Disturbance: 150 dB RMS Upstream Downstream	113 - 243 102 - 237 3,058 3,981-5,632	2–5 min/ week	18 - 31	50 - 156 9 - 111 858 - 3,058 858 - 5,412	0.66 hr / day	134
< 2 grams	Injury: 206 dB Peak Injury: 183 SEL _{cum}	25 - 34 205 - 446	2 5 min/		5 - 7 50 - 235	0.66 hr /	
	Disturbance: 150 dB RMS Upstream Downstream	3,058 3,981-5,632	2–5 min/ week	18 - 31	858 - 3,058 858 - 5,412	day	134

a As part of the hydroacoustic monitoring program and to account for equipment failure, impact pile driving is assumed to occur for up to 3.75 minutes one day per week.

Impact pile driving would result in effects to fish that may range from behavioral disturbance to immediate death, depending on size of the fish, duration of exposure to sound pressure, proximity to the strike site, size of the pile, and number of strikes in a given time frame (e.g., per 12-hour period).

Actual exposure to noise above the injury thresholds and disturbance guidance would be fairly limited, restricted to the periods when impact pile driving is occurring: 138 days in the Columbia River and 134 days in North Portland Harbor interspersed over the entire four-year in water construction period from roughly mid-September through mid-April of each year (Exhibit 5-23 and Exhibit 5-24). Within this time period, exposure would be further restricted to no more than approximately 40 minutes per 12-hour work day.

Project-generated noise above the injury threshold may cause a range of lethal and sublethal injuries to fish, as outlined in Appendix K of the BA (CRC 2010). Effects may include damage to non-auditory tissues, including rupture of air-filled organs, such as the swim bladder. Damage to the swim bladder may lead to loss of control over vertical movement or may result in mortality. Loud noise may cause damage to the skin, nerves, and eyes of fish. Elevated sound levels may also result in the formation of gas bubbles in tissue, causing inflammation, cellular damage, and blockage or rupture of blood vessels. These injuries may lead to immediate or delayed mortality.

Intense sound may lead to hearing loss in fish. Such hearing loss may be temporary and reversible, known as temporary threshold shift (TTS). TTS and represents fatigue of the hair cells in the inner ear and is not considered tissue damage (Carlson et al. 2007). Intense sound may also reach levels that cause permanent threshold shift (PTS): permanent hearing loss resulting from the irreversible death of sensory hair cells in the inner ear. Such auditory damage may result in a general decrease in fitness, foraging success, ability to avoid predators, and ability to

b Accumulated SEL (injury) threshold distances are calculated based on the construction scenario presented in Exhibit 5-4.

c. Distances show extent of calculated values or where noise stops at landforms.

communicate. Thus, even if intense noises do not directly result in death, auditory damage could result in delayed mortality to fish.

Project-generated noise above the disturbance guidance may cause behavioral effects to fish. Literature related to the effect of pile driving on fish behavior is extremely limited and somewhat conflicting. Effects could be relatively minor, limited to startling, disruption in feeding, or avoidance of the project area (WSDOT 2008). Other effects could be more significant, with consequences for survival and reproduction. For example, while exposure to noise levels above 150 dB RMS is not likely to directly cause mortality or injury, it could result in an impaired ability to avoid predators, indirectly resulting in death (WSDOT 2008). Additionally, avoidance of the project area could presumably cause delays in migration for those species that migrate. Migration delays, in turn, may present a variety of risks for fish including: depletion of energy reserves; delayed or reduced spawning; increased exposure to predation, disease, and thermal stress; disruption of arrival timing to the estuary (which may desynchronize arrival with prey availability); and an increase in residualism in some steelhead and Chinook (NMFS 2009).

Lampreys do not have swimbladders and it is therefore difficult to determine the extent of this impact. Fish species without swimbladders are thought to be at lower risk from underwater sound than fishes with swimbladders (Stadler, pers. comm. 2010, Hastings and Popper 2005, Coker and Hollis 1952, Gaspin 1975, Baxter et al. 1982, Goertner 1994). Hydroacoustic impacts to lamprey should not be discounted, but they cannot be quantified or analyzed with any level of certainty.

Overall, this element of the project is likely to appreciably impact individuals of all listed salmon, steelhead, eulachon, and resident fish present in the areas exposed to noise above the injury threshold and disturbance guidance during impact pile-driving activities. Exhibit 5-28 summarizes the species and life stages of listed fish likely to be exposed to this effect.

Due to the extremely limited numbers of green sturgeon and bull trout present in the project area, risk of exposure is discountable. Thus, this element of the project is not likely to appreciably impact green sturgeon and bull trout.

Exhibit 5-28. Species and Life Stages Expected to be Present in the Project Area during Pile Driving

	Life Stage				
Species	Spawning	Incubation	Rearing	Outmigrating Juveniles	Migrating/ Holding Adults
Chinook					
Lower Columbia River ESU ^a			X	X	Χ
Upper Columbia River Spring-Run ESU			Х	Х	Х
Upper Willamette River ESU			Х	Х	Х
Snake River Fall-Run ESU				Х	X
Snake River Spring/Summer-Run ESU				X	X
Steelhead					
Lower Columbia River DPS ^a			Х	Х	Х
Middle Columbia River DPS				X	Х

			Life Sta	ge	
Species	Spawning	Incubation	Rearing	Outmigrating Juveniles	Migrating/ Holding Adults
Upper Willamette River DPS				Х	Х
Upper Columbia River DPS				X	Х
Snake River Basin DPS				X	X
Sockeye					
Snake River ESU				X	X
Coho					
Lower Columbia River ESU			Х	Х	Х
Chum					
Columbia River ESU	X	X	X	X	X
Bull Trout					
Columbia River DPS					X
Green Sturgeon					
Southern DPS					X
Eulachon					
Southern DPS	X	X		X	X
White sturgeon	Х	Х	Х	Х	Х
Lamprey Species			Х	Х	Х
Resident fish (e.g., sculpin, dace, threespine stickleback, sucker, shiner)	Х	Х	Х		

a ESU = evolutionarily significant unit; DPS = distinct population segment.

5.2.1.2 Hydroacoustic Impacts to Fish from Vibratory Pile Driving and Removal

Vibratory pile driving would be used to install cofferdams and temporary piles throughout the in-water project area in the Columbia River and North Portland Harbor. Load bearing piles (used for temporary work platforms, work bridges, tower cranes, and oscillator platforms) would be vibrated into place before being proofed with an impact hammer. Piles that are not load bearing (mooring piles) would be installed using vibration only.

Vibratory pile driving produces lower peak noise levels than impact pile driving of the same sized pile, and this generally results in fewer injuries to fish (USFWS 2009). Rise time is also much slower during vibratory pile driving, decreasing the potential for injury (Carlson et al. 2001, Nedwell and Edwards 2002, as cited in USFWS 2009). USFWS states that there are no documented kills attributed to the use of a vibratory hammer (USFWS 2004, as cited in WSF 2009).

Currently there are no established thresholds for noise levels generated by vibratory pile driving that are likely to cause injury or behavioral disturbance to fish. Additionally, there are no established threshold distances at which vibratory noise is likely to harm fish. However, NMFS offers the guidance that vibratory pile driving noise at 150 dB RMS may cause behavioral disturbance to fish.

Vibratory pile driving on the CRC project is likely to create noise above 150 dB RMS. Exhibit 5-29 outlines a range of typical noise levels produced by vibratory pile driving as measured by Caltrans during hydroacoustic monitoring of several construction projects (Caltrans 2009). The monitoring showed that vibratory driving of pipe pile (up to 72 inches in diameter) is likely to generate initial sound levels of up to 180 dB RMS, and vibratory driving of sheet pile is likely to generate initial sound levels of 160 to 165 dB RMS.

Exhibit 5-29. Summary of Unattenuated Underwater Sound Pressures for Vibratory Pile Driving

Pile Type and Approximate Size	Water Depth (meters)	SPLs (dB RMS)a
0.30-meter (12-inch) steel H-type	<5	150
0.30-meter (12-inch) steel pipe pile	<5	155
0.6-meter (24-inch) AZ steel sheet – typical	~15	160
0.6-meter (24-inch) AZ steel sheet – loudest	~15	165
1.0-meter (36-inch) steel pipe pile – typical	~5	170
1.0-meter (36-inch) steel pipe pile – loudest	~5	175
1.8-meter (72-inch) steel pipe pile – typical	~5	170
1.8-meter (72-inch) steel pipe pile – loudest	~5	180

Source: Caltrans 2009, CRC 2010.

On the CRC project, vibratory pile driving is likely to occur frequently during installation of temporary structures throughout the four-year in water construction period and the 18-month in-water demolition period. Vibratory pile driving for installation of temporary structures would likely take place up to approximately 5 hours per day during the in-water construction period and may occur during any hour of day.

Vibration may also be used to install the 10-foot-diameter steel casings for the drilled shafts of the permanent structures in the Columbia River and North Portland Harbor. No data were available regarding the initial SPLs generated by steel casings of this size. Therefore, it is not possible at this time to calculate the extent of noise generated from vibratory installation of 10-foot-diameter casings. However, it seems reasonable that vibration of the 10-foot steel casings would produce at least as many initial SPLs as 72-inch steel pipe pile (180 dB RMS at 5 meters), and therefore, noise from 10-foot casings would extend at least as far as that from 72-inch steel pipe pile. The design team estimates that vibratory installation of 10-foot casings would take approximately 90 days in the Columbia River and 31 days in North Portland Harbor. Vibratory installation of 10-foot casings is not restricted to the in water work window and therefore may take place any time during the four-year in-water construction period.

All of the species and life stages of salmon, steelhead, eulachon, and resident fish shown in Exhibit 5-28 could be exposed to this effect when they are present in this portion of the project area. However, fish kills attributed to the use of a vibratory hammer have never been documented (USFWS 2004, as cited in WSF 2009), this activity is unlikely to injure fish and is not expected to significantly interfere with behaviors such as migration, rearing, or foraging. Thus, vibratory pile driving is not likely to appreciably impact any of these species.

Due to the extremely limited numbers of green sturgeon and bull trout present in this portion of the project area, risk of exposure is discountable. Thus, this element of the project is not likely to appreciably impact green sturgeon or bull trout.

a Impulse level (35 millisecond average).

5.2.1.3 Noise Impacts to Fish from Excavation Drilled Shaft Casings

After the casings are installed, the project would excavate the material from inside of the permanent shafts. Hydroacoustic impacts from drilling and excavating inside of casings have not been well documented but would be far less than impacts from impact pile driving. Drilling shafts would likely elevate in-water noise levels, causing disturbance to fish, but the extent of this disturbance cannot be calculated. Lethal effects from drilling of shafts have not been documented on other projects and are not likely to occur. Shafts would be excavated year-round during the in water construction period (roughly, January 2014 to August 2017 in the Columbia River and September 2013 to February 2016 in North Portland Harbor). Effects to fish are expected to be insignificant.

5.2.2 Acoustic Impacts to Pinnipeds

5.2.2.1 Acoustic Effects from Pile Driving

Project-generated noise, including impact and vibratory pile driving, may have impacts to Steller sea lions, California sea lions, and harbor seals (referred to in this section collectively as pinnipeds), which migrate through the project area. The following sections present background information about how pinnipeds respond to noise, criteria for noise levels likely to cause injury or disturbance to sea lions, and an analysis of how pile-driving noise is likely to affect pinnipeds present in CRC project area.

5.2.2.2 How Pinnipeds Respond to Noise

There are few studies that quantify reactions of pinnipeds to noise, and even fewer that have directly observed reactions of pinnipeds to pile-driving noise (Southall et al. 2007). (Pinnipeds are a taxonomic category of marine mammals that includes seals and sea lions). Southall et al. (2007) performed a literature review of all known studies on the effects of noise on marine mammals. The review offers guidelines on how pinnipeds exhibit behavioral effects, temporary hearing loss, and injury resulting from elevated levels of underwater and airborne noise.

Behavioral Effects

Behavioral response to sound is dependent on a number of site-specific characteristics, including the intensity of the noise source, the distance between the noise source and the individual, and the ambient noise levels at the site (Southall et al. 2007). Behavioral response is also highly dependent on the characteristics of the individual animal. Marine mammals that have been previously exposed to noise may become habituated, and therefore may be less sensitive to noise. Such animals are less likely to elicit a behavioral response.

Behavioral responses have been observed experimentally and have been determined to be highly variable. In some cases, marine mammals may detect a sound and exhibit no obvious behavioral responses. In other cases, marine mammals may exhibit minor behavioral responses, including annoyance, alertness, visual orientation towards the sound, investigation of the sound, change in movement pattern or direction, habituation, alteration of feeding and social interaction, and temporary or permanent avoidance of the area affected by sound. Minor behavioral responses do not necessarily cause long-term effects to the individuals involved. Severe responses include panic, immediate movement away from the sound, and stampeding, which could potentially lead to injury or mortality (Southall et al. 2007).

In their comprehensive review of available literature, Southall et al. (2007) noted that quantitative studies on behavioral reactions of seals to underwater noise are rare. A subset of only three

studies observed the response of pinnipeds to underwater multiple pulses of noise (a category of noise types that includes impact pile driving) and were also deemed by the authors as having results that are both measurable and representative.

Harris et al. (2001) observed the response of ringed, bearded, and spotted seals to underwater operation of a single airgun and an 11-gun array. Received exposure levels were 160 to 200 dB RMS re: (referenced to) 1 μ Pa. Results fit into two categories. In some instances, seals exhibited no response to noise. However, the study noted significantly fewer seals during operation of the full array in some instances. Additionally, the study noted some avoidance of the area within 150 meters of the source during full array operations.

Blackwell et al. (2004) is the only study directly related to pile driving. The study observed ringed seals during impact installation of steel pipe pile. Received underwater SPLs were measured at 151 dB RMS re: 1 μ Pa at 63 meters. The seals exhibited either no response or only brief orientation response (defined as "investigation or visual orientation"). It should be noted that the observations were made after pile driving was already in progress. Therefore, it is possible that the low-level response was due to prior habituation.

Miller et al. (2005) observed responses of ringed and bearded seals to a seismic airgun array. Received underwater sound levels were estimated at 160 to 200 dB RMS re: 1 μ Pa. There were fewer seals present close to the noise source during airgun operations in the first year, but in the second year the seals showed no avoidance. In some instances, seals were present in very close range of the noise. The authors concluded that there was "no observable behavioral response" to seismic airgun operations.

Southall et al. (2007) conclude that there is little evidence of avoidance of SPLs from pulsed noise ranging between 150 and 180 dB RMS re: 1 µPa. Additionally, they conclude that behavioral response in ringed seals is likely to occur at 190 dB RMS. It is unclear whether or not these data apply to Steller and California sea lions. Given that there are so few data available, it is difficult to draw conclusions about what specific behaviors pinnipeds would exhibit in response to underwater noise.

Southall et al. (2007) also compiled known studies of behavioral responses of marine mammals to airborne noise, noting that studies of pinniped response to airborne pulsed noises are exceedingly rare. The authors deemed only one study as having quantifiable results.

Blackwell et al. (2004) studied the response of ringed seals within 500 meters of impact driving of steel pipe pile. Received levels of airborne noise were measured at 93 dB RMS re: $20 \mu Pa$ at a distance of 63 meters. Seals had either no response or limited response to pile driving. Reactions were described as "indifferent" or "curious."

Due to the extremely limited data on this topic, it is not possible to draw definitive conclusions about what specific behaviors pinnipeds would exhibit in response to airborne noise generated by impact pile driving.

Several field observations indicate that sea lions exhibit mixed responses to elevated noise levels.

During a Caltrans installation demonstration project for retrofit work on the East Span of the San Francisco Oakland Bay Bridge, California, sea lions responded to pile driving by swimming rapidly out of the area, regardless of the size of the pile-driving hammer or the presence of sound attenuation devices (74 FR 63724).

Dyanna Lambourn, marine mammal research biologist at WDFW, noted that Steller sea lions generally avoid unfamiliar loud noises. In response to pile driving, they would be likely to exit areas exposed to elevated noise, unless there were a particularly strong attraction, such as an abundant food source (Lambourn 2010 pers. comm.). Lambourn also stated that Steller sea lions could become habituated to noises that are continuous and occurring over longer periods of time.

The USACE has conducted hazing of sea lions at Bonneville Dam since 2004 in an attempt to decrease rates of predation on listed salmonids and sturgeon. The 2010 monitoring report (Stansell et al. 2010) documented the response of both California and Steller sea lions to several types of deterrents, including Acoustic Deterrent Devices (ADDs). These devices produce noise levels of 205 dB in the frequency range of 15 kHz. (The report did not specify whether these values referred to airborne or underwater noise). The crews also employed above-water pyrotechnics (cracker shells, screamer shells, or rockets) and underwater percussive devices called seal bombs. Hazing occurred seven days a week from March 2 to the end of May. The study did not differentiate between Steller sea lions and California sea lions, so it is uncertain whether these two species respond differently to hazing.

The observers reported that sea lions tended to spend more time underwater and temporarily avoided the area while hazing activities were occurring, but returned to forage soon after the activities ceased. They concluded that hazing only slowed the rate of predation, rather than effectively deterring it. The sea lions slightly shifted foraging times, preying more heavily at dawn and dusk, when hazing activities were beginning or ending. Nevertheless, despite active hazing, the rate of predation on salmon and sturgeon was still quite high. Observers noted that sea lions swam to within 20 feet of the ADDs to forage.

The explosive and percussive noises produced during these hazing activities are quite different from pile-driving noise, as they are abrupt and non-pulsed. These results may not be applicable to pile-driving projects; however, the results were included to demonstrate that high SPLs alone do not necessarily cause significant behavioral responses in sea lions. Also, the study is specific to sea lion behavior in the lower Columbia River, and it observed the same individuals that transit through the CRC project area. The results suggest that these individuals either are already habituated to some loud noises or could readily become habituated.

Temporary Threshold Shift

Temporary Threshold Shift (TTS) is reversible hearing loss caused by fatigue of hair cells and supporting structures in the inner ear. Technically, TTS is not considered injury, as it consists of fatigue to auditory structures rather than damage to them. Pinnipeds have demonstrated complete recovery from TTS after multiple exposures to intense noise, as described in the studies below (Kastak et al. 1999, 2005).

There are no studies of the underwater noise levels likely to cause TTS in Steller sea lions. However, TTS studies have been conducted on harbor seals, California sea lions, and northern elephant seals. Southall et al. (2007) report several studies on non-pulsed noise (a category that includes vibratory pile-driving noise), but only one study on pulsed noise.

• Finneran et al. (2003) studied responses of two individual California sea lions. The sea lions were exposed to single pulses of underwater noise, and experienced no detectable TTS at received noise level of 183 dB peak re: 1 μPa, and 163 dB SEL re: 1 μPa²-s.

There were three studies of pinniped TTS in response to non-pulsed underwater noise. All of these studies were performed in the same lab and on the same test subjects, and therefore the results may not be applicable to all pinnipeds or in field settings.

- Kastak and Schusterman (1996) studied the response of harbor seals to non-pulsed construction noise, reporting TTS of about 8 dB.
- Kastak et al. (1999) exposed a harbor seal, California sea lion, and elephant seal to octave-band noise at 60 to 70 dB above their hearing thresholds. After 20 to 22 minutes, the subjects experienced TTS of 4 to 5 dB.
- Kastak et al. (2005) used the same test subjects above, exposing them to higher levels of
 noise for longer durations. The animals were exposed to octave-band noise for up to
 50 minutes of net exposure.
 - The study reported that the harbor seal experienced TTS of 6 dB after a 25-minute exposure to 2.5 kHz of octave-band noise at 152 dB re: 1 μ Pa and 183 dB SEL re: 1 μ Pa2-s.
 - The California sea lion demonstrated onset of TTS after exposure to 174 dB re: 1 μ Pa and 206 dB SEL re: 1 μ Pa2-s.
 - The northern elephant seal demonstrated onset of TTS after exposure to 172 dB re: 1 μ Pa and 204 dB SEL re: 1 μ Pa2-s.

Combining the above data, Southall et al. (2007) assume that pulses of underwater noise result in the onset of TTS in pinnipeds when underwater noise levels reach 212 dB peak or 171 dB SEL. They did not offer criteria for non-pulsed sounds.

Southall et al. 2007 reported only one study of TTS in pinnipeds resulting from airborne pulsed noise:

• Bowles et al. (unpubl. data) exposed pinnipeds to simulated sonic booms. Harbor seals demonstrated TTS at 143 dB peak re: 20 μPa and 129 dB SEL re: 20 μPa²-s. California sea lions and northern elephant seals experienced TTS at higher exposure levels than the harbor seals.

Two studies examined TTS in pinnipeds resulting from airborne non-pulsed noise. These studies may not be relevant to the CRC project, but are provided for general reference.

- Kastak et al. (2004) used the same test subjects as in Kastak et al. 2005, exposing the animals to non-pulsed noise (2.5 kHz octave-band noise) for 25 minutes.
 - The harbor seal demonstrated 6 dB of TTS after exposure to 99 dB re: 20 μ Pa and 131 dB SEL re: 20 μ Pa2-s.
 - The California sea lion demonstrated onset of TTS at 122 dB re: $20~\mu Pa$ and 154 dB SEL re: $20~\mu Pa2$ -s.
 - The northern elephant seal demonstrated onset of TTS at 121 dB re: 20 μ Pa and 163 dB SEL re: 20 μ Pa2-s.
- Kastak et al. (2007) studied the same California sea lion as in Kastak et al. 2004 above, exposing this individual to 192 exposures of 2.5 kHz octave-band noise at levels ranging from 94 to 133 dB re: 20 μPa for 1.5 to 50 minutes of net exposure duration. The test subject experienced up to 30 dB of TTS. TTS onset occurred at 159 dB SEL re: 20 μPa²-s. Recovery times ranged from several minutes to 3 days.

Southall et al. (2007) assume that multiple pulses of airborne noise result in the onset of TTS in pinnipeds when levels reach 143 dB peak or 129 dB SEL.

Lambourn (2010) noted that, in a field setting, sea lions are unlikely to remain in areas exposed to noise levels high enough to cause hearing loss, unless there is a particular attraction keeping them in the area.

Injury - Permanent Threshold Shift

Permanent threshold shift (PTS) is irreversible loss of hearing sensitivity at certain frequencies caused by exposure to intense noise. It is characterized by injury to or destruction of hair cells in the inner ear. Southall et al. (2007) note that there are no empirical studies demonstrating the noise levels that prompt PTS in marine mammals. Furthermore, they found that there is virtually no understanding of the relationship between TTS and PTS in marine mammals, as no studies have been performed.

Southall et al. (2007) propose that noise levels inducing 40 dB of TTS may result in onset of PTS in marine mammals. The authors present this threshold with precaution, as there are no specific studies to support it. Because direct studies on marine mammals are lacking, the authors base these recommendations on studies performed on other mammals. Additionally, the authors assume that multiple pulses of underwater noise result in the onset of PTS in pinnipeds when levels reach 218 dB peak or 186 dB SEL. In air, noise levels are assumed to cause PTS in pinnipeds at 149 dB peak or 144 dB SEL (Southall et al. 2007).

5.2.2.3 Criteria for Injury and Disturbance

NMFS is currently developing comprehensive guidance on sound levels likely to cause injury and behavioral disruption in the context of the MMPA. Until formal guidance is available, NMFS uses conservative thresholds of sound pressure levels likely to cause injury or disturbance to sea lions (Exhibit 5-30) (NMFS 2009; WSDOT 2009a).

Exhibit 5-30. Injury and Disturbance Thresholds for Pinnipeds

Location	Threshold
Underwater – impact pile driving	Injury: 190 dB RMS re: 1 µPa
	Disturbance: 160 dB RMS re: 1 µPa
Underwater – vibratory pile driving	Injury: None designated
	Disturbance: 120 dB RMS re: 1 µPa
Above water	Injury: None designated
	Disturbance: 90 dB RMS re: 20 µPa (unweighted) for harbor seals
	100 dB $_{RMS}$ re: 20 μPa (unweighted) for all other pinnipeds

Source: NMFS (2009), WSDOT (2009a).

5.2.2.4 Estimating Noise Levels and Acoustic Area of Effect

The extent of in-water and airborne project-generated noise was calculated for the locations where pile driving would occur in the Columbia River and North Portland Harbor.

The extent of underwater noise was modeled for several pile driving scenarios:

- For two sizes of pile: 18- to 24-inch pile and 36- to 48-inch pile.
- For impact pile drivers operating both with and without an attenuation device. Use of an attenuation device was assumed to decrease initial SPLs by 10 dB.
- For all vibratory pile driving of pipe pile and sheet pile used for temporary structures.

Although two impact pile drivers would operate simultaneously in close proximity to one another in the Columbia River, the two drivers are not expected to generate noise levels greater than a single pile driver. Pile strikes from both drivers would need to be synchronous (within 0.0 and approximately 0.1 second apart) in order to produce higher noise levels than a single pile driver operating alone. Because it is highly unlikely that two pile drivers would operate in exact synchronicity, we assume that two pile drivers would not generate noise levels greater than that of a single pile driver. Therefore, initial noise levels for multiple pile drivers are assumed to be the same as for a single pile driver.

No data were available regarding the initial SPLs generated by vibratory installation of 10-foot diameter steel casings that are proposed for the drilled shafts. Therefore, the project team extrapolated initial SPLs from published values, as described in the subsection on vibratory pile driving below.

The extent of airborne noise was modeled for impact pile driving only.

Impact Pile Driving - Underwater Noise

Exhibits 5-31 and 5-32 quantify the extent, timing, and duration of impact pile-driving noise that would exceed threshold levels for disturbance and injury to seals and sea lions. Impact pile driving is expected to take place only within a 31-week in-water work window, ranging from Week 38 of one year to Week 16 of the next (or approximately from September 15 to April 15) over the bridge construction period. There will be a total of about 138 days of impact pile driving in the Columbia River and about 134 days of impact pile driving in North Portland Harbor for the entire project from the start of bridge construction in 2013 to its anticipated completion in 2017 (approximately 4.25 years for both Columbia River and North Portland Harbor bridges). Impact pile driving in the mainstem Columbia River would occur at more than one pier complex on about 1 or 2 days total during the course of the approximately 4-year construction period. Impact pile driving would be restricted to approximately 45 minutes per 12-hour work day. After initial hydroacoustic monitoring to test its effectiveness, a noise attenuation device would be used during all other impact pile driving. Each work day would include a period of at least 12 consecutive hours with no impact pile driving in order to minimize disturbance to aquatic animals. Impact pile driving would only occur during daylight hours.

Exhibit 5-31. Summary of Impact Pile-Driving Noise Above 190 dB RMS Underwater Injury Threshold

	Columbia River			North Portland Harbor			
Pile Size and Number	Distance (m)	Duration	No. Days	Distance (m)	Duration	No. Days	
Without Attenuation Device							
18- to 24-inch pile	9	7.5 min/week	38	9	2.5 – 5 min/week	18	
36- to 48-inch pile	54	7.5 min/week	38	54	2.5 – 5 min/week	31	
With Attenuation Device							
18- to 24-inch pile	2	45 min/day	138	2	45 min/day	72	
36- to 48-inch pile	12	45 min/day	138	12	45 min/day	62	

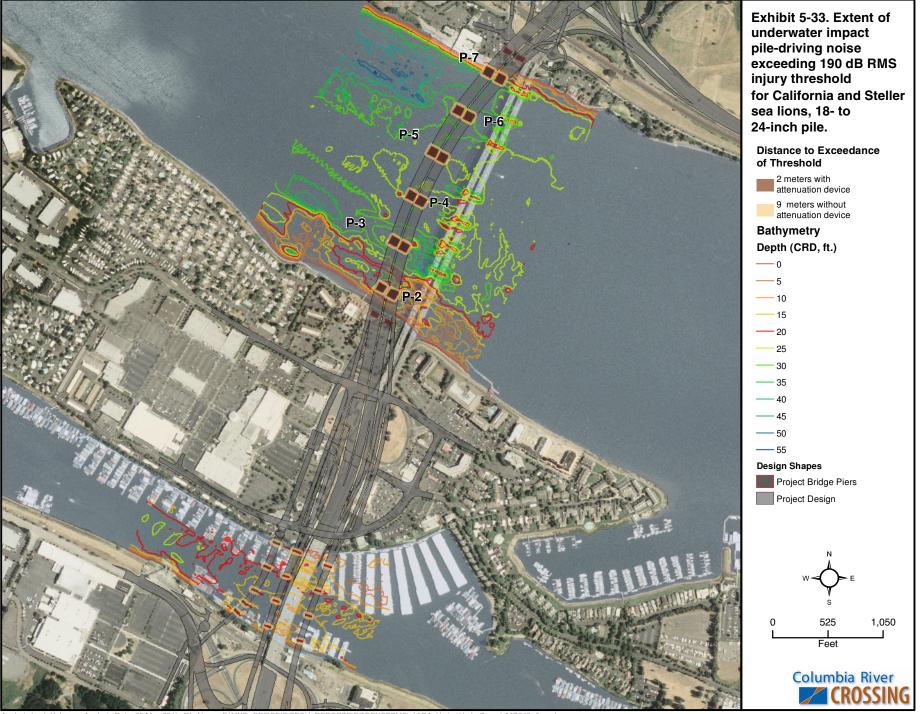
Note: Elevated noise levels would occur throughout the 5-year in-water work period. Potential exposure may only occur from approximately September through May, the range of months when seals and sea lions have been observed at Bonneville Dam and could be present in the project area. However, seals and sea lions would actually not be exposed to injurious levels of noise, because impact pile driving would stop when seals and sea lions approach the injury isopleth, i.e. the area where underwater noise is at or above 190 dB RMS.

Exhibit 5-32. Summary of Impact Pile-Driving Noise Above 160 dB RMS Underwater Disturbance Threshold

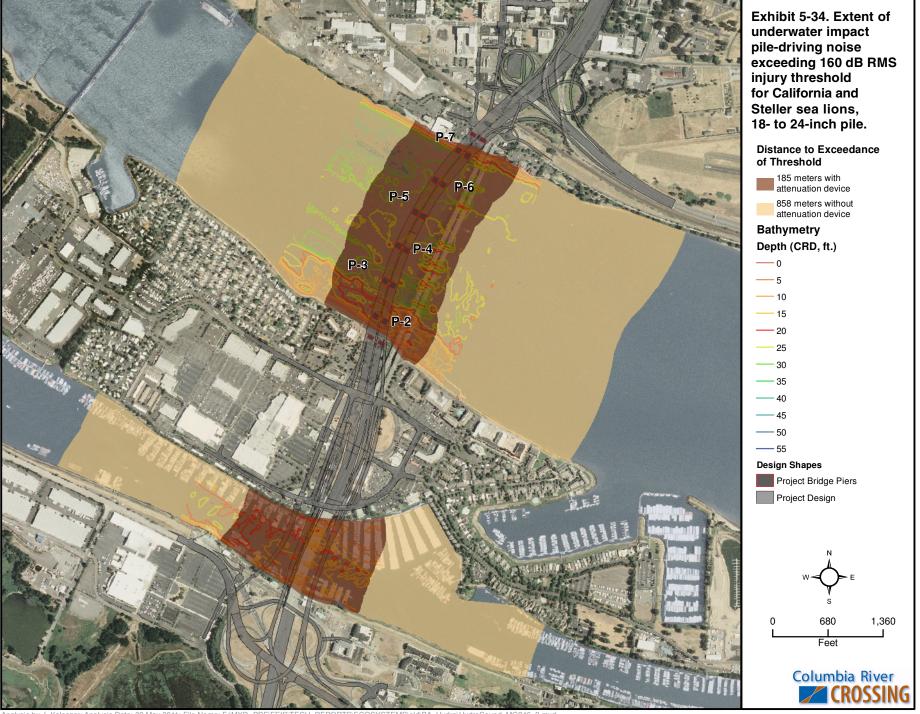
	Columbia River			North Portland Harbor		
Pile Size and Number	Distance (m)	Duration	No. Days	Distance (m)	Duration	No. Days
Without Attenuation Device						
18- to 24-inch pile	858	7.5 min/week	38	858	2.5 – 5 min/week	18
36- to 48-inch pile	5,412	7.5 min/week	38	3,058 – u 5,412 – d	2.5 – 5 min/week	31
With Attenuation Device						
18- to 24-inch pile	185	45 min/day	138	185	45 min/day	72
36- to 48-inch pile	1,166	45 min/day	138	1,166	45 min/day	62

Note: Elevated noise levels would occur throughout the 5-year in-water work period. Potential exposure may only occur from approximately September through May, the range of months when seals and sea lions have been observed at Bonneville Dam and could be present in the project area.

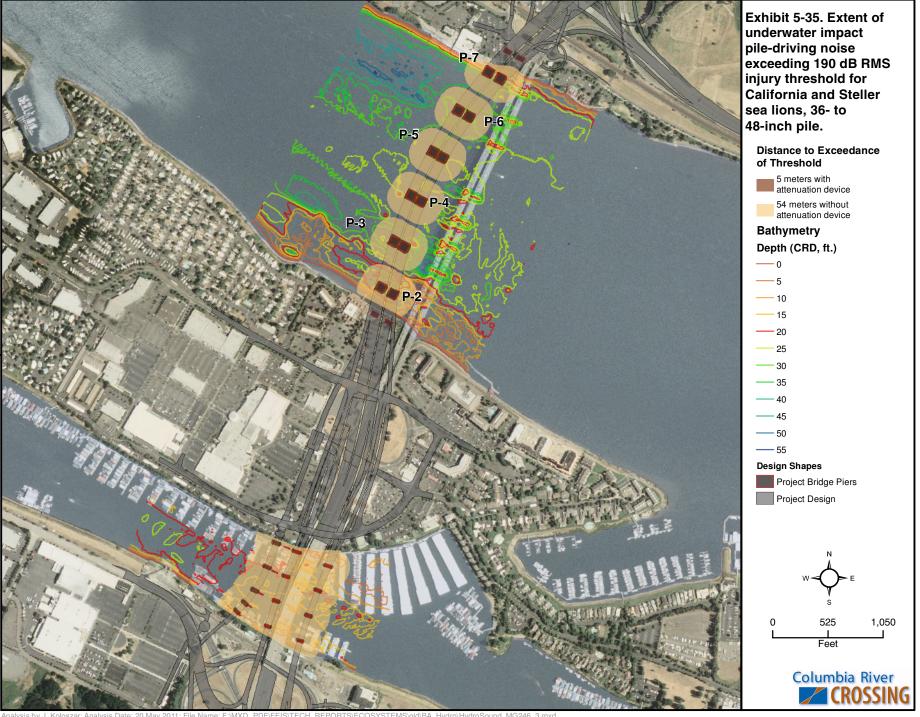
u = upstream, d = downstream



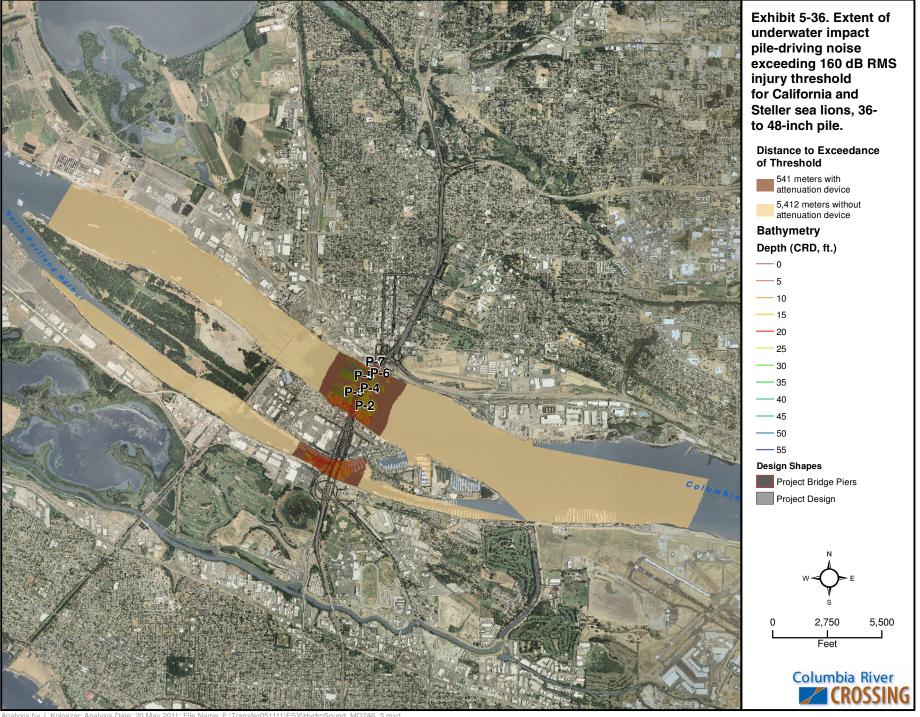
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For 18- to 24-inch pile in both water bodies, and for 36- to 48-inch pile in the Columbia River, the actual, site-specific distances are the same as the calculated distances (Exhibit 5-33 and Exhibit 5-35).

Vibratory Pile Driving - Underwater Noise

No studies were available that measured site-specific initial noise levels generated by vibratory pile driving in the CRC project area. However, Exhibit 5-37 outlines a range of typical noise levels produced by vibratory pile driving as measured by Caltrans during hydroacoustic monitoring of several construction projects (Caltrans 2009).

Exhibit 5-37. Summary of Unattenuated Underwater Sound Pressures for Vibratory Pile Driving

Pile Type and Approximate Size	Water Depth	SPLs (dB RMS) ^a
0.30-meter (12-inch) steel H-type	<5 meters	150
0.30-meter (12-inch) steel pipe pile	<5 meters	155
1-meter (36-inch) steel pipe pile – typical	~5 meters	170
0.6-meter (24-inch) AZ steel sheet - typical	~15 meters	160
0.6-meter (24-inch) AZ steel sheet - loudest	~15 meters	165
1-meter (36-inch) steel pipe pile – loudest	~5 meters	175
1.8-meter (72-inch) steel pipe pile – typical	~5 meters	170
1.8-meter (72-inch) steel pipe pile – loudest	~5 meters	180

Source: Caltrans 2009, CRC 2010.

Pipe Pile

Exhibit 5-38 summarizes the extent, timing, and duration of noise above the 120 dB RMS disturbance threshold during vibratory installation of pipe pile and sheet pile. Vibratory installation of pipe pile would be likely to occur throughout the entire 5-year construction period of all new in-water piers or bents and for installation of mooring piles. Vibratory installation of sheet pile would only occur in the Columbia River during construction of the new Columbia River bridges and demolition of the existing Columbia River bridges. This activity would occur intermittently throughout the construction and demolition period. This activity would not be restricted to an in-water work window and therefore may take place during any time of the year.

Exhibit 5-38. Summary of Vibratory Pile -Driving Noise Above 120 dB RMS Underwater Disturbance Threshold – Pipe Pile and Sheet Pile

			Columbia River		North	n Portland Ha	arbor
Pile Type	Timing	Distance (m)	Hours/ Day	No. Days	Distance (m)	Hours/D ay	No. Days
Pipe Pile	Year-round	20,166 – u 8,851 – d	Up to 5	1,470– 1,620	3,058 – u 5,632 – d	Up to 5	~334
Sheet Pile	Year-round	6,962	Up to 24	99	N/A	N/A	N/A

Note: Elevated noise levels will occur throughout the in-water work period. Potential exposure may only occur from approximately September through May, the range of months when seals and sea lions have been observed at Bonneville Dam and could be present in the project area.

a Impulse level (35 millisecond average).

u = upstream, d = downstream

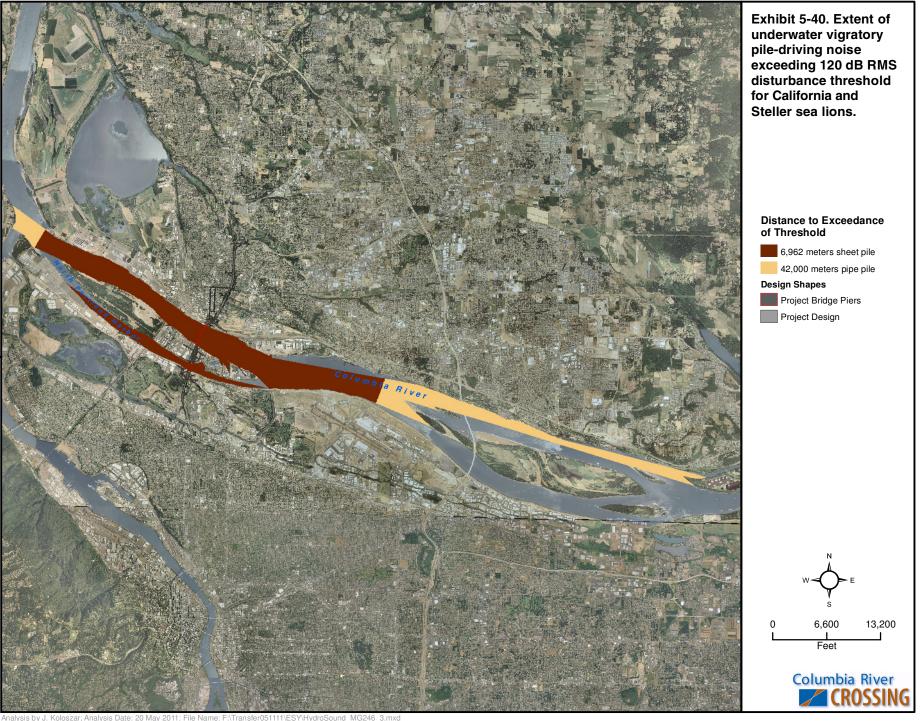
5.2.3 Vibratory Installation of Steel Casings

If steel casings for drilled shafts are vibrated into place, the CRC project design team estimates that installation of the 10-foot diameter casings will take approximately 90 days in the Columbia River and 31 days in North Portland Harbor. Vibratory installation of casings is not restricted to the in-water work window and therefore may take place any time during the in-water construction period. Exhibit 5-39 summarizes the estimated extent, timing, and duration of noise above the injury and disturbance thresholds during vibratory installation of steel casings. Hydroacoustic monitoring will be conducted to field verify the distances within which noise exceeds these thresholds.

Exhibit 5-39. Summary of Vibratory Pile-Driving Noise above Disturbance and Injury Thresholds – Steel Casings

		Columbia River		North Portla	and Harbor
Threshold	Timing	Distance (m)	No. Days	Distance (m)	No. Days
120 dB RMS	Year-round	20,166 – u 8,851 – d	90	3,058 – u 5,632 – d	31
190 dB RMS	Year-round	5	90	5	31

Note: Elevated noise levels will occur throughout the 4-year in-water construction period. Potential exposure may only occur from approximately September through May, the range of months when seals and sea lions have been observed at Bonneville Dam and could be present in the project area.u = upstream, d = downstream



Steel Casings

Vibration may also be used to install the 10-foot-diameter steel casings for the drilled shafts of the permanent structures in the Columbia River and North Portland Harbor. No data were available regarding the initial SPLs generated by installation of steel casings of this size. Therefore, the design team extrapolated from published values, assuming that vibratory driving of 10-foot casings would generate noise at levels of up to 10 dB RMS (an order of magnitude) higher than the highest value for vibratory installation of a 72-inch pile (Exhibit 5-37). That is, vibratory installation of 10-foot diameter steel casing may yield a maximum value of 190 dB RMS at 5 meters from the pile.

Therefore, it is assumed that vibratory installation of 10-foot-diameter steel pile would exceed the 190 dB RMS injury threshold for sea lions at 5 meters from the source (Exhibit 5-41). Exhibit 5-41 also shows the distance within which noise is calculated to attenuate to the 120 dB RMS vibratory pile driving disturbance threshold, as per the Practical Spreading Model.

Exhibit 5-41. Distance to Underwater Noise Thresholds from Source for Vibratory Driving of Steel Casings

	Distance from Source (m)
Estimated Noise Level (dB RMS)	Initial SPL 190 dB RMS at 5 m
190 (injury threshold)	5
120 (disturbance threshold)	233,000

Landforms in the Columbia River and North Portland Harbor would completely block underwater noise well before it reaches the 233,000-meter distance calculated for the 120 dB RMS disturbance threshold. Exhibit 5-42 shows site-specific values for the maximum distance at which noise is likely to exceed the injury and disturbance thresholds.

Exhibit 5-42. Distance to Underwater Noise Thresholds for Vibratory Driving of Steel Casings – Site-Specific Values

	Distance from Source Distance from Source (m)			
Estimated Noise Level (dB RMS)	Columbia River	North Portland Harbor		
190 (injury threshold)	5	5		
120 (disturbance threshold)	20,166 Upstream 8,851 Downstream	3,058 Upstream 5,632 Downstream		

Without a precise estimate of initial SPLs, the values shown in Exhibit 5-42 are rough estimates. To refine these estimates, the CRC team proposes to perform hydroacoustic monitoring during vibratory installation of the first steel casing in order to verify: 1) the initial SPLs generated by this activity and 2) the potential injury zone for sea lions. Additionally, hydroacoustic monitoring is likely to be required under the terms of a Letter of Authorization issued by NMFS under the Marine Mammal Protection Act.

Airborne Noise

Exhibits 5-43 and 5-44 summarize the extent, timing, and duration of airborne impact pile-driving noise above disturbance thresholds for sea lions and seals, respectively. Airborne noise effects would occur on the same schedule as those described for impact pile driving above.

Exhibit 5-43. Summary of Impact Pile-Driving Noise Above 100 dB RMS Airborne Disturbance Threshold for Sea Lions

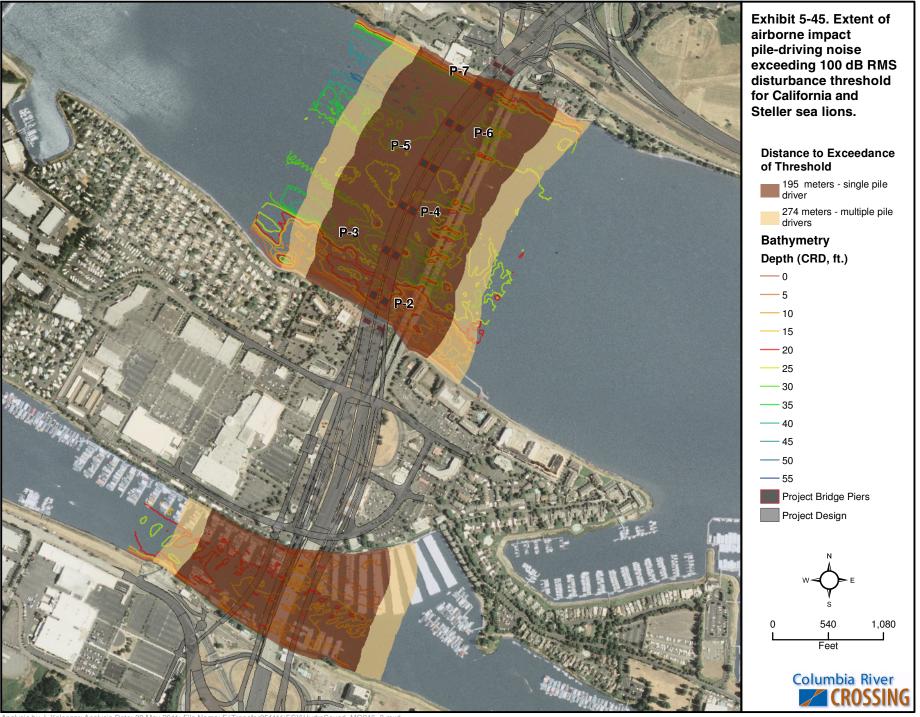
Location	Distance from Source (m)	Min/Day	No. Days
Columbia River	196	≤45	138
North Portland Harbor	196	≤45	134

Note: Elevated noise levels will occur throughout the approximately 4-year in-water work period. Potential exposure may only occur from approximately September through May, the range of months when seals and sea lions have been observed at Bonneville Dam and could be present in the project area.

Exhibit 5-44. Summary of Exposure to Impact Pile-Driving Noise Above 90 dB RMS Airborne Noise Disturbance Threshold for Harbor Seals

Location	Distance from Source (m)	Min/Day	No. Days
Columbia River	650	≤45	138
North Portland Harbor	650	≤45	134

Note: Elevated noise levels will occur throughout the approximately 4-year in-water work period. Potential exposure may only occur from approximately September through May, the range of months when seals and sea lions have been observed at Bonneville Dam and could be present in the project area.



Analysis by J. Koloszar; Analysis Date: 20 May 2011: File Name: F:\Transfer051111\ESY\HydroSound_MG246_3.mx

5.2.3.1 Effects of Noise on Pinnipeds in the CRC Project Area

Steller and California sea lions and harbor seals are likely to be exposed to elevated noise levels in the project area. Exposure is likely to occur from November through May when primarily adult and subadult male Steller and California sea lions typically forage at Bonneville Dam. Sea lions are known to migrate through the project area between the dam and the ocean during this time period, often making multiple round-trip journeys. Individual sea lions also are occasionally present from September to November (Tackley et al. 2008). Harbor seals are also observed at the dam between February and May, and may be present in the river during the fall and winter. Therefore exposure during this time is possible, but less likely.

It is not certain how many sea lions would be exposed to elevated noise levels. Since counts at the dam began in 2002, numbers of Steller sea lions have ranged from 3 to 75 individuals; numbers of California sea lions have ranged from 30 to 104 individuals; numbers of harbor seals have ranged from 1 to 3 individuals (Stansell et al. 2010). Based on trends in the number of pinnipeds identified at Bonneville Dam in recent years, we estimate up to 89 California sea lions, 225 Steller sea lions, and six harbor seals may travel through the project area, annually, in future years. These figures account for animals making round-trips between Astoria and Bonneville Dam more than once in a given year.

There are no pinniped haulouts or breeding sites in areas likely to be exposed to elevated noise. The nearest known sea lion haulout is located approximately 32 miles upstream of the project area (Tennis 2009 pers. comm.). The nearest sea lion breeding site is located more than 200 miles from the project area (NMFS 2008b). The nearest known harbor seal haulout is located at Carroll Slough at the confluence of the Cowlitz and Columbia Rivers, approximately 45 miles west of the westernmost edge of the project area. Therefore, elevated noise levels would have no effect on individuals at breeding or haulout sites.

Sea lions use the project area primarily for transiting only and are expected to be highly mobile when present in portions of the project area exposed to noise above the threshold levels for injury and disturbance. Additionally, Lambourn (2010 pers. comm.) notes that sea lions are likely to avoid unfamiliar noises unless there is a particular attraction keeping them in the area. As the CRC project area does not contain any such attractions (e.g., an especially rich food source, breeding area, or haulout site), sea lions would presumably avoid portions of the project area exposed to high levels of elevated noise (for example, noise generated by impact pile driving). Therefore, they would likely experience only brief, temporary behavioral disturbance or harassment as a result of impact pile-driving noise. Lambourn (2010) also added that Steller sea lions could become habituated to noises that are continuous and occurring over longer periods of time (such as vibratory pile-driving noise).

Exposure to Underwater Impact Pile-driving Noise

Exhibit 5-46 and Exhibit 5-47 below quantify the extent, timing, and duration of impact pile-driving noise that would exceed threshold levels for disturbance and injury to sea lions. Impact pile driving is expected to take place over the 4-year in-water construction period. During each year, work would likely occur within a 31-week in-water work window, ranging from week 38 of one year to week 16 of the next (or approximately from September 15 to April 15). There would be a total of about 138 days of impact pile driving in the Columbia River and 134 days of impact pile driving in North Portland Harbor over the 4-year construction period (Exhibit 5-23). Impact pile driving would be restricted to approximately 40 minutes per 12-hour work day. During most of this 40-minute period, pile driving would occur only with the use of a noise

attenuation device; however, for a short duration (about 7.5 minutes per week in the Columbia River and roughly 2.5 to 5 minutes per week in North Portland Harbor), unattenuated pile driving may occur either during routine testing of the attenuation device or accidentally in the case of equipment failure. Each work day would include a period of at least 12 consecutive hours with no impact pile driving in order to minimize disturbance to aquatic animals. Likewise, each 7-day work week would include a 2-day pile-driving recess. Impact pile driving would occur only during daylight hours.

Exhibit 5-46. Summary of Extent, Timing, and Duration of Impact Pile-Driving Noise above 190 dB RMS Underwater Injury Threshold^a

	Columbia River			North Portland Harbor			
Pile Size and Number	Distance (m)	Duration	No. Days	Distance (m)	Duration	No. Days	
Without Attenuation Device							
18- to 24-inch pile	9	7.5 min/week	38	9	2.5–5 min/week	18	
36- to 48-inch pile	54	7.5 min/week	38	54	2.5–5 min/week	31	
With Attenuation Device							
18- to 24-inch pile	2	40 min/day	138	2	40 min/day	72	
36- to 48-inch pile	12	40 min/day	138	12	40 min/day	62	

Note: Elevated noise levels would occur throughout the 4-year in-water construction period. Potential exposure may only occur from approximately September to May, when Steller sea lions are typically present in the project area.

Exhibit 5-47. Summary of Extent, Timing, and Duration of Impact Pile-Driving Noise above 160 dB RMS Underwater Disturbance Threshold

	Co	lumbia River		North Portland Harbor			
Pile Size and Number	Distance (m)	Duration	No. Days	Distance (m)	Duration	No. Days	
Without Attenuation Device							
18- to 24-inch pile	858	7.5 min/week	38	858	2.5–5 min/week	18	
36- to 48-inch pile	5412	7.5 min/week	38	3058 – u 5412 – d	2.5–5 min/week	31	
With Attenuation Device							
18- to 24-inch pile	185	40 min/day	138	185	40 min/day	72	
36- to 48-inch pile	1166	40 min/day	138	1166	40 min/day	62	

Note: Elevated noise levels would occur throughout the 4-year in-water construction period. Potential exposure may only occur from approximately September to May, when Steller sea lions are typically present in the project area.

Exposure to Underwater Vibratory Pile-driving Noise

Pipe Pile and Sheet Pile

Exhibit 5-48 summarizes the extent, timing, and duration of noise above the 120 dB RMS disturbance threshold generated by vibratory pile driving during installation of pipe pile and sheet

a Sea lions would actually not be exposed to injurious levels of noise, because impact pile driving would stop when sea lions are present in the injury zone.

u = upstream, d = downstream

pile. Vibratory driving of pipe pile and sheet pile is not expected to exceed the 190 dB RMS injury threshold, but it is likely to exceed the 120 dB RMS disturbance threshold.

Vibratory driving of pipe pile is likely to occur intermittently throughout the entire in-water project area at all new in-water piers or bents. Vibratory driving of sheet pile would occur along the same timeline, but only at pier complexes 2 and 7 in the Columbia River. These activities would occur continually throughout the 4-year in-water construction period over approximately 49 to 54 months. These activities are not restricted to an in-water work window, and therefore may take place during any of the 52 weeks of the year.

Vibratory driving of pipe pile and sheet pile is also likely to occur during demolition of the existing Columbia River bridge piers to install barge moorings and cofferdams. Pipe piles for barge moorings would be installed and removed continually throughout the entire 18-month demolition period, during any of the 52 weeks of the year. Cofferdams would each require about 10 days to install and would likely be installed during the last 13 months of the 18-month demolition period. Exhibit 5-48 shows the estimated extent, timing, and duration of noise above the 120 dB RMS disturbance threshold.

Exhibit 5-48. Summary of Exposure to Vibratory Pile-Driving Noise Above 120 dB RMS Disturbance Threshold – Pipe Pile and Sheet Pile

		Columbia River			North Portland Harbor		
Pile Type	Timing	Distance (m)	Hours /Day	No. Days	Distance (m)	Hours /Day	No. Days
Pipe Pile	Year-round	20166 – u 8851 – d	Up to 5	1470 - 1620	3058 – u 5632 – d	Up to 5	1470 – 1620
Sheet Pile	Year-round	6962 to 15000 – u 6962 to 8851 – d	Up to 24	99	N/A	N/A	N/A

Note: Elevated noise levels would occur throughout the 4-year in-water construction period. Potential exposure may only occur from approximately September to May when sea lions are typically present in the project area.

Steel Casings

Exhibit 5-49 summarizes the extent, timing, and duration of noise above the injury and disturbance thresholds during vibratory installation of steel casings. The design team estimates that vibratory installation of 10-foot casings would take approximately 90 days in the Columbia River and 31 days in North Portland Harbor. Vibratory installation of 10-foot casings is not restricted to the in-water work window and therefore may take place any time during the 4-year in-water construction period.

Exhibit 5-49. Summary of Exposure to Vibratory Pile-Driving Noise above Disturbance and Injury Thresholds – Steel Casings

		Columbia River		North Portla	and Harbor
Threshold	Timing	Distance (m)	No. Days	Distance (m)	No. Days
120 dB RMS	Year-round	20,166 – u 8,851 - d	90	3,058 – u 5,632 - d	31
190 dB RMS	Year-round	5	90	5	31

Note: Elevated noise levels would occur throughout the 4-year in-water construction period. Potential exposure may only occur from approximately October to May when Steller sea lions are typically present in the project area.

u = upstream, d = downstream

u = upstream, d = downstream

As stated earlier, hydroacoustic monitoring would be conducted to field verify the distances within which noise exceeds these thresholds and to satisfy the conditions of the Incidental Harassment Authorization.

Exposure to Airborne Impact Pile-Driving Noise

Exhibit 5-50 summarizes the extent, timing, and duration of airborne impact pile-driving noise. Airborne noise effects would occur on the same schedule as those described for underwater impact pile driving.

Exhibit 5-50. Summary of Exposure to Airborne Impact Pile-Driving Noise above 100 dB RMS Disturbance Threshold Generated by Impact Pile Driving

Location	Distance from Source (m)	Mins/Day	No. Days
Columbia River	195	≤40	138
North Portland Harbor	195	≤40	134

Note: Elevated noise levels would occur throughout the 4-year in-water construction period. Potential exposure may only occur from approximately September to May, when Steller sea lions are typically present in the project area.

<u>Injury</u>

The CRC project is not likely to injure pinnipeds. Although impact pile-driving noise is likely to exceed the injury threshold, this effect will be limited to a distance of 2 to 54 meters from the noise source, depending on the number and size of the piles. Vibratory installation of steel casings may also exceed the injury threshold within 5 meters of the source. Additionally, as impact pile-driving noise will be sporadic, occurring only about 45 minutes per day, pinnipeds will likely avoid it as an unfamiliar source of disturbance. Similarly, installation of the steel casings will be very limited. Pinnipeds are expected to avoid the injury zone rather than becoming habituated, thus reducing the potential for exposure.

The CRC project will further limit the potential for injury to pinnipeds through the implementation of a monitoring plan. Marine mammal monitors will ensure that the CRC project curtails impact pile driving if seals and sea lions approach the 2- to 54-meter injury isopleth, known as the safety zone, for impact pile driving. For added protection, the safety zone will be a minimum of 50 meters even though the injury isopleth may actually be smaller. Additionally, if vibratory installation of 10-foot diameter steel casings produces noise above the injury threshold, this activity will cease before seals or sea lions enter the potential safety zone for vibratory pile driving. The project will perform hydroacoustic monitoring to confirm injury zone isopleths. Monitoring zones may be refined accordingly, but will never be less than 50 meters.

Measures to avoid injury to seals and sea lions, including details of the monitoring plan and shut-down procedures, are described in Section 6. Because injurious noise levels will extend only a short distance and marine mammals will be monitored approaching these areas, it is reasonable to expect that qualified marine mammal monitors will be able to detect seals and sea lions within these areas. Impact pile driving will not occur at night, making the probability of detection very high. Vibratory installation of 10-foot-diameter steel casings may occur at night; however, marine mammal monitors will use night-vision/night-detection equipment to ensure detection of seals or sea lions within the safety zone while this activity is taking place. For these reasons, avoidance of injury through implementation of a monitoring plan would be attainable. While injury is theoretically possible, it is not probable. Therefore, CRC project-generated noise is not likely to injure seals and sea lions.

Behavioral Effects

The project is likely to create noise above threshold levels for airborne and underwater behavioral disturbance to pinnipeds. Exhibits 46 through Exhibit 49 outline the extent, timing, and duration of this effect.

Studies on behavioral effects to seals and sea lions are limited (Southall et al. 2007), and because the few available studies show wide variation in response to underwater and airborne noise, it is difficult to quantify exactly how pile-driving noise will affect pinnipeds. The literature shows that elevated underwater noise levels could prompt a range of effects, including no obvious visible response, brief visual orientation towards the noise, curiosity (or movement towards the source), or habituation to the sound (Southall et al. 2007). For underwater noise, Southall et al. (2007) note that there is little evidence that high levels of pulsed noise, ranging between 150 and 180 dB RMS relative to 1 μ Pa, will prompt avoidance of an area. For airborne noise, Southall et al. (2007) note there is extremely limited data suggesting very minor, if any, observable behavioral responses by pinnipeds exposed to airborne pulses of 60 to 80 dB relative to 20 μ Pa; however, given the paucity of data on the subject, we cannot rule out the probability that avoidance of noise in the project area could occur.

In an effort to gauge potential sea lion response to disturbance, CRC reviewed reports of sea lion response to harassment from hazing techniques just below Bonneville Dam. The deterrence efforts below Bonneville Dam began in 2005 and have used acoustic deterrent devices (ADDs), boat chasing, above-water pyrotechnics (cracker shells, screamer shells or rockets), rubber bullets, rubber buckshot, and beanbags (Stansell et al. 2009). Review of deterrence activities by the West Coast Pinniped Program noted "USACE observations from 2002 to 2008 indicated that increasing numbers of California sea lions were foraging on salmon at Bonneville Dam each year, salmon predation rates increased, and the deterrence efforts were having little effect on preventing predation" (Scordino 2010). In the USACE status report through May 28, 2010, boat hazing was reported to have limited, local, short-term impact in reducing predation in the tailrace, primarily from Steller sea lions. ODFW and WDFW reported that sea lion presence did not appear to be significantly influenced by boat-based activities and several "new" sea lions (initially unbranded or unknown from natural markings) continued to forage in the observation area in spite of shore- and boat-based hazing. They suggested that hazing was not effective at deterring naïve sea lions if there were large numbers of experienced sea lions foraging in the area (Brown et al. 2010). Observations on the effect of ADDs, which were installed at main fishway entrances by mid-June of 2007, noted that pinnipeds were observed swimming and eating fish within 20 feet of some of the devices with no deterrent effect observed (Tackley et al. 2008a; Tackley et al. 2008b; Stansell et al. 2009; Stansell et al. 2010). Many of the animals returned to the area below the dam despite hazing efforts (Stansell et al. 2009; Stansell and Gibbons 2010). Relocation efforts to Astoria and the Oregon coast were tried in 2007; all but 1 of 14 relocated animals returned to Bonneville Dam within days (Scordino 2010).

No information on in-water noise levels of hazing activities at Bonneville Dam has been published other than that ADDs produce underwater noise levels of 205 dB in the 15 kHz range (Stansell et al. 2009). Durations of boat-based hazing events were reported at less than 30 minutes for most of the 521 boat-based events in 2009, but ranged up to 90 minutes (Brown et al. 2009). Durations of boat-based hazing events were not reported for 2010. However, 280 events occurred over 44 days during a 5-month period using a total of 4,921 cracker shells, 777 seal bombs, and 97 rubber buckshot rounds (Brown et al. 2010). Based on knowledge of in-water noise from construction activities, the CRC project assumes that pinniped exposure to project-related in-water noise (e.g., in-water construction and demolition) will be less than that of exposure to hazing techniques.

In addition, sea lions are expected to traverse through and not remain in the project area. Tagging studies of California sea lions indicate they pass hydrophones upriver and downriver of the CRC project site quickly. Wright et al. (2010a) reported minimum upstream and downstream transit times between the Astoria haulout and Bonneville Dam (river distance ~20 kilometers) were 1.9 and 1 day, respectively, based on 14 trips by 11 sea lions. The transit speed was calculated to be 4.6 km/hr in the upstream direction and 8.8 km/hr in the downstream direction. Graphics of the six individuals acoustically tagged in 2009 show they made a combined total of 11 upriver or downriver trips quickly through the CRC project site to or from Bonneville Dam and Astoria (Brown et al. 2009). Graphics from four acoustically tagged California sea lions in 2010 also indicate that the animals moved though the area below Bonneville Dam down to the receivers located below the CRC project site rapidly both in the upriver or downriver directions (Wright draft report graphics from pers. comm. 2010b). Although the data apply to California sea lions, Steller sea lions and harbor seals have no incentive to stay near the CRC project area, have no haulouts near the project area, and are expected to also pass the project area quickly. Therefore, sea lions and seals are not expected to be exposed to a long duration of construction noise.

Underwater noise generated by impact pile driving may cause minor disruption of movement through the area and of feeding activities, but based on travel data and haulout locations, exposure is expected to be brief. Additionally, because many of the individuals transiting the area are already habituated to high ambient disturbance levels from existing commercial and recreational vessel traffic and to hazing at Bonneville Dam, it is expected that they would not be sensitive to pile-driving noise. Although brief, temporary, behavioral harassment would occur within the disturbance threshold areas, elevated noise levels from impact pile driving are expected to have only a negligible effect on foraging and transiting of individual seals and sea lions, and no effect on the overall populations.

Safe passage concerns during pile installation and removal at more than one pier complex were raised by NMFS. Given the 800-meter width of the Columbia River and the rarity of impact pile driving on opposite sides of the river (approximately 1 or 2 days total throughout the approximately 4-year construction period), passage should not be hindered. Vibratory installation or removal of piles at more than one pier complex would likely occur at the same time on occasion during construction and demolition. During construction and demolition, space limitations due to barge size and limitations on the amount of equipment available are anticipated to be limiting factors for the contractor. Vibratory installation of steel casings, pipe piles, and sheet piles are calculated to exceed behavioral disturbance thresholds at between 6,962 and over 50,000 meters. In this case, the entire width of the channel would be affected by noise above the disturbance threshold even if only one pier complex was being worked on. As stated above, not enough information on Columbia River pinniped reaction to vibratory driving is available to determine whether individuals would alter their movement patterns in this industrial area and some seals or sea lions may not pass the construction site due to noise and general construction activity. However, the safety of the animals would not be compromised by these noise levels.

If in-water work for pile installation or removal occurs simultaneously on both sides of the river and results in changes to pinniped behavior so they do not pass the project site, it would be due to behavioral harassment from incidental exposure to a short duration of non-injurious noise levels and general construction activity. Sea lions would be expected to traverse through the area and not be exposed to a long duration of noise. Also, based on observations from the Bonneville Dam sea lion removal program over 6 years, many individual sea lions appear motivated to reach the Bonneville Dam tailrace to forage, undeterred by directed noise hazing techniques. Many individual sea lions return repeatedly, even after being exposed to hazing and being captured, herded, and branded. Therefore, it is anticipated that the less intense, short duration, non-injurious

noise levels from installation and removal of pipe piles, casings, and cofferdams will not affect the behavior of many of the individual sea lions in the project area.

Harbor seals occur sporadically in low numbers in the project area, so they are less likely to be exposed to the short periods when impact or vibratory installation and removal noise could occur on both sides of the Columbia River simultaneously.

Temporary Threshold Shift

Temporary Threshold Shift (TTS) is reversible hearing loss caused by fatigue of hair cells and supporting structures in the inner ear. Technically, TTS is not considered injury, as it consists of fatigue to auditory structures rather than damage to them. Impact pile driving would produce maximum underwater source pulsed noise levels estimated at 210 dB peak and 176 dB SEL with 10 dB of attenuation from an attenuation device (214 dB peak and 186 dB SEL without an attenuation device). Summarizing existing data, Southall et al. (2007) assumed that pulses of underwater noise result in the onset of TTS in pinnipeds when levels reach 212 dB peak or 171 dB SEL. They did not offer criteria for non-pulsed sounds. Although these suggested criteria have not been adopted by any regulatory body, they are presented as a starting point to discuss the likelihood of TTS occurring during the CRC project. The literature has not drawn conclusions on levels of underwater non-pulsed noise (e.g., vibratory pile installation) likely to cause TTS. With a noise attenuation device, TTS is not likely to occur based on our estimated source levels. Without a noise attenuation device, we estimate that the extent of the area in which underwater noise levels could potentially cause TTS is somewhere in between the extent of where the injury threshold occurs (2 to 54 meters from the noise source) and the extent of where the disturbance threshold occurs (74 FR 63724).

Although underwater noise levels produced by the CRC project may exceed levels produced in studies that have induced TTS in pinnipeds (Southall et al. 2007), there is a general lack of controlled, quantifiable field studies related to this phenomenon. Existing studies have had varied results (Southall et al. 2007). Therefore, it is difficult to extrapolate from these data to site-specific conditions on the CRC project. For example, because most of the studies have been conducted in laboratories, rather than in field settings, the data are not conclusive as to whether noise will cause seals and sea lions to avoid the project area, thereby reducing the likelihood of TTS, or whether noise will attract seals and sea lions, increasing the likelihood of TTS. In any case, there are no universally accepted standards for the amount of exposure time likely to induce TTS. Lambourne (2010 personal communication) posits that, in most circumstances, free-roaming Steller sea lions are not likely to remain in areas subjected to high noise levels long enough to experience TTS unless there is a particularly strong attraction, such as an abundant food source. While we may infer that TTS could conceivably result from the CRC project, it is impossible to exactly quantify the magnitude of exposure, the duration of the effect, or the number of individuals likely to be affected.

Impact pile driving would produce initial airborne noise levels of approximately 112 dB peak at 160 feet from the source, as compared to the level suggested by Southall et al. (2007) of 143 dB peak referenced to 20 μ Pa for onset of TTS in pinnipeds from multiple pulses of airborne noise. It is not expected that airborne noise levels would prompt TTS in individual seals and sea lions. Exposure is likely to be brief because seals and sea lions use the project area for transiting, rather than breeding or hauling out. In summary, it is expected that elevated noise would have only a negligible probability of causing TTS in individual seals and sea lions.

5.2.3.2 Conclusion

Injury to California and Steller sea lions is avoidable through the implementation of a monitoring plan that requires a cessation of impact pile driving before individuals enter the underwater injury zone, defined as from 2 to 54 meters from impact pile driving. Additionally, if vibratory installation of 10-foot-diameter steel casings produces noise above the injury threshold, this activity would cease before sea lions enter the potential injury zone (anticipated to be 5 meters from the activity).

Noise above the behavioral disturbance threshold is probably unavoidable during both impact and vibratory pile driving, but effects to sea lions are expected to be brief and temporary, impacting only a small number of adult and subadult sea lions transiting the project area. No noise disturbance would occur at breeding areas or haulouts. Noise is not expected to significantly interfere with foraging, transiting, breathing, or other essential life functions.

5.2.4 Noise from Underwater Debris Removal

Debris removal may occur in North Portland harbor at the location of each of the new piers where there is anecdotal evidence that riprap occurs within the pier footprints. Debris removal in the North Portland Harbor, if it occurs, would be likely to create noise at or above the 120 dB RMS disturbance threshold for continuous noise in underwater portions of the project area.

Few studies have been conducted on noise emissions produced by underwater debris removal. A review of the literature indicates that underwater debris removal will produce noise in the range of 135 dB to 147 dB RMS at 10 meters (Dickerson et al. 2001; OSPAR 2009; Thomsen et al. 2009), i.e., greater than the 120 dB RMS disturbance threshold for non-pulsed noise.

Underwater debris removal is not expected to generate significant airborne noise. The air-water interface creates a substantial sound barrier and reduces the intensity of underwater sound waves by a factor of more than 1,000 when they cross the water surface. The abovewater environment is, thus, virtually insulated from the effects of underwater noise (Hildebrand 2005). Therefore, underwater debris removal is not expected to measurably increase ambient airborne noise.

Exhibit 5-51 shows the calculated distance at which underwater debris-removal noise attenuates to the underwater disturbance threshold for continuous noise (120 dB RMS).

Exhibit 5-51. Underwater Noise Attenuation for Debris-Removal Noise – Calculated Values

	Distance from Source (m)
Noise Level (dB RMS)	Bucket Dredge Source Sound Pressure Level 147 at 10 m
150	7
140	30
130	136
120	631

5.2.4.1 Potential Exposure of Steller and California Sea Lions to Underwater Debris-Removal Noise

Exhibit 5-52 summarizes potential exposure of pinnipeds to underwater debris-removal noise in North Portland Harbor. Exposure is presented as an overlap of the areal extent of noise at or

above the disturbance threshold, combined with the duration and timing of the impact and the time periods when seals and sea lions are likely to be present in the project area.

Debris removal is not certain to occur, but is included to present the fullest disclosure of effects. It is possible that debris removal would occur in North Portland harbor at the location of each of the new piers where there is anecdotal evidence that riprap occurs within the pier footprints. The exact quantity of this material is unknown, but as a worst-case scenario, this activity would remove approximately 90 cubic yards of material over an area of approximately 2,433 square feet from all piers combined.

Exhibit 5-52. Summary of Debris-Removal Noise Above 120 dB RMS Underwater Disturbance Threshold

Noise Source	Location ^a	Underwater Distance (m)	Hours/ Day	No. Days	Timing⁵
Bucket dredge	Potentially at all new NPH piers	631	≤12	7 days	Nov 1–Feb 28

a NPH = North Portland Harbor

5.2.4.2 Effects of Exposure to Debris-Removal Noise

The reactions of pinnipeds to debris-removal noise have received virtually no study. Previous studies indicate that dredging noise has resulted in avoidance reactions in marine mammals; however, the number of studies is few, limited to only a handful of locations. Thomsen et al. (2009) caution that given the limited number of studies, the existing published data may not be representative, and that it is therefore impossible to extrapolate the potential effects from one area to the next.

In a review of the available literature regarding the effects of dredging noise on marine mammals, Richardson et al. (1995) found studies only related to whales and porpoises and none related to pinnipeds. The review did, however, find studies related to the response of pinnipeds to "other construction activities," which may be applicable to dredging noise. Three studies of ringed seals during construction of artificial islands in Alaska showed mostly mild reactions ranging from negligible to temporary local displacement. Green and Johnson (1983, as cited in Richardson et al. 1995) observed that some ringed seals moved away from the disturbance source within a few kilometers of construction. Frost and Lowry (1988, as cited in Richardson et al. 1995) and Frost et al. (1988, as cited in Richardson et al. 1995) noted that ringed seal density within 3.7 kilometers of construction was less than seal density in areas located more than 3.7 kilometers away. Harbor seals in Kachemak Bay, Alaska, continued to haul out despite construction of hydroelectric facilities located 1,600 meters away. Finally, Gentry and Gilman (1990) reported that the strongest reaction to quarrying operations on St. George Island in the Bering Sea was an alert posture when heavy equipment occurred within 100 meters of northern fur seals.

In their study about sea lion hazing at Bonneville Dam, Stansell et al. (2009) note that sea lions showed only temporary behavioral responses to underwater loud noises, such as ADDs and seal bombs, and above-water pyrotechnics, which did not cause any measurable interference with foraging or transiting. Sea lions quickly habituated to the noise, some foraging within 20 feet of intense noise. The results suggest that some of individuals that transit through the project area either are already habituated to some loud noises or could readily become habituated.

b Over the course of in-water construction period in the North Portland Harbor: 2013 to 2017.

5.2.1 Effect of Exposure to Debris-Removal Noise

There are no established levels of underwater debris-removal noise shown to cause injury to seals and sea lions. However, since the maximum expected debris-removal noise levels on the CRC project are below any known injury thresholds (190 dB RMS, for impulsive noises), it is unlikely that this activity would produce noise levels that are injurious to seals and sea lions. Additionally, the limited body of literature does not include a single report of injuries caused by noise from underwater excavation.

Debris-removal noise is likely to exceed the disturbance threshold (120 dB RMS for non-pulsed continuous noises) for only a short distance from the source (approximately 631 meters). Specific responses to noise above this level may range from no response to avoidance to minor disruption of migration and/or feeding. Alternatively, seals and sea lions may become habituated to elevated noise levels (NMFS 2005; Stansell 2009). This is consistent with the literature, which reports only the following behavioral responses to these types of noise sources: no reaction, alertness, avoidance, and habituation. NMFS (2005) posits that continuous noise levels of 120 dB RMS may elicit responses such as avoidance, diving, or changing foraging locations.

Debris removal is only estimated to occur for up to 7 days over the 4-year construction period in North Portland Harbor. If this activity overlaps with pinniped presence, behavioral disturbance is expected to be brief and temporary, restricted to individuals that are transiting the North Portland Harbor portion of the project area. Because many of the individual sea lions transiting the project area are already habituated to hazing at Bonneville Dam and to high levels of existing noise throughout the lower Columbia River, we expect that they would not be especially sensitive to a marginal increase in existing noise. Thus, due to the short duration of this noise, its location only in North Portland Harbor and the high level of existing disturbance throughout the lower Columbia River, noise generated from debris removal is not expected to result in behavioral disturbance that would rise to the level of harassment.

5.2.2 Vessel Noise

Various types of vessels, including barges, tug boats, and small craft, will be present in the project area at various times. Vessel traffic will continually traverse the in-water CRC project area, with activities centered on piers 2 through 7 of the Columbia River and the new North Portland Harbor bents. Such vessels already use the project area in moderately high numbers; therefore, the vessels to be used in the project area do not represent a new noise source, only a potential increase in the frequency and duration of these noise types.

There are very few controlled tests or repeatable observations related to the reactions of pinnipeds to vessel noise. However, Richardson et al. (1995) reviewed the literature on reactions of pinnipeds to vessels, concluding overall that seals and sea lions showed high tolerance to vessel noise. One study showed that, in water, sea lions tolerated frequent approach of vessels at close range, sometimes even congregating around fishing vessels. Because the project area is heavily traveled by commercial and recreational craft, it seems likely that seals and sea lions that transit the project area are already habituated to vessel noise, thus the additional vessels that will occur as a result of CRC project activities will likely not have an effect on these pinnipeds. Therefore, CRC project vessel noise in the project area would be unlikely to rise to the level of harassment.

5.2.3 Physical Disturbance

Vessels, in-water structures, and overwater structures have the potential to cause physical disturbance to seals and sea lions, although in-water and overwater structures will cover no more

than 20 percent of the entire channel width at one time (CRC 2010). As previously mentioned, various types of vessels already use the project area in high numbers. Tug boats and barges are slow moving and follow a predictable course. Seals and sea lions will be able to easily avoid these vessels while transiting through the project area, and they are probably already habituated to the presence of numerous vessels, as the lower Columbia River and North Portland Harbor receive high levels of commercial and recreational vessel traffic. Therefore, vessel strikes are extremely unlikely and, thus, discountable. Potential encounters will likely be limited to brief, sporadic behavioral disturbance, if any at all. Such disturbances are not likely to result in a risk of harassment of seals and sea lions transiting the project area.

5.2.4 Effects on Prey

Fish are the primary dietary component of all of the pinniped species in the project area. The Columbia River and North Portland Harbor provide migration and foraging habitat for sturgeon and lamprey, migration and spawning habitat for eulachon, and migration habitat for juvenile and adult salmon and steelhead, as well as some limited rearing habitat for juvenile salmon and steelhead.

There are no physical barriers to fish passage within the project area, nor are there fish passage barriers between the project area and the Pacific Ocean. The proposed project would not involve the creation of permanent physical barriers and, thus, long-term changes in seal and sea lion prey species distribution would not be expected to occur.

Any adverse effects to prey species would occur during project construction and are temporary. All project activities would be conducted using the BMPs and minimization measures outlined in Section 6. Given the large numbers of fish in the Columbia River, the short-term nature of effects to fish populations, and extensive BMPs and minimization measures to protect fish during construction, as well as conservation and habitat mitigation measures that will continue into the future, the project would not be expected to have measurable effects on the distribution or abundance of potential prey species in the long term. Therefore, temporary habitat impacts would be expected to have a negligible impact to habitat for pinniped prey species. These effects to prey species are summarized below and are outlined in more detail in Sections 6.1 to 6.3 of the CRC project BA.

Noise from pile installation could harm (impact driving) or cause behavioral disturbance to fish (impact or vibratory installation of steel pipe pile, vibratory installation of steel pipe pile, sheet pile, and steel casings for drilled shaft placement). Avoidance and minimization measures would be implemented to limit effects to fish due to noise from impact pile driving. These measures include the use of drilled shafts for bridge foundation rather than impact driving of 8-footdiameter steel pipe piles, minimizing of the number of in-water piers or bents, restricting impact pile driving to the amount needed for proofing of load-bearing piles only, timing windows for inwater impact pile driving to avoid the majority of fish runs, use of noise attenuation devices, vibratory installation of piles to the extent practicable, a hydroacoustic performance measure to monitor and limit the extent of potential incidental fish take, and on-site biological monitors. Nevertheless, impact pile driving would likely create a temporary migration barrier to all life stages of fish using the Columbia River and North Portland Harbor, although this would be localized. Cofferdams and temporary in-water work structures also could create partial barriers to the migration of juvenile fish in shallow-water habitat. Impacts to fish species distribution would be temporary during in-water work and hydroacoustic impacts from impact pile driving would only occur for limited periods during the day and only during the in-water work window established for this activity in conjunction with ODFW, WDFW, and NMFS. The overall effect to the prey base for seals and sea lions would be insignificant.

Prey may be affected by turbidity, contaminated sediments, or other contaminants in the water column. The CRC project would minimize, avoid, or contain all potential sources of contamination, minimizing the risk of exposure to prey species of seals and sea lions. The CRC project involves several activities that could potentially generate turbidity in the Columbia River and North Portland Harbor, including pile installation, pile removal, installation and removal of cofferdams, installation of steel casings for drilled shafts, and debris removal. Turbidity would not be expected to cause mortality in the fish species using the project area, and effects would probably be limited to temporary avoidance of the discrete areas of elevated turbidity (no more than 300 feet from the source) for approximately 4 to 6 hours at a time (CRC 2010). Therefore, turbidity will have only insignificant effects to fish and, thus, insignificant effects on seals and sea lions.

In-water work would be extremely unlikely to mobilize contaminated sediments (CRC 2010). Well in advance of in-water work, the CRC project team would perform an extensive search for evidence of contamination, pinpointing the location, extent, and concentration of the contaminants. Then, BMPs would be implemented to ensure that the CRC project (1) avoids areas of contaminated sediment or (2) enables responsible parties to initiate clean-up activities for contaminated sediments occurring from construction activities within the project area. These BMPs would be developed and implemented in coordination with regulatory agencies. Because the CRC project would identify the locations of contaminated sediments and use BMPs to ensure that they do not become mobilized, there is little risk that the prey base of seals and sea lions would be greatly affected by or exposed to contaminated sediments.

In-water and near-water construction would employ numerous BMPs and would comply with numerous regulatory permits to ensure that contaminants do not enter surface water bodies. In the unlikely event of accidental release, numerous BMPs and a Pollution Control and Contamination Plan (PCCP) would be implemented to ensure that contaminants are prevented from spreading and are cleaned up quickly. Therefore, contaminants would not be likely to significantly affect fish and, thus, effects on the seal and sea lion prey base would also be insignificant.

5.3 Physical Loss of Prey Species Habitat

The project would lead to temporary physical loss of approximately 20,700 square feet of shallow-water habitat. Project elements responsible for temporary physical loss include the footprint of the numerous temporary piles associated with in-water work platforms, work bridges, tower cranes, oscillator support piles, cofferdams, and barge moorings in the Columbia River and North Portland Harbor.

The in-water portions of the new structures would result in the permanent physical loss of approximately 250 square feet of shallow-water habitat at pier complex 7 in the Columbia River. Demolition of the existing Columbia River structures would permanently restore about 6,000 square feet of shallow-water habitat. Overall, there would be a net permanent gain of at least 5,945 square feet of shallow-water habitat in the Columbia River (CRC 2010). At North Portland Harbor, there would be a permanent net loss of about 2,435 square feet of shallow-water habitat at all of the new in-water bridge bents. Note that all North Portland Harbor impacts would be in shallow water.

Physical loss of shallow-water habitat is of particular concern for rearing or subyearling migrant salmonids. In general, in-water structures that completely block the nearshore could force these juveniles to swim into deeper-water habitats to circumvent the structures. Deep-water areas represent lower quality habitat because predation rates are higher there. Numerous studies show that predators such as walleye and northern pikeminnow occur in deep-water habitat for at least

part of the year (Johnson 1969; Ager 1976; Paragamian 1989; Wahl 1995; Pribyl et al. 2004). In the case of the CRC project, in-water portions of the structures would not pose a complete blockage to nearshore movement anywhere in the project area. Although these structures would cover potential rearing and nearshore migration areas, the habitat is not rare and is not of particularly high quality. These juveniles would still be able to use the abundant shallow-water habitat available for miles in either direction. Neither the permanent nor the temporary structures would force these juveniles into deeper water, and therefore pose no added risk of predation.

Physical loss of shallow-water habitat would have only negligible effects on foraging, migration, and holding of salmonids that are of the yearling age class or older. These life functions are not dependent on shallow-water habitat for these age classes. Furthermore, the lost habitat is not of particularly high quality. There is abundant similar habitat immediately adjacent along the shorelines of the Columbia River and throughout North Portland Harbor. The lost habitat represents only a small fraction of the remaining habitat available for miles in either direction. There would still be many acres of habitat for yearling or older age-classes of salmonid foraging, migrating, and holding in the project area. Physical loss of shallow-water habitat would have only negligible effects on eulachon and green sturgeon for the same reason as above. The effects to these elements of seal and sea lion habitat would, thus, be minimal.

The CRC project would cause a temporary physical loss of approximately 16,635 square feet of deep-water habitat, consisting chiefly of coarse sand with a small proportion of gravel. CRC project elements responsible for temporary physical loss include the cofferdams and numerous temporary piles associated with in-water work platforms and moorings. The in-water portions of the new structures would result in the permanent physical loss of approximately 6,300 square feet of deep-water habitat at pier complexes 2 through 7 in the Columbia River. Demolition of the existing Columbia River piers would permanently restore about 21,000 square feet of deep-water habitat. Overall, there would be a net permanent gain of about 15,000 square feet of deep-water habitat in the Columbia River.

Although there would be a temporary net physical loss of deep-water habitat, this would not be expected to have a significant impact on listed fish. The lost habitat is not rare or of particularly high quality, and there is abundant similar habitat in immediately adjacent areas of the Columbia River and for many miles both upstream and downstream. The lost habitat would represent a very small fraction (far less than 1 percent) of the remaining habitat available. Additionally, the in-water portions of the permanent and temporary in-water structures would occupy no more than about 1 percent of the width of the Columbia River. Therefore, the structures would not pose a physical barrier to fish migration.

In addition, compensatory mitigation for direct permanent habitat loss to jurisdictional waters from permanent pier placement would occur in accordance with requirements set by USACE, DSL, Ecology, ODFW, and WDFW. To meet these requirements, CRC would propose restoring habitat in the lower Lewis River and lower Hood River. At the Hood River site, 1 mile of a historic side channel would be reconnected to the lower Hood River and an existing 21-acre wetland resulting in habitat benefits to salmonids and eulachon. At the Lewis River site, restoration of 18.5 acres of side channels would occur between the lower Lewis River and the lower Columbia River resulting in habitat benefits to salmonid and other native species. Therefore, permanent habitat loss would be expected to have a negligible impact to habitat for pinniped prey species.

Due to the small size of the impact relative to the remaining habitat available, and the permanent benefits from habitat restoration, both temporary and permanent physical habitat loss would be insignificant to fish and, thus, to the habitat and foraging opportunities of seals and sea lions.

5.3.1 Work-Area Isolation and Fish Salvage

The project would use cofferdams to isolate the in-water work area from active flow during construction in the Columbia River. Cofferdams would be used during demolition of the existing bridge in the Columbia River, if a wire saw is not used to cut the existing piers into pieces. The purpose of the cofferdams is to avoid contaminating the Columbia River with work materials or wastes, to contain re-suspended sediments, and to minimize disturbance and injury to fish. Cofferdams would be installed in a manner that minimizes fish entrapment. Sheet piles would be installed from upstream to downstream and lowered slowly until in contact with the substrate.

Up to 11 cofferdams are anticipated. The two cofferdams used during construction of piers 2 and 7 in the Columbia River would cover a combined area of approximately 15,750 square feet. The nine cofferdams used during demolition of the existing in-water Columbia River bridge piers 2 through 10 would each encompass an area of 7,500 square feet (for a total area of 67,500 square feet). Cofferdams would likely be installed and removed at any time of year, pending approval from USFWS and NMFS. ODFW and WDFW have both agreed that performing this activity outside of the standard work window would not cause significant harm to fish. Installation would use low-impact methods such as vibrating or pressing into place.

Cofferdams used for construction would each require 10 days to install, be in place for approximately 330 to 470 calendar days apiece, and would require 15 days for removal.

Each cofferdam used for demolition would require 10 days to install, be in place for approximately 20 additional work days apiece, and require approximately 10 work days to remove.

Installation of the cofferdams would likely generate low-level noise and visual disturbance. For this reason, fish are likely to actively avoid the work area during the construction of cofferdams. Nevertheless, due to the large size of the cofferdams, it is impossible to guarantee that no fish would become trapped inside. To minimize impacts to fish, the project would perform measures to remove fish from the work area during and after the installation of the cofferdams. Fish salvage would be conducted by qualified biologists in compliance with protocols outlined in Section 6. Methods could include seining, electrofishing, trapping, and encouraging volitional movement of fish away from the work area. Captured fish would be released outside of the work area. To avoid entrainment of fish, pump intakes would be screened according to ODFW and WDFW standards and ODOT and WSDOT protocols outlined in Section 6.

The salvage involves capture, direct handling, and transporting of fish; therefore, there is a reasonable risk that the operation may harass, injure, or kill fish. If fish remain trapped in a cofferdam during construction, they would likely perish.

Because the fish salvage operation may take place year-round, individuals from any life stage of the fish species using the Columbia River and North Portland Harbor may be exposed to this effect. Exhibit 5-53 shows the species and life stages of fish SOI that may potentially be present within the project area during work-area isolation.

Exhibit 5-53. Species of Interest and Life Stages Expected to be Present in the Project Area During Work Isolation

			Lif	e Stage		
Species	Spawning	Incubation	Rearing	Outmigrating Juveniles	Migrating/ Holding Adults	Transiting
Chinook						
Lower Columbia River ESU			Χ	Χ	X	
Upper Columbia River Spring-Run ESU			Х	Χ	Х	
Snake River Fall-Run ESU			Х	Χ	Х	
Snake River Spring/Summer-Run ESU				Χ	Х	
Steelhead						
Lower Columbia River DPS			Х	Χ	Х	
Middle Columbia River DPS				Χ	Х	
Upper Columbia River DPS				Χ	Х	
Snake River Basin DPS				X	Х	
Sockeye						
Snake River ESU				X	X	
Coho						
Lower Columbia River ESU			X	Χ	Х	
Chum						
Columbia River ESU			Χ	Χ	X	
Bull trout						
Columbia River DPS					X	
Green sturgeon						
Southern DPS					Χ	
White sturgeon	Х	Х	Х	Х	Х	Х
Eulachon						
Southern DPS	Х	X		X	Χ	
Lamprey Species			Х	Х	Х	Х
Resident fish (e.g., sculpin, dace, threespine stickleback, sucker, shiner)	Х	Х	Х			

Adult sea lions are likely to be in the Columbia River during installation and removal of cofferdams. However, they would only use the river for transiting between the Bonneville Dam foraging area and haulout sites further downstream. Sea lions are unlikely to swim near these construction activities and would probably actively avoid them (see discussion in 5.2.2 above regarding effects of noise on sea lions). Given the project's marine mammal monitoring efforts

and sea lions' ability to avoid areas of high noise and disturbance, the project would not create a barrier to normal transit through the project area. Sea lions would not be trapped within cofferdams. Although sea lions would probably be aware of the construction, their avoidance of the area means that the disturbance would be insignificant. The likelihood of exposure to harmful levels of disturbance is discountable. This element of the project is not likely to significantly impact sea lions.

5.3.2 Increase in Overwater Coverage and Shading

The project would create several temporary sources of new overwater coverage in the Columbia River and North Portland Harbor and would increase the overall shade footprint in the project area. Temporary overwater structures include work platforms, work bridges, tower cranes, oscillator support platforms, and barges.

Studies have shown that fish communities under overwater structures differ from those in adjacent areas, in part due to the effect of shading (Southard et al. 2006). In general, shade may affect listed fish by increasing habitat for predators, causing visual disorientation, and decreasing primary productivity.

5.3.2.1 General Effects of Shading on Fish

Overwater coverage increases the amount of shade in the water column. Fish rely on visual cues when performing life functions such as foraging, schooling, avoiding predators, and migration. The literature shows that changes in light conditions can alter fish behavior (Simenstad et al. 1999; Simenstad and Nightingale 2001). Overwater structures that alter the existing light regime may limit the ability of fish to perform essential life functions (Southard et al. 2006). Shade may also affect the productivity of underwater plants: the basis of the food web for many juvenile fish (Simenstad and Nightingale 1999). Finally, shade may affect fish by providing cover for predators (Carrasquero 2001).

Predation

Shade attracts and provides cover for many species of predatory fish, including northern pikeminnow, smallmouth bass, and largemouth bass (Pribyl et al. 2004; Celedonia et al. 2008). The literature does not draw a clear, consistent relationship between an increase in predation and an increase in shade; predation rates have been shown to both increase and decrease with increasing light (Carrasquero 2001). In a review of the available literature, researchers concluded that the effect of shading on predation is "inconclusive" (WSF 2009). However, a literature review conducted by Carrasquero (2001) shows that largemouth and smallmouth bass have a strong affinity for piers and overwater structures, potentially using the cover of darkness to ambush fish. In a study in the Columbia River, Beamesderfer and Riemen (1991, as cited in Celedonia et al. 2008) noted that northern pikeminnow selected low-velocity microhabitats created by in-water structures where juvenile salmonids were congregating. In a study conducted in the lower Columbia River, Zimmerman (1999) found that smallmouth bass consumed salmonids averaging 119 millimeters in length, and pikeminnow consumed salmonids averaging 167 millimeters in length. Relatively few salmonids consumed by pikeminnow were greater than 250 millimeters in length (Zimmerman 1999). This indicates that predation risks are greater for juvenile salmonids and for other small fish such as dace, threespine stickleback, redside shiners, suckers, and sculpin.

Migration and Orientation

The literature provides empirical evidence that juvenile salmonids, and presumably other fish, may become disoriented beneath overwater structures or other shaded areas with sharp contrast between light and dark. Heiser and Finn (1970), Weitkamp (1982), and Pentec (1997) reported that fish were reluctant to enter shadow zones under docks and or other sources of intense shade. Pentec (1997), Taylor and Willey (1997), Simenstad et al. (1999), Williams et al. (2003), and Toft et al. (2004) reported observing fish movement along the shadow zone boundary without penetration into the shadow. Shreffler and Moursund (1999) found that juvenile Chinook ceased directional movement at the shadow line rather than immediately continuing under an overwater structure. Juvenile salmon consistently swam from the shadow line into the light, then immediately darted down and back into the light-dark transition area again.

Other literature suggests that a sharp light-dark interface caused by overwater structures may interfere with migration in juvenile salmonids. Response of fish to overwater structures is complex, as some fish will readily pass under structures, and others will not. Schools may either disband upon encountering an overwater structure, or they may pause and proceed as a group (Southard et al. 2006). A study conducted by Pacific Northwest National Laboratory (PNNL) (Williams et al. 2003) concluded that overwater structures are likely to be impediments to juvenile migration, depending on numerous factors such as light levels, angle of the sun, cloud cover, current velocity and direction, and tidal stage. For example, the study indicated that effects of shading were reduced during low tide when more light can dissipate beneath overwater structures. The same study also observed that juvenile chum would not cross into shade when the decrease in light level was 85 percent over a horizontal distance of approximately 5 meters. Acoustic tagging at Port Townsend revealed that juvenile Chinook and coho passed under overwater structures more quickly in the evening when the light-dark interface is indistinct (Southard et al. 2006). On the other hand, Weitkamp (1982) found that juvenile salmonids will readily swim under overwater structures. Williams et al. (2003) found that salmon fry were not inhibited by the 33-foot-wide shadow cast by an overwater structure at the Mukilteo Ferry Terminal, even though light levels under the structure were 97 percent lower than ambient levels.

Thus, although the literature is not in agreement regarding the effects of shade on orientation, there appears to be some evidence that a shadow line under overwater structures could interfere with the migration of salmonid juveniles during some daylight hours. Studies have suggested that this may prompt fish to enter deeper water, where they could presumably be exposed to predation from birds, mammals, and other fish (WSF 2009). Additionally, juveniles may congregate at the edge of the shadow line, making them more vulnerable to predation (Southard et al. 2006).

Primary Productivity

Shading may result in decreased productivity of underwater vegetation. Macrophytes, benthic algae, and phytoplankton contribute to aquatic habitat complexity and form the basis of the food web for many species of fish. Carrasquero (2001) notes that lowered light levels may reduce or eliminate macrophyte beds, algae, and other aquatic vegetation beneath overwater structures. This may, in turn limit the amount of prey available to fish (Simenstad and Nightingale 2001). Epibenthic crustaceans are of most concern because they are typically associated with nearshore plants (Simenstad et al. 1999). Loss of underwater vegetation may also reduce cover for juvenile fish, potentially increasing exposure to predation (Carrasquero 2001). Furthermore, shading underneath overwater structures may reduce primary production in phytoplankton. However, this relationship is complex and poorly understood (Carrasquero 2001). For example, there is evidence that primary productivity of phytoplankton may be greater at the edge of overwater structures than in areas outside of the structure (White 1975, as cited in Carrasquero 2001). On

the other hand, Mulvihill et al. (1980, as cited in Carrasquero 2001) report that pilings and piers beneath overwater structures may provide substrate for algal growth where bottom depths are below the photic zone or where bottom substrates are unstable. The increase in algal growth may potentially compensate for loss of phytoplankton primary productivity.

5.3.2.2 Sources of Shade on the CRC Project

The CRC project would create several temporary sources of in-water shade: barges, in-water work platforms, work bridges, tower cranes, and oscillator support platforms. There would be a net increase in temporary shade Exhibit 5-54..

Exhibit 5-54. Summary of Temporary Shade Sources in the Columbia River and North Portland Harbor

	Columbia River		North Portland Harbor		
Туре	Area (sq. ft.)	Duration in Water (days)	Area (sq. ft.)	Duration in Water (days)	
Work platforms/bridges for drilling shafts	148,000	120	29,640	up to 42	
Tower cranes	2,400	600	N/A	N/A	
Oscillator support platforms	N/A	N/A	27,900	Up to 33	
Barges for construction	106,432	Varies	1,085,000	up to 42	
Barges for demolition	42,000	~1	N/A	N/A	
Total	256,432		1,142,540		

Construction Barges

Barges would be anchored in the Columbia River and North Portland Harbor to serve as in-water work platforms during construction of in-water and overwater bridge elements. Stationary barges would be used at each of the in-water piers or bents. The shade footprint of moving barges (such as materials and spoils barges) was not included in this analysis. These barges move more or less constantly and on an unpredictable schedule, so it is impossible to quantify the extent or duration of shade cast by these sources.

Although the project would use numerous barges, there would be a limited number of barges in place at any one time. During construction in the Columbia River, there would likely be one to four stationary barges operating in the Columbia River at one time, casting no more than 120,000 square feet of shade at once. In North Portland Harbor, there would likely be no more than six crane barges operating at one time, creating a maximum of approximately 105,000 square feet of shade at one time.

In-Water Structures

The project would use temporary in-water work platforms, work bridges, tower cranes, and oscillator support platforms to support the equipment used to drill shafts in the Columbia River and North Portland Harbor.

In the Columbia River, there would be six temporary work platforms/bridges, one at each of the in-water pier complexes. At pier complexes 2 and 7, the work bridges would be L-shaped, approximately 17,500 square feet and 18,500 square feet in size, respectively. At pier complexes 3 through Pier 6, each work platform would cover an area of approximately 29,000 square feet (Exhibit 5-54.). Up to four platforms would be in place at one time. Once drilled shafts are

completed, the platforms would be removed. Six temporary tower cranes would be installed, one for each in-water pier complex. Each would shade an area of approximately 400 square feet. Including the work platforms, work bridges, and the tower cranes, roughly 125,000 square feet would be shaded at one time in the Columbia River.

In North Portland Harbor, the project would use nine work bridges of different sizes to build the nine bents nearest the shorelines. Only one or two work bridges would be in place at any given time. Additionally, the project would use 31 oscillator support platforms (900 square feet), one for each in-water shaft in North Portland Harbor. Only one to three oscillator support platforms would be in place at once. At any one time, in-water structures in North Portland Harbor would shade no more than 7,180 square feet altogether.

Demolition in the Columbia River would not require shade-producing in-water structures.

Demolition Barges

Demolition of the existing structures in the Columbia River would require one to three stationary barges at any one time, with a maximum shade footprint of approximately 21,000 square feet at once.

There would be no demolition or demolition barges in North Portland Harbor.

5.3.2.3 Potential Effects of Shading on Fish in the CRC Project Area

Temporary overwater structures, including the barges, work platforms, work bridges, oscillator support platforms, and tower cranes, are located at the water line and therefore could create new, high-intensity shade, potentially generating the type of intense light-dark contrast that could attract predators or cause visual disorientation to fish in the Columbia River and North Portland Harbor. This impact would be temporary, limited to the time that these structures are in the water (Exhibit 5-54.).

Temporary shading would not be uniform over all of the in-water construction years. In the Columbia River, shading would be limited to the first three pier complexes during the first year, expand to all six in the second, and taper off to three or fewer during the last two years. In North Portland Harbor, temporary shade would be distributed more or less evenly over the first two years of the in-water construction periods with more shade-producing activities concentrated in the last in-water construction year. Temporary shading would be evenly dispersed over the in-water demolition period.

Effects to Predation

The existing Columbia River and North Portland Harbor bridges likely attract predators, such as largemouth bass, smallmouth bass, and northern pikeminnow. The project increases the amount of shade in the project area compared to existing levels, but chiefly on a temporary basis.

It is impossible to quantify the extent to which increased shade may affect predation rates on juveniles. However, it is probable that an increase in predator habitat would increase predation pressure on juvenile salmonids and larval eulachon in the project area during daylight hours. Project-related sources of shade likely to attract predators (barges, temporary overwater structures, and shaft caps) are located in juvenile migration routes, creating an opportunity for predators to forage on juveniles during migration. Additionally, rearing juvenile salmonids are present in the project area and could experience increased predation pressure as a result of increased shade. Green sturgeon and bull trout would unlikely be subjected to increased predation

pressure, because only adult and subadults may use the project area and the risk of predation is extremely low for fish of this size (Zimmerman 1999). Likewise, adult salmonids are unlikely to be exposed, for the same reason.

The increase in shade along the nearshore could have particularly adverse effects on certain life stages of juvenile salmonids. In general smaller, rearing and subvearling migrant salmonids are highly dependent on the nearshore. Overwater structures that create a shadow line completely blocking the nearshore could force these runs into deeper water where they could be subjected to higher levels of predation. (It should be noted that the literature does not show widespread agreement on this effect, and therefore this result is not certain to occur). This scenario could occur in several locations: at the temporary work bridges at Columbia River pier complexes 2 and 7, the permanent new shaft cap at pier complex 7 in the Columbia River, and at all of the temporary work bridges in North Portland Harbor. While all runs of juvenile salmonids could be exposed to increased predation, species that rear in this portion of the project area (LCR Chinook, UCR spring run Chinook, UWR Chinook, LCR coho, and LCR steelhead) and species that migrate as subvearlings through this portion of the project area (CR chum and a portion of the LCR Chinook run) are generally more vulnerable to this effect both because they are dependent on the nearshore and because they are of a small size more easily captured by predators. It is not possible to quantify how many of these individuals would be exposed to increased predation in shallow water.

Effects to Orientation and Migration

As stated earlier, the literature is not in agreement as to whether the light-dark interface definitively causes visual disorientation or interference with migration in juvenile salmonids. This analysis assumes a worst case scenario, that is, that all new intense shade sources in the project area could result in visual disorientation during the day time. Assuming this is true, juvenile salmonids could be exposed to this effect during daylight hours when they are present in the Columbia River and North Portland Harbor portions of the project area.

For juvenile salmonids, visual disorientation could presumably lead to delayed migration and increased vulnerability to predation. The literature indicates that these effects are not certain to occur, and in any case, it is impossible to quantify the magnitude of these effects. The project would not create a swath of dense shade that completely spans either the Columbia River or North Portland Harbor stream channel. Therefore, even if the light-dark interface would prompt avoidance of the shadow zone, it would not likely completely block migration. Nighttime migration would be unaffected.

Eulachon larvae do not have volitional movement (Langness 2009 pers. comm.), and are therefore not subject to disorientation.

Green sturgeon are bottom feeders (NMFS 2008e) that inhabit portions of the stream channel with low light levels. Shade effects (particularly, a sharp light-dark interface) would not likely extend to the depths that green sturgeon inhabit. In addition, their presence in the project area is extremely limited. Therefore, green sturgeon would not be likely to experience visual disorientation as a result of increased shade in the project area.

Because bull trout abundance is extremely low in the project area and because the proportion of the project area that likely could be exposed to increased shade is very limited, the risk of exposure to this effect is discountable. Additionally, only adult and subadult bull trout could potentially occur in this portion of the project area, and these age classes are not subject to visual disorientation from shade.

The increase of shade in shallow-water habitat could have particularly adverse effects to species that are highly dependent on the nearshore for migration: CR chum and the portion of the LCR run that migrates as subyearlings. Shade could completely overlap shallow-water habitat at the temporary work bridges at the Columbia River pier complexes 2 and 7 and at all of the temporary work bridges in the North Portland Harbor, potentially prompting these salmonids to swim into deeper water to circumvent the shadow line. It is not possible to quantify how many individuals could experience delayed migration due to the presence of shade in the nearshore.

Effects to Primary Productivity

The project would not be expected to cause significant impacts to primary productivity or the food web for any of the fish species using the project area. The project could reduce the productivity of plants, algae, and phytoplankton occurring both within the photic zone and beneath overwater structures. However, shade would be limited to localized, discrete areas, measuring no more than several hundred to several thousand square feet. Barges, work platforms, tower cranes, and oscillator support platforms would be temporary sources of shade on the CRC project.

Although the project could result in loss of primary production in shaded areas, this loss is not likely to significantly impact the food web. The project area does not contain habitats that are known to support high primary and secondary productivity for fish. In northwest estuaries, such habitats include areas that produce and retain high levels of detritus: floodplains, vegetated riparian areas with overhanging vegetation, shallow marshes, tidal creeks, dendritic channel networks, low intertidal and subtidal eelgrass beds, emergent vegetation in tidal wetlands, and macroalgal beds (such as mudflats and sandflats). In the Columbia River estuary, detritus is concentrated in low-velocity peripheral bay habitats (Bottom et al. 2005). These habitats are completely lacking in the project area, which is dominated by high-velocity open water that is severed from the historical floodplain and lacks emergent vegetation, structural complexity, and riparian areas with overhanging vegetation. In areas of the upper estuary that lack these habitat features, there has been a shift from detritus-based primary production to production dominated by phytoplankton. This has led to widespread loss of food webs supporting epibenthic feeders such as juvenile salmonids (Bottom et al. 2005). This type of food web also favors production of a microdetrital food web dominated by simple-celled plants and organic particles (NMFS 2005b), as well as calanoid copepods and other organisms that are not consumed by juvenile salmon (Bottom et al. 2005). Because of the shift in the food web, the suspension/deposit feeder Corophium salmonisis is now the most abundant prey item of juvenile salmonids in the estuary. This species is a poor food source because it is low in protein and high in chitin (NMFS 2005b). Because the project area lacks detritus-rich habitat types and harbors a microdetrital food web, it provides only limited, low-quality foraging habitat and food web support for salmonids.

In shallow-water areas of the lower Columbia River, the majority of primary productivity is driven by benthic algae, with some contribution from filamentous algae and flowering grasses. Within the water column, primary productivity is driven by phytoplankton (NMFS 2005b). Because shallow-water habitat is limited in the project area, the majority of primary productivity is likely driven by phytoplankton.

There have been no known surveys of underwater vegetation or periphyton in the project area. However, in the lower Columbia, small diatoms (*Achnanthes, Cocconeis*, and some filamentous blue greens) are expected to be present. Other grazing-resistant algae are expected to be present on the riprap along shorelines and on bridge piers, together with filamentous green algae (such as *Cladophora*) and its associated epiphyton (for example, *Rhoicosphenia, Cocconeis*, and *Epithemia*). Red algae are also probably very common (Carpenter 2010 pers. comm.).

Typical macrophytes in the lower Columbia River include *Potomogeton crispus*, *Elodea* (*cascadensis*, *nuttallii*, and others), *Ceratophyllum*, and possibly *Hetheranthera dubia*. However, macrophytes are likely not present or are very limited in the project area, as they typically occur in backwater areas (Carpenter 2010 pers. comm.). Because underwater portions of the project area are characterized by high current velocity and an armored streambank, backwater areas are generally lacking, and thus, macrophytes are limited. Additionally, substrate is unstable sand, and the underwater topography slopes off steeply, reducing the size of the photic zone.

Because the project area lacks high-productivity, detritus-based plant communities and is dominated by plankton and periphyton, shading would not impact habitats of particularly high quality. Additionally, outside of the areas potentially influenced by shading, the surrounding area contains dozens of square miles of water available for primary production both upstream and downstream. Shading would only impact a tiny fraction of the remaining area available for primary production, such that there is no measurable reduction in baseline levels of production. All of the listed species that forage in the project area are highly mobile and can readily move to these nearby areas in response to localized impacts to vegetation or the food web. Because the impact is small relative to the amount of habitat present in the surrounding area, this effect would be insignificant.

Beneficial Effects of Shade

Shade may also confer benefits to salmonids using the project area. Salmonids require cool water to perform life history functions. Temperatures of 50 to 57°F are considered adequate to support spawning, migration, and rearing (Bjornn and Reiser 1991). The 303(d) listings for the Columbia River portion of the project area indicate that temperature exceeds standards for spawning, migrating, and rearing salmonids during summer months in the Columbia River and North Portland Harbor (DEQ 2009), with measured temperatures ranging as high as 72°F (USGS 2007). Overwater structures may create shade, resulting in localized areas of cooler water. The temporary overwater structures would create new areas of dense shade that could potentially provide an increase in summertime cool-water refugia compared to the current condition. These increases in shade may confer a benefit to migrating and rearing salmon, although it is impossible to quantify to what extent.

Pacific lamprey are light-sensitive and migrate primarily at night. Lights that have been used at the Columbia River dams for night video migration counts have been shown to repel Pacific lampreys (Ocker et al. 2001; NPCC 2004). Therefore, shading would not affect Pacific lamprey but could be affected by artificial lights used at night during construction.

Increased shade would have little or no effect on sea lions since predation rates of salmon and sturgeon, two prey species, are not anticipated to change in a way that would affect population size. Furthermore, Steller sea lions have been shown to rely more on white sturgeon as a food source than salmon (Tackley et al. 2008).

5.3.3 Artificial Lighting over Water

The project would require several new sources of overwater artificial lighting to be used during nighttime construction. The following sections outline the general effects of lighting on fish and provide an analysis of the likely effects on fish in the CRC project area.

5.3.3.1 General Effects of Artificial Lighting on Fish

Artificial light sources associated with overwater structures or construction activities may attract fish. Because salmon rely on vision for capturing prey, the artificial lights may improve both prey

detection and predator avoidance (Tabor et al. 1998, as cited in Carrasquero 2001). During a study of the Columbia River at Bonneville Pool, Collis et al. (1995) observed that juvenile salmon were attracted to work lights directed at the water surface. In Lake Washington, juvenile Chinook have been observed congregating at night near streetlights on the SR 520 bridge (Celedonia et al. 2008). Tabor et al. (2004) observed sockeye fry in the Cedar River, noting that they were significantly more abundant under city street lights than at nearby sites that were not illuminated. Light levels as low as 0.22 lux (0.020 foot candle) appeared to influence fry behavior. In one location, turning off the streetlights resulted in a significant decrease in the number of sockeye fry present.

Artificial lights may create sharp boundaries between dark and light areas under water. This, in turn, may cause juvenile fish to become disoriented or avoid crossing the light-dark interface, as outlined in detail in Section 5.3.2.1. Williams and Thom (2001) noted that artificial lighting on docks may change nighttime movement patterns in juvenile salmon. Numerous other studies (Fields 1966; Prinslow et al. 1979; Weitkamp 1982; Ratte and Salo 1985; Pentec 1997; Taylor and Willey 1997; and Johnson et al. 1998, as cited in Southard et al. 2006) corroborate these findings, noting behavioral changes in juvenile salmon in response to artificial lighting. McDonald (1960, as cited in Tabor et al. 2004) found that sockeye fry will stop swimming downstream upon encountering artificial lighting, and was able to completely stop nightly migration of sockeye salmon fry with artificial lighting kept on all night at 30 lux (2.8 foot candles). A USFWS (1998) literature review noted that sockeye fry moved through experimental streams more quickly in complete darkness than under bright lights (Tabor et al. 1998). Increased light appeared to inhibit migration of sockeye fry, with significant effects to migration when light levels reached 2.0 lumens/ft² (2.0 foot candles). A later study (Tabor et al. 2004) corroborated the finding that fewer sockeye moved through illuminated artificial streams than in darkness, and those that did move, moved more slowly. In this study, light intensity levels from 1.08 to 5.40 lux (0.1 to 0.5 foot candle) appeared to inhibit migration. The same study noted that the delay in outmigration in sockeye fry increased their vulnerability to predation.

Another USFWS study (Tabor and Piaskowski 2001) observed juvenile Chinook in nearshore habitat in Lake Washington, noting that individuals became active when light levels reached 0.08 to 0.21 foot candle and were scarce in the study area when light levels were between 2.2 to 6.5 foot candles. A review of the impact of ferry terminals on juvenile migration in Puget Sound (Simenstad and Nightingale 1999) cites Ali (1958, 1960, and 1962) as stating that light is tremendously important for numerous life functions of chum, coho, sockeye, and pink salmon, noting that feeding, minimum prey capture, and schooling are dependent on light levels lower than 10⁻⁴ foot candles (similar to a clear, moonless night) and that maximum prey capture for chum and pink fry occurs when the light level is 1.0 foot candle (similar to light levels at dawn and dusk).

Artificial light sources may provide an advantage to predators such as smallmouth bass, largemouth bass, northern pikeminnow, and salmonids. Rainbow trout predation on sockeye fry in artificial streams increased with increased lighting at levels of less than 1.1 lux (Ginetz and Larkin 1976, as cited in Tabor et al. 2004). Northern pikeminnow are attracted to areas where juvenile salmonids congregate, such as hatchery release sites and dams (Collis et al. 1995; Beamesderfer and Rieman 1991). If light sources attract congregations of juvenile salmonids, this could cause an increase in predation by northern pikeminnow. Celedonia et al. (2008) found that smallmouth bass may feed at night in the vicinity of artificial light or under moonlight. Largemouth bass have been shown to forage efficiently at light levels ranging from low-intensity daylight to full moonlight, with less foraging at light levels equivalent to a starlit, moonless night (McMahon and Holanov 1995).

Tabor et al. (2004) observed the effect of light intensity on cottid predation of sockeye fry in artificial streams, noting that cottids consumed 45 percent of the fry under intense illumination (5.4 lux or 0.50 foot candle), 28 percent under dim light (0.22 lux or 0.020 foot candle), and 5 percent in complete darkness (0 lux or 0 foot candle). The study also observed that fewer fry emigrated in illuminated streams and did so at a faster rate when predators were present than in lighted streams where predators were not present, indicating that the presence of predators may inhibit migration in some individuals. In a field study in the Cedar River, Washington, Tabor et al. (2004) further noted that the number of shoreline fry and rates of predation by cottids increased with an increase in light levels. At one site, shielding the lights to levels of 0.1 to 0.32 lux (0.013 to 0.030 foot candle) substantially reduced predation.

The literature is not in complete agreement about light levels that are likely to impede migration or increase predation on juvenile fish. However, data from Tabor et al. (2004) may present a worst-case scenario; that is, light levels as low as 0.22 lux (0.20 foot candle) may delay migration or increase predation on juvenile salmonids.

5.3.3.2 Effects of Lighting on Fish in the CRC Project Area

Temporary overwater lighting sources would include the cofferdams, barges, work platforms/bridges, oscillator platforms, and tower cranes. Temporary lighting would not be uniform over all of the in-water construction years. During the Columbia River in-water construction period, temporary lighting would be limited to the first three pier complexes during the first year, expand to all six in the second, and taper off to three or fewer during the last 2 years. In North Portland Harbor, temporary lighting would be distributed more or less evenly over the first 2 years of the in-water construction periods with illumination-producing structures concentrated in the last in-water construction year. Temporary lighting would be distributed evenly across the Columbia River in-water demolition period.

The barges and temporary in-water structures would cast light at the water surface during construction and demolition in the Columbia River and North Portland Harbor. At this stage in the project design, the intensity of light likely to be cast on the water surface is not known. However, to the extent practicable, the project would implement conservation measures that minimize the effects of lighting on fish. Measures may include using directional lighting with shielded luminaries to control glare and to direct light onto work areas instead of surface waters.

It is impossible to quantify how many fish would be exposed to increased lighting; however, all of the juvenile fish that use the project area could be exposed to this effect when they are rearing in or migrating through the project area.

It is possible that the increase in lighting in the project area could cause some interference with juvenile salmonid migration. Overwater structures would be limited to discrete locations measuring from several hundred to several thousand square feet and would only span a fraction of the entire channel. While lighting may prompt juvenile fish to avoid the illuminated area, it would not constitute a complete barrier to migrating juvenile fish.

It is also possible that rearing and migrating juvenile salmonids could congregate under light sources, potentially becoming exposed to an increased risk of predation than they are currently exposed to. As with effects to migration, it is impossible to quantify the extent to which predation would increase. However, it seems likely that an increase in the conditions that confer an advantage to visual predators could increase levels of predation. Rearing juveniles (LCR Chinook, coho, and steelhead) are present in the area for a relatively large proportion of the year, and therefore could be especially vulnerable to this effect.

Illumination in shallow water may place subyearling migrants (LCR Chinook and CR chum) at particular risk, as these individuals are highly dependent on nearshore areas. This effect is discussed in greater detail in Section 5.3.2.3.

The project would implement BMPs during in-water and upland construction activities to avoid and minimize impacts to water quality. Without implementation of BMPs, water quality could be impacted in a number of ways. Chemical contamination could potentially occur through the accidental release of construction materials or wastes. Upland excavation could lead to erosion, causing turbidity in adjacent water bodies. In-water work (such as pile driving, demolition, debris removal, barge use, and installation of bridge piers) could generate turbidity directly in waterways. The implementation of BMPs would help ensure that these effects would be localized and temporary, limited to the duration of the project, and would result in minimal impacts to water quality.

This section describes the sources of effects to water quality, outlines the BMPs that would be used to contain them, and analyzes the potential effects to listed fish. Section 6 outlines the BMPs in further detail.

5.3.3.3 Chemical Contamination

There are numerous potential sources of chemical contamination associated with in-water work in the Columbia River and North Portland Harbor. Potential contamination sources include the following:

- Equipment located in or over water (such as barges or equipment operating on barges, temporary work platforms, the existing structure, or the new structure) are potential sources of contamination.
- Uncured concrete would be used in numerous locations both in and over water for the construction of the piers and superstructure for the new bridges.
- Construction of the superstructure would involve the use of numerous other potential contaminants, including various petroleum products, adhesives, metal solder, concrete and metal dust, asphalt, and others.
- Bridge demolition would occur both in and over water and may release contaminants such as concrete debris, concrete dust created by saw cutting, lead paint, creosote-treated wood, and others.
- There are a total of approximately 1,800 timber piles at the nine existing Columbia River bridge piers. It is assumed that these piles have been chemically treated, based on their age and intended purpose. Contaminants from the piles could be mobilized during demolition of the piers.

Although there are several sources of chemical contaminants, there is a low risk that chemicals would actually enter the Columbia River and North Portland Harbor. A spill prevention, control, and countermeasures (SPCC) plan would be implemented to completely contain sources of chemical contamination such as equipment leaks, uncured concrete, and other pollutants.

During construction of the drilled shafts, uncured concrete would be poured into water-filled steel casings, creating a mix of concrete and water. As the concrete is poured into the casing, it would displace this highly alkaline mixture. The project would implement BMPs to contain the mixture and ensure that it does not enter any surface water body. Once contained, the water would be treated to meet state water quality standards and then either released to a wastewater treatment facility or discharged to a surface water body.

In-water bridge demolition would take place only in the Columbia River. All demolition activities would be completely contained within cofferdams. The contractor is required to prepare a demolition plan according to ODOT and WSDOT standard specifications. The plan would be submitted to ODOT and WSDOT and would not be implemented without being approved and stamped by a registered professional engineer. The demolition plan would specify containment methods to ensure that bridge elements and wastes do not enter the Columbia River. Breaking up the concrete piers with an excavator or saw cutter could potentially introduce concrete dust into the water; however, because of the containment proposed, there is minimal risk that dust or debris would enter the Columbia River during demolition. Any concrete wastes would be allowed to settle in the cofferdams before the cofferdams are disassembled. During removal of the cofferdams, released water would meet state water quality standards.

Removal of the timber piles that are deemed navigational hazards and located beneath the existing Columbia River bridge piers would be contained within cofferdams during the demolition of the rest of the piers. There may be, however, some piles that must be removed and are located outside of the cofferdam footprint. These would likely be cut off at or below the mudline. No containment is proposed for the removal of these pilings. However, given the high flow in the Columbia River, dilution of contamination is likely to be high, and the extent of the contamination is expected to be minimal.

The project would obtain several regulatory permits that include terms and conditions for controlling and containing chemical releases to surface water bodies. These permits include Ecology's 401 Water Quality Certification, WDFW hydraulic project approval (HPA), DEQ's 401 Water Quality Certification, DSL's Removal/Fill Permit, and USACE's 404 Removal/Fill Permit. The project would adhere to the terms and conditions of all of these permits further minimizing risks to water quality in the Columbia River and North Portland Harbor.

In general, construction equipment operating on land poses a low risk of releasing chemical contaminants (such as petroleum fuel or other fluids) that could enter surface water bodies by way of stormwater inlets, ditches, or other forms of conveyance. Implementation of a pollution control plan would minimize the risk of landward contaminants entering water, to ensure that the risk of contaminant release is discountable. These measures are outlined in greater detail in Section 6. Overall, this aspect of the project is not likely to appreciably impact fish.

5.3.3.4 Turbidity and Suspended Sediment

The project is likely to generate turbidity during the course of in-water work in the Columbia River and North Portland Harbor. Exhibit 5-55 lists the activities that could potentially generate turbidity downstream of each activity and summarizes the effect to the environmental baseline in the Columbia River and North Portland Harbor.

Exhibit 5-55. Potential Sources of Turbidity

			Likely Extent of	Duration	
Activity	Timing ^a	Location ^b	Downstream Turbidity	of Effect (hr/day)	Number of Work Days
Install temporary piles, impact methods	9/15–4/15	Adjacent to P2 – P7 in CR	~25 feet	0.66	138 in CR 134 in NPH
		Adjacent to new NPH shafts			
Install temporary piles and cofferdams, vibratory	Year-round	Adjacent to P2 – P7 in CR	~25 feet	Up to 24	Continually over ~1015 days in CR
methods		Adjacent to new NPH shafts			~334 in NPH
Remove temporary piles and cofferdams, direct pull	Year-round	Adjacent to P2 – P7 in CR	Minimal	Up to 24	Continually over ~1015 days in CR
or vibratory		Adjacent to new NPH shafts			~334 days in NPH
Install steel casings to drill	Year-round	P2 – P7 in CR	~25 feet	8 – 10	250 / CR pier
permanent shafts – vibratory hammer, oscillator, or rotator		New NPH shafts			<1 / NPH shaft
Drill and excavate permanent shafts	Year-round	P2 – P7 in CR New NPH shafts	Minimal (contained)	n/a	100 / CR pier ≤8 / NPH shaft
Operate stationary and	Year-round	P2 – P7 in CR	<300 feet	Varies	Continually over
moving barges in shallow water		new NPH shafts			~1015 days (CR) ~640 in NPH
Debris removal (clamshell)	11/1–02/28	Potentially at 31 locations in NPH.	~300 feet (or as prescribed by permits)	4-6 hr/day, ≤ 4x/day	<7 days
Demolish existing Columbia River bridges' piers (includes installation of cofferdams)	Year-round	Existing Piers 2 - 11 in CR	Minimal	8 – 10	~266

a All activities likely to take place within the 4-year in-water construction period.

Potential Effects to the Environmental Baseline

The project would employ numerous BMPs to minimize the extent and duration of turbidity. These BMPs may include (but would not be limited to) a spill prevention/pollution control plan, an erosion control plan, and others as outlined in Section 6. The exact BMPs have not yet been determined. However, these BMPs would ensure that the amount and extent of turbidity would meet the terms and conditions of the two Section 401 Water Quality Certifications that would be obtained from DEQ and Ecology. The certifications would specify a mixing zone for turbidity: that is, a specified distance beyond which turbidity may not exceed ambient levels downstream of the source. We anticipate that the permits would specify a mixing zone of 300 feet downstream of turbidity-generating activities, as this is typical for water bodies the size of the Columbia River and North Portland Harbor (that is, with flows of 300 cubic feet per second [cfs] or greater). Typically, these permits allow exceedance of ambient levels of turbidity for a period of 4 hours within the mixing zone and 2 hours outside of the mixing zone, after which the applicant must stop work until the turbidity dissipates to ambient levels. The project would implement regular water quality monitoring in accordance with the permits to ensure that the project adheres to the permit conditions, with cessation of work if conditions are not met.

b CR = Columbia River; NPH = North Portland Harbor, P = pier complex.

In actuality, many of the activities listed in Exhibit 5-55 are not expected to generate large amounts of turbidity. The following activities are expected to generate turbidity at far shorter distance than the anticipated 300-foot mixing zone: installation of piles and cofferdams using impact or vibratory methods; removal of piles and cofferdams using direct pull or vibratory methods; installation of large diameter steel casings using an oscillator, rotator, or vibratory hammer; and demolition activities contained within a cofferdam. These activities do not involve in-water excavation and disturb relatively small amounts of material; therefore, the potential for generating turbidity is greatly reduced.

EPA advises that turbidity extend no more than 25 feet from the source during impact or vibratory pile installation (WSF 2009). Assuming that this is an average value observed over a range of substrate types and flow levels, we expect this threshold distance to be achievable on the CRC project. The Columbia River and North Portland Harbor are large water bodies, providing very high levels of dilution, and reducing size of the potential mixing zone. Additionally, substrates in these water bodies are coarse sand, which settles in relatively short distances compared to finer sediments. Given these mitigating circumstances, we expect that turbidity levels in the CRC project area would be similar to average conditions in other streams, or at least not exceed them. Therefore, we expect that the turbidity would extend to no more than 25 feet from installation of piles, cofferdams, and the steel casings for drilled shafts.

Few studies document the magnitude or extent of turbidity resulting from pile removal. Roni and Weitkamp (1996) reported that pile removal in Manchester, Washington, generated turbidity at less than 1 nephelometric turbidity unit (NTU) above background levels. Washington State Ferries (WSF) performed water quality monitoring during pile removal at Friday Harbor Ferry Terminal; they reported that turbidity levels did not exceed 1 NTU above background levels and were less than 0.5 NTU above background for most of the samples. WSF also performed water quality monitoring during pile removal at Eagle Harbor Maintenance Facility in 2005, reporting that removal of steel and creosote pile resulted in turbidity levels of no more than 0.2 NTU. These values represent extremely small increases above background turbidity levels. Given that the Columbia River and North Portland Harbor have very high dilution capacity and given that substrate in the project area is coarse sediment that settles readily, it is expected that turbidity generated by removal of piles and cofferdams would dissipate within a minimal distance Specifically, it is assumed that this distance would be less than that for pile installation (25 feet), as pile removal displaces less sediment than pile installation.

Drilling and cleaning the permanent shafts would introduce only minimal amounts of sediment into the water. All of the drilling and excavation would occur within the closed steel casings. To the extent practicable, excavated materials would not be allowed to enter the water, but would be stored in contained areas on the barges or work platforms and transported to a permitted upland disposal site.

Debris removal is the only aspect of in-water work likely to generate significant amounts of turbidity. Debris removal could potentially occur at discrete locations in North Portland Harbor. While debris removal is not certain to occur, this information is presented as a worst-case analysis.

There are anecdotal reports that remnant pieces of the original North Portland Harbor bridge (including riprap used as scour protection), still remain on the stream floor. The exact location of the material is not known, but the design team believes that it occurs in several scattered locations, potentially within the footprint of any of the new North Portland Harbor bridge shafts. If this is the case, the material must be removed before drilled shafts can be installed in these locations. Before debris removal begins, divers would pinpoint the locations of the material.

Debris removal would be performed only in the precise locations where the material occurs within the footprint of the new bents, greatly minimizing the areal extent of the activity. As stated previously, the amount of material in this location is not known. Assuming a worst-case scenario, that the area of the material is the same as the footprint of the drilled shafts, the project would remove debris at each of the 31 new bridge shafts (encompassing an area of roughly 2,433 square feet, total). The design team estimates that no more than 90 cubic yards of material would be removed.

Due to the large size of North Portland Harbor, the design team anticipates that it would not be possible to install physical BMPs to contain turbidity during debris removal in these locations. Regardless, the project would comply with the terms of all permits related to in-water turbidity, and turbidity would not exceed the levels, distance, or duration specified by the permits. Depending on the permit specifications, the turbidity plumes are expected to reach no more than 300 feet downstream of the source for a duration of no more than 4 to 6 hours. In all cases, debris removal would be performed using a clamshell at a slow, controlled pace to minimize turbidity.

Barges operating in shallow water have the potential to produce turbidity at pier complexes 2 and 7 in the Columbia River and at all of the new North Portland Harbor bents. Barges would have a draft depth of about 13 feet and would operate in water as shallow as 20 feet deep. Therefore, barge propellers may produce turbulence that causes sediments to become suspended. Additionally, tug boats that position barges may also have propellers that generate suspended sediment. Tug boats would operate only during discrete time periods to (1) position the work barges at each of the shallow-water piers (pier complex 7 in the Columbia River and all North Portland Harbor bents) and (2) to remove them when work is completed. These barges would remain stationary for the duration of the work, and therefore have little potential to produce turbidity. Additionally, there would be one or two barges at each of the shallow-water piers used to store and move materials and dredge spoils. These barges would make numerous trips, as needed, operating on a sporadic schedule. Because the schedule is unknown, it is not possible to predict the timing and duration of the turbidity plumes. In any case, the size of the plumes is expected to be much smaller than the typical plume created by dredging (estimated to be no more than 300 feet). Given that sediment in this portion of the project area consists mainly of coarse material with only minor amounts of fines, suspended sediment is expected to settle quickly, further restricting the size of the potential turbidity plume. Additionally, compared with the existing energy generated by high velocity flow in this portion of the project area, disturbance of sediment by tug and work boat propellers is expected to be minimal. Because little aquatic vegetation is present in this portion of the project area, turbidity generated by barges and tug boats is not expected to have a significant impact on underwater vegetation. In any case, turbidity would not exceed the levels, distance, or duration specified by the permits. Construction barges would not be grounded.

Demolition would involve cutting, breaking, and removing the nine existing Columbia River bridge piers. Exact demolition methods are unknown at this time and would be determined by the contractor at a later date. However, the CRC team anticipates that all demolition work would be performed from barges and would be completely contained inside of enclosed cofferdams. Installation and removal of the cofferdams is the only aspect of bridge demolition likely to cause turbidity. Turbidity is likely to extend only a minimal distance from the source (Exhibit 5-55) and could potentially be present for the duration of the time it takes to install or remove each cofferdam. Installation of the cofferdam, demolition of the pier, and removal of the cofferdam is expected to take 40 days throughout the 18-month in-water demolition period. In any case, turbidity would not exceed the levels, distance, or duration specified by the permits.

In general, upland excavation has the potential to cause erosion, which in turn may introduce suspended sediments into water bodies by way of stormwater inlets, ditches, or other forms of conveyance. However, it is not likely that upland construction would cause turbidity in the CRC project area water bodies. To prevent the introduction of sediments into waterways from upland excavation, the project would adhere to an erosion control plan that specifies the type and placement of BMPs, mandates frequent inspections, and outlines contingency plans in the event of failure. Additionally, in many cases, there would likely be numerous other barriers between the potential sources and the project area water bodies. Therefore, there is only a discountable risk that upland excavation would generate turbidity in project area water bodies. Erosion control specifications are outlined in further detail in Section 6.

General Effects of Turbidity on Fish

Turbidity is a naturally occurring phenomenon; however, turbidity above background levels may harm fish. Several factors contribute to turbidity levels in water, including suspended sediments, dissolved particles, finely divided organic and inorganic matter, chemicals, plankton, and other microscopic organisms. Not all of these materials are necessarily harmful, meaning that turbidity levels alone may not accurately indicate the effect on fish. TSS, a direct measure of particles transported in the water column, may be a more useful indicator of the effect to fish. However, due to the ease of taking turbidity measurements, turbidity is in widespread use throughout the literature as an indicator of the effect of suspended sediments on fish (Bash et al. 2001).

The response of fish to turbidity is complex. High levels of turbidity may be fatal to salmonids, but salmonids may also be affected by turbidity at relatively low levels (Lloyd 1987). Juvenile salmonids have been observed in naturally turbid estuaries and highly turbid glacial streams, which indicates that that salmon are able to cope with elevated turbidity during certain life stages (Gregory and Northcote 1993, as cited in Bash et al. 2001). In contrast, salmonids not normally exposed to elevated turbidity levels may be appreciably impacted at relatively low levels (Gregory 1992, as cited in Bash et al. 2001). The severity of effect depends on a variety of factors, such as the turbidity level, extent of the turbidity plume, the duration and frequency of exposure, the toxicity and angularity of the particles, life stage of the fish, and access to "turbidity refugia" (Bash et al. 2001). Depending on the amount of exposure, turbidity above background levels may prompt the following effects: direct mortality, gill tissue damage, physiological stress, and behavioral effects.

Numerous studies document that direct mortality for juvenile salmonids occurs at a 96-hour median sediment concentration of 6,000 mg/L (Stober et al. 1981 as cited in Bash et al. 2001; Salo et al. 1980; LeGore and DesVoigne 1973 as cited in WSF 2009).

Suspended sediments have been shown to damage gill structure (Noggle 1978). When the filaments of salmonid gills are clogged with sediment, fish attempt to expunge the sediment by opening and closing their gills excessively, in a physiological process known as "coughing." In response to the irritation, the gills may secrete a protective layer of mucus. Although this may interfere with respiration, it is not a lethal effect (Berg 1982, as cited in Bash et al. 2001). Servizi and Martens (1992) noted a significant increase in coughing in subyearling coho when turbidity measured 30 NTU. Berg (1982, as cited in Bash et al. 2001) observed a significant increase in coughing in juvenile coho at 60 NTU, with a decline or return to pre-exposure levels of coughing at 10 NTU. This indicates that turbidity somewhere between 10 and 30 NTUs may cause onset of coughing. Servizi and Martens (1987) found that gill trauma occurred in subyearling sockeye at suspended sediment concentrations of 3,148 mg/L.

The literature indicates that exposure to suspended sediments may cause stress response in both adult and juvenile salmonids. Physiological stress generally manifests itself as elevated blood sugar, plasma glucose, and plasma cortisol (Bash et al. 2001). Redding et al. (1987) observed physiological stress in subyearling coho after exposure to sediment concentrations of 2,000 mg/L for 7 to 8 days. Servizi and Martens (1987) observed elevated blood glucose levels in adult and juvenile sockeye after contact with fine sediment. In adults, this response occurred at concentrations of 500 to 1,500 mg/L after exposure for 2 to 8 days. At levels of 150 to 200 mg/L, no stress response was observed (Redding et al. 1987; Servizi and Martens 1987). At the individual level, stress may reduce growth, increase the likelihood of disease, inhibit the development from parr to smolt, disrupt osmotic balance, impair migration, and reduce survival (Wedermeyer and McLeay 1981, as cited in Bash et al. 2001). At the population level, stress may reduce spawning success, increase larval mortality, and decrease overall population abundance (Bash et al. 2001).

Turbidity may also prompt behavioral responses in fish, including avoidance, migration delays, and changes in foraging and predation. Numerous studies document salmonids avoiding suspended sediments and migrating to less turbid areas (Berg 1982; Sigler et al. 1984). Lloyd et al. (1987) showed that juvenile salmonids avoid streams that are chronically turbid unless they cannot avoid these areas on their migration path. Cederholm and Salo (1979) showed that the upstream migration of salmonids in the lower Columbia River may be delayed when water clarity is reduced. On the other hand, adult male Chinook experienced no disruption in migration to spawning grounds after exposure to sediment concentrations of 650 mg/L over 7 days.

The literature is not in complete agreement as to whether or not turbidity increases the rate of prey capture in salmonids. Some studies reveal that fish have decreased foraging success in response to increased turbidity (Berg 1982; Berg and Northcote 1985; Redding et al. 1987; Gardner 1981 as cited in Bash et al. 2001; Boehlert and Morgan 1985 as cited in Bash et al. 2001; Vogel and Beauchamp 1999 as cited in Bash et al. 2001). One study showed decreased foraging at levels as low as 20 NTU (Berg 1982). In contrast, other studies show that juvenile coho, steelhead, and Chinook have increased foraging success in "slightly to moderately turbid" water (Sigler at al. 1984; Gregory and Levings 1998). There is also evidence that suspended sediments may offer cover from predators (Gregory 1993; Gregory and Levings 1996; Davies-Colley and Smith 2001), which may both enhance survival and increase foraging success.

Turbidity and concurrent sedimentation may negatively affect survival of eggs and emergence of fry or larvae. After being deposited in spawning areas, high levels of fines may become embedded in the substrate, reducing the permeation of oxygen into eggs, potentially resulting in mortality. Additionally, deposition of sediment may physically block the emergence of fry or larval fish (Cederholm and Salo 1979).

Effects on Fish in the CRC Project Area

There are few water quality monitoring studies that cite turbidity levels encountered during installation and removal of piles, cofferdams, and steel casings. Due to the lack of data, the analysis of the effects of turbidity on fish is based on turbidity levels observed during dredging, for which there are numerous monitoring studies. Havis (1988, as cited in WSF 2009), Salo et al. (1979, as cited in WSF 2009), and Palermo et al. (1990, as cited in WSF 2009) note that typical samples collected within 150 feet of dredging contain sediment concentrations between 50 and 150 mg/L. LaSalle (1988, as cited in WSF 2009) concluded that maximum sediment concentrations resulting from dredging range between 700 and 1,100 mg/L at a distance of approximately 300 feet from the source, based on monitoring data from seven clamshell dredging operations. These levels would be expected for dredging of fine sediments such as silt or clay.

Much lower concentrations, 50 to 150 mg/L, would be expected for dredging in coarser substrates (LaSalle 1988). The CRC in-water project area contains a mixture of coarser sediments and silty sand. Therefore, the amount of turbidity encountered during debris removal is likely to be more than 50 to 150 mg/L but is not expected to exceed 700 to 1,100 mg/L. Turbidity levels for the other activities listed in Exhibit 5-55 (installation and removal of piles and cofferdams, installation of large steel casings, barge use, and drilling shafts) are expected to be much lower than levels resulting from dredging.

Turbidity levels on the CRC project would not be expected to reach levels that cause mortality in fish. The highest sediment concentrations expected to occur (1,100 mg/L) would be well below levels known to kill fish (6,000 mg/L). Likewise, turbidity levels on the CRC project would not likely cause gill trauma, as gill trauma occurs at roughly 3,000 mg/L, well above the highest levels of turbidity expected on the project. However, turbidity would likely reach levels that could cause "coughing." Coughing may occur at 30 NTU, a value roughly estimated to be greater than 100 mg/L (Lloyd 1987). Actual exposure to these levels would be expected to be minimal, however. Regulatory permits would require restricting the size of the plumes (probably to about 300 feet from the source) and their duration (about 4 to 6 hours). Additionally, because of the large size and the high dilution capacity of the Columbia River and North Portland Harbor, there are abundant turbidity refugia, and listed fish should not become trapped in turbid water. The turbidity would be localized in areas downstream of specific activities (Exhibit 5-55) and would not extend across the entire width of the Columbia River or North Portland Harbor. Therefore, it would not cause a complete barrier to movement. Thus, while turbidity levels could theoretically be high enough to prompt coughing in fish, it is unlikely that the duration and extent of exposure would be great enough to cause gill damage.

The project could produce turbidity at levels that could cause physiological stress in fish. Of the studies available, the data indicate that stress may occur at a minimum level of 500 mg/L after several days of exposure. The project could generate a maximum of 1,100 mg/L of sediment concentration, but most activities would generate levels more on the order of 50 to 150 mg/L. On the CRC project, the actual duration of exposure to elevated turbidity could be quite low, as regulatory permits would restrict the size and duration of the turbidity plumes, probably to about 300 feet and to about 4 to 6 hours at a time. Additionally, because of the large size and the high dilution capacity of the Columbia River and North Portland Harbor, listed fish would be able to avoid the turbidity plumes and not become trapped in turbid water. The turbidity would not cause a complete migration barrier. Thus, while turbidity levels could theoretically be high enough to prompt stress in fish, it is unlikely that the duration and extent of exposure would be great enough to cause stress.

It is highly likely that turbidity generated by the project would cause both adult and juvenile fish to avoid discrete portions of the work area (Exhibit 5-55), as avoidance has been documented at very low turbidity levels. Turbidity-generating activities would be ongoing for the duration of the 4-year in-water construction period, and, therefore, these activities are likely to intersect up to four migration periods of juvenile salmon and steelhead. The exception is debris removal, which would likely intersect only about 7 days of one juvenile migrational period. Fish would likely circumvent the turbidity plumes and swim into less turbid areas. Whether this avoidance would result in a biologically significant effect is less clear. Although the literature shows that juvenile salmonids may delay migration in response to high turbidity, this may not necessarily be true in the CRC project area for two reasons. First, due to the large size of the Columbia River and North Portland Harbor, turbidity refugia would be abundant, and juvenile fish would probably circumvent the plumes with no significant delay to migration. Second, larger sediment plumes (anticipated to be no more than 300 feet) would occur in the project area for no more than roughly 4 to 6 hours at a time. Therefore, there is ample time for juveniles to migrate between sediment

pulses, and even if there were a delay, it would only be for a matter of hours. Adults have not been shown to delay migration even after many days of exposure to high turbidity. Because the CRC project would cause only low exposure (due to the abundance of turbidity refugia) over a limited spatial extent and over short durations, delays to adult migration are not probable.

Turbidity would likely reach levels that have been shown both to enhance and impede foraging abilities in fish. Therefore, we can expect that turbidity generated by the project would cause fish in the project area to increase foraging in some circumstances and decrease foraging in others. There is also evidence that turbidity may provide cover from predators, creating a benefit to juvenile fish. However, due to the uncertainty in the literature, and due to the wide variations in the levels of turbidity shown to cause either of these outcomes, it is impossible to quantify this effect.

Turbidity and resulting sedimentation may affect spawning eulachon in the project area. (Other listed fish would not be exposed to this effect because none spawn in portions of the project area downstream of activities likely to generate turbidity.) High levels of turbidity have the potential to smother eggs and block the emergence of larvae (Langness 2009 pers. comm.). There are no known eulachon spawning concentrations in portions of the project area likely to be exposed to elevated turbidity and sedimentation. Given the lack of precise spawning locations, it is assumed that spawning could potentially occur anywhere in the portions of the Columbia River and North Portland Harbor with water depths of 8 to 20 feet, and if spawning occurs in this area, it would likely be exposed to elevated turbidity. In other words, exposure could result from turbidity-generating activities at pier 7 in the Columbia River and throughout North Portland Harbor. Actual exposure is expected to be quite low, as high levels of turbidity would be limited to approximately 300 feet downstream of the discrete areas where debris removal would occur and would be restricted to a much smaller area for other in-water activities (Exhibit 5-55). This represents a minuscule proportion of the channel and an insignificant fraction of the total available spawning habitat immediately surrounding the affected area for many miles upstream and downstream.

Exposure to eulachon eggs or larvae would be limited to the overlap of (1) the incubation and emergence period, approximately from January through June, with (2) the 4-year in-water construction period. Other resident fish that utilize habitat in the project area for their full life cycle (e.g., sculpins, threespine sticklebacks, suckers, dace, and shiners) would also be exposed to turbidity and sedimentation. Exhibit 5-56 summarizes the effect of turbidity and sedimentation on various life functions of fish.

Exhibit 5-56. Summary of Effect of Turbidity and Sedimentation on Life Functions of Fish

Activity/ Timing ^a	Mortality ^b	Gill Damage ^c	Stress ^c	Avoidance	Migration Delay ^c	Foraging/ Predation ^d	Spawning ^e
Debris Removal 11/1–2/28	No	Not likely	Not likely	Likely (~300 ft, 4-6 hrs, ~4x/day)	Not likely	Likely	Likely (~300 feet)
Impact installation 9/15–4/15	No	Not likely	Not likely	Likely (25 ft, ~1 hr/day)	Not likely	Likely	Likely (~25 feet)
Vibratory installation year-round	No	Not likely	Not likely	Likely (25 ft, ≤24 hr/day)	Not likely	Likely	Likely (~25 feet)
Pile/cofferdam removal year-round	No	Not likely	Not likely	Likely (minimal, ≤24 hr/day)	Not likely	Likely	Likely (minimal)

Activity/ Timing ^a	Mortality ^b	Gill Damage ^c	Stress ^c	Avoidance	Migration Delay ^c	Foraging/ Predation ^d	Spawning ^e
Drilled shafts year-round	No	Not likely	Not likely	Not likely (contained)	Not likely	Likely	Not likely (contained)
Demolition year-round	No	Not likely	Not likely	Likely (minimal, ~8-10 hr/day)	Not likely	Likely	Likely (minimal)
Barges, shallow water year-round	No	Not likely	Not likely	Likely <300 feet	Not likely	Likely	Likely (<300 feet)

a All activities to occur within 4-year in-water constriction period.

Exhibit 5-57 summarizes the species and life stages of fish that could potentially be exposed to turbidity and sedimentation in the Columbia River and North Portland Harbor.

Exhibit 5-57. Fish Species Potentially Exposed to Project-generated Turbidity in the Columbia River and North Portland Harbor

			Life Sta	ige	
Species	Spawning	Incubation	Rearing	Juvenile Outmigration	Migrating/ Holding Adults
Chinook					
LCR ESU			Χ	X	X
UCR Spring-Run ESU			Χ	X	X
UWR ESU				X	X
SR Fall-Run ESU				Χ	X
SR Spring/Summer-Run ESU				X	X
Steelhead					
LCR DPS			Χ	X	X
MCR DPS				X	X
UWR DPS				X	X
UCR DPS				X	X
SR DPS				X	X
Sockeye					
SR ESU				X	X
Coho					
LCR ESU			Χ	Χ	X
Chum					
CR ESU			Χ	Χ	X
Bull Trout (exposure is discountable	ole due to extrer	nely low number	ers in project	area)	
CR DPS					X ^a
Green Sturgeon (exposure is disce	ountable due to	extremely low	numbers in p	roject area)	
Southern DPS					X ^a
Eulachon					
Southern DPS	X	X		X	X

b Turbidity would not reach levels known to cause mortality.

c Exposure unlikely due to avoidance, dilution, turbidity refugia, and limited extent and duration of effect.

d Effect likely but not quantifiable.

e Applies to eulachon only.

	Life Stage				
Species	Spawning	Incubation	Rearing	Juvenile Outmigration	Migrating/ Holding Adults
Lamprey species			Χ	Х	Х
Resident fish (e.g., sculpin, dace, threespine stickleback, sucker, shiner)	Х	Х	Х		

a Includes subadults.

Turbidity is not expected to have an effect on invertebrate distribution or abundance.

Summary of Effects to Aquatic Species

Bull trout and green sturgeon could potentially be exposed to turbidity effects, but due to extremely low numbers of these species in the very limited areas subject to elevated turbidity, exposure would be insignificant.

Adult and juvenile salmon and steelhead (Exhibit 5-57) are likely to be exposed to elevated turbidity, but not at levels likely to cause mortality, gill damage, stress, or migratory delay. Turbidity may reach levels that could cause temporary avoidance of the areas within the discrete mixing zones and timelines outlined in Exhibit 5-55 and 5-56. This is likely an adverse effect.

Adult and larval eulachon, as well as resident fish, are likely to be exposed to elevated turbidity in the same manner as described for salmon and steelhead. Additionally, turbidity and sedimentation may have adverse effects on spawning and potential spawning habitat, but these effects would be limited to discrete areas, representing a miniscule proportion of available spawning habitat. Turbidity is not expected to interfere with migration of larval eulachon, which do not have volitional movement.

The temporary turbidity that is likely to occur from project activities is not expected to reach a level that would impact sea lions. Sea lions in the Columbia River are known to habituate to very high levels of turbidity; for example, at the Bonneville Dam spillway, where they congregate to feed during maximum spring spill (Tennis 2010 pers. comm.). The level of turbidity resulting from project activities is not expected to interfere with sea lions' ability to navigate, respire, avoid predators, or find prey. Therefore, this element of the project is not likely to significantly impact sea lions.

Turbidity is not likely to measurably affect any life stages of lamprey.

5.3.3.5 Contaminated Sediments

State and federal databases have identified upland sites in the project area or immediate vicinity that are known or suspected to contain contaminated media (Parcel Insight 2009). Parcel Insight (2009) compiled information from all of the regulatory databases related to chemical contamination in the project area, including the federal Comprehensive Environmental Response, Compensation, and Liability Information System (CERCLIS) database, Oregon State Environmental Cleanup Site Information (ECSI) database, Oregon and Washington State Leaking Underground Storage Tank (LUST) database, and Oregon State Hazardous Materials (HAZMAT) database. DEQ suspects that four sites in the project area may contain contaminated sediments due to their proximity to the contaminated upland sites and due to available information about past activities on the sites (Parcel Insight 2009).

- Schooner Boat Works Pier 99 is a marine repair facility located on the south bank of North Portland Harbor, east of I-5. The facility appears in the ECSI and CERCLIS databases. Metals and petroleum products were detected in on-site soils. Groundwater and sediment at the site have not yet been analyzed. Considering the types of activities conducted at the site and the length of time that these activities occurred, other potential site contaminants may include: organotoxins, toxic metals (such as arsenic, lead, cadmium, chromium, mercury, and zinc), volatile organic compounds, semi-volatile organic compounds, and PCBs. Additionally, regulatory agencies have received complaints about this site releasing materials into the water (Parcel Insight 2009).
- Diversified Marine is a second marine repair facility located on the south bank of North Portland Harbor, west of the I-5 bridge. This facility also appears in the Oregon State HAZMAT and ESCI databases and in the federal CERCLIS database. As for Pier 99, regulatory agencies have received complaints about the Diversified Marine site releasing materials into the water. The record of Pollution Complaints and Spill Reports suggests that on-site activities could have contaminated the site soils and nearby sediments with any of a variety of contaminants used in boat building, maintenance, and repair. These contaminants may include paint chips, toxic metals (such as copper oxide, organotins, lead, cadmium, chromium, mercury, and zinc), petroleum constituents (such as benzene, toluene, ethylbenzene, and PAHs), and organic contaminants such as phthalates, pentachlorophenol, chlorinated solvents, and PCBs (Parcel Insight 2009).
- The site of a former landfill is located on Hayden Island near the Columbia River shoreline and to the west of I-5 at the current location of the Thunderbird Hotel. This unregulated landfill was located in a seasonal lake basin and probably operated between 1950 and 1970, after which it was covered with a 7- to 8-foot layer of clean fill. In 1989, an ARCO gas station that later opened on the eastern edge of the former landfill initiated a study and detected gasoline contamination in the groundwater. Borings also revealed a layer of landfill debris beneath clean fill. The DEQ LUST program (file #26-89-0149) requested a Corrective Action Plan from ARCO, leading to pump-and-treat remediation that began operating in August 1990. Groundwater samples from eight monitoring wells contained dissolved metals, which are most likely a result of leachate percolating through unknown solid wastes in the unsaturated zone (Parcel Insight 2009). Because there is a high connectivity between the groundwater and the Columbia River in this location, it is suspected that metals could be present in the river sediments immediately adjacent to the site.
- The former site of the Boise-Cascade Lumber Mill is located in Vancouver on the north shore of the Columbia River, about 1,500 feet to the west of the I-5 bridge and to the west of the Red Lion Hotel. Based on the industrial history and type of activities conducted on the site, it is possible that these contaminants may have impacted nearby sediments in the Columbia River. However, the USACE performed in-water sediment sampling near the site, but did not detect contaminated sediments (USACE 2008, 2009).

The project would implement several measures to prevent the mobilization of contaminated sediments in the project area. First, the project would complete a Phase I Environmental Site Assessment or each acquired property that could reasonably contain contaminated materials. The Phase I Environmental Site Assessment may identify possible contamination based on the site history, a visual inspection of the site, and a search of federal and state databases of known or suspected contamination sites. If there is evidence of contamination, a Phase II Environmental Site Assessment may be performed to pinpoint the location of the contaminated sediments as well as to measure the extent and concentration of the contaminants. The Phase II Environmental Site Assessment would also identify the specific areas recommended for remedial action.

The project would implement BMPs to ensure that the project either: 1) avoids areas of contaminated sediment or 2) enables responsible parties to initiate cleanup activities for contaminated sediments occurring within the project construction areas. The exact BMPs are not yet determined, but the contractor would be required to develop mitigation and remediation measures in accordance with ODOT and WSDOT standard specifications and all state and federal regulations. The plan would also comply with all regulatory criteria related to contaminated sediments. There would be coordination with regulatory agencies such as DEQ and Ecology on the assessment of site conditions and the cleanup of contaminated sediments. If contaminated sediments are removed from the site, they would be disposed of at a permitted upland disposal site.

Because the project would identify the locations of contaminated sediments and use BMPs to ensure that they do not become mobilized, there is little risk that aquatic species would be exposed to contaminated sediments. This aspect of the project is not likely to measurably affect any fish species.

5.3.4 Avian Predation

Project-related in-water and overwater structures may have an effect on avian predation in the CRC project area. Such structures may include the temporary work platforms/bridges, tower cranes, oscillator support platforms, barges, and cofferdams, as well as the permanent new bridge spans.

Avian predation is known to be a factor that limits salmon recovery in the Columbia River basin (NMFS 2008e). Throughout the basin, birds congregate near man-made structures and eat large numbers of migrating juvenile salmonids (Ruggerone 1986, Roby et al. 2003, and Collis et al. 2002 cited in NMFS 2008e). Basin wide, avian predation is high enough to constitute a substantial portion of the mortality rate of several runs of salmon and steelhead (Roby et al. 2003 cited in NMFS 2008e). Predation rates are particularly high in impoundments upstream of dams, dam bypass systems, and dredge spoil islands (NMFS 2008e). Additionally, local environmental factors may exacerbate avian predation. In particular, mainstem dams in the lower Columbia detain suspended sediments, a condition that has increased water clarity, potentially enhancing the foraging success of predaceous birds (NMFS 2008e).

The effects of overwater structures on interactions between salmonids and avian predators are widely recognized but have not been the subject of extensive study (Carrasquero 2001). In a 2001 literature review Carrasquero (2001) determined that there is no quantitative or qualitative evidence that docks, piers, boathouses, or floats either increase or decrease predation on juvenile salmonids. Additionally, the review found no studies related to predator-caused mortality specifically associated with overwater structures. Caspian terns, double-crested cormorants, and various gull species are the principal avian predators in the Columbia River basin (NMFS 2000 cited in NMFS 2008e). Populations in the basin have increased as a result of nesting and feeding habitats caused by the creation of dredge spoil islands, reservoir impoundments, and tailrace bypass outfalls (Roby et al. 2003). However, no studies have demonstrated the use of overwater structures by predaceous birds (Carrasquero 2001).

The overwater structures in the CRC project area are not likely to attract large concentrations of avian predators as do such features as nesting islands, impoundments, or tailraces. Nevertheless, because avian predators are known to congregate on overwater structures and because the project would increase the number of available perches, it is possible that the avian predation rates could increase to some extent within the project area. Specifically, the new bridges could create a permanent increase in the number of perches available. Additionally, the work platforms/bridges,

tower cranes, oscillator support platforms, and barges would temporarily increase the number of perches available in the Columbia River and North Portland Harbor. Presumably, avian predation may occur during the overlap of: 1) when overwater structures are present in the project area and 2) when juvenile fish are present in the project area; however, it is impossible to quantify how many individual fish would be affected.

5.3.5 Effects to Aquatic Habitat

5.3.5.1 Shallow-water Habitat

The project would have temporary effects on shallow-water habitat in the Columbia River and North Portland Harbor. Temporary impacts to shallow water include: in-water and overwater structures (work platforms, work barges, tower cranes, oscillator support piles, cofferdams, and barges), turbidity, and elevated underwater noise. Section 4.1.2.1 outlines the role of shallow-water habitat in the life history of fish.

The following ESA-listed and SOI and life stages of fish could be exposed to these effects:

- Holding, feeding, and migration habitat for juveniles and holding and migration habitat
 for adults in several ESUs/DPSs: LCR coho; CR chum; SR sockeye; LCR, MCR, UCR,
 and SR steelhead; and LCR, UCR spring-run, SR fall-run, and SR spring/summer-run
 Chinook.
- Rearing habitat for juvenile Chinook (LCR, UCR spring-run, and UWR), LCR coho, CR chum, and LCR steelhead.
- Adult bull trout migration and holding habitat. Because of the extremely low numbers of bull trout in this portion of the project area, risk of exposure to this effect is discountable.
- Adult and subadult green sturgeon feeding and migration habitat. Because of the
 extremely low numbers of green sturgeon in this portion of the project area, risk of
 exposure to this effect is discountable.
- All life stages of white sturgeon.
- Adult and larval eulachon spawning and migration habitat.
- Lamprey ammocoetes may be present in the substrate within the project area. Project work would remove or disturb substrate that may contain ammocoetes. Therefore, ammocoetes may be injured or killed. Potential project impacts to this life stage should not be discounted, but because abundance and distribution data are so limited, impacts cannot be quantified at this time. Pacific lamprey are discussed in more detail in Appendix A.
- Impacts to shallow-water habitat would have no effect on sea lions.

Since shallow-water impacts would occur continually throughout the 4-year in-water construction period, as many as four migration cycles of salmon, steelhead, and eulachon could be exposed to these effects. Resident fish may be exposed to these effects year-round during the in-water construction period.

All of these species and life stages may use shallow-water habitat at some point during their presence in the project area. Of these life stages, rearing juvenile salmonids and subyearling migrant salmonids (CR chum and LCR Chinook) may be closely dependent on shallow-water habitat, and therefore are more vulnerable to these effects.

This section outlines the project's short-term effects on fish in shallow-water portions of the CRC project area. These effects include physical loss of habitat, increase in the area of overwater coverage, turbidity, and underwater noise.

Physical Loss of Shallow-Water Habitat

Exhibit 5-58 and Exhibit 5-59 quantify the area affected by short-term physical loss of shallow-water habitat. The project would lead to temporary physical loss of approximately 20,700 square feet of shallow-water habitat. Project elements responsible for temporary physical loss include the footprint of the numerous temporary piles associated with in-water work platforms, work bridges, tower cranes, oscillator support piles, cofferdams, and barge moorings in the Columbia River and North Portland Harbor. Note that all North Portland Harbor impacts are in shallow water.

Exhibit 5-58. Physical Impacts to Shallow-water Habitat in the Columbia River

Structure	Area	Time in Water
Work Platforms – construction (P2 & 7)	728 sq. ft.	150 - 300 days each
Barge moorings – construction (P7)	25 sq. ft.	120 days each
Cofferdams – construction (P7) (about ¼ is in shallow water)	2,000 sq. ft.	240 days each
Barge moorings - demolition (existing Pier 10, 11)	200 sq. ft.	30 days each
Coffer dams – demolition (existing Pier 10, 11)	15,000 sq. ft.	45 days each
Total	17,753 sq. ft.	

Exhibit 5-59. Physical Impacts to Shallow-water Habitat in North Portland Harbor

Structure	Area	Time in Water	
Temporary			
Work Platforms – construction (9 locations)	400 to 710 sq. ft.	Up to 42 days each	
Oscillator Platforms (31 locations)	1,200 to 1,560 sq. ft.	Up to 34 days each	
Barge moorings – construction (8 locations)	318 to 678 sq. ft.	Up to 34 days each	
Total	1,970 to 2,940 sq. ft.		

The effect of physical loss of shallow-water habitat is described in detail in Section 4.1.2.1.

In the case of the CRC project, in-water portions of the structures would not pose a complete blockage to nearshore movement anywhere in the project area. Although these structures would cover potential rearing and nearshore migration areas, the habitat is not rare and is not of particularly high quality. These juveniles would still be able to use the abundant shallow-water habitat available for miles in either direction. Neither the permanent nor the temporary structures would force these juveniles into deeper water, and therefore pose no added risk of predation. Additionally, northern pikeminnow and walleye tend to avoid high-velocity areas during the spring juvenile salmonid outmigration (NMFS 2000; Gray and Rondorf 1986; Pribyl et al. 2004). The high velocities present in deep-water portions of the CRC project area may limit the potential for actual predation in deep-water areas.

Physical loss of shallow-water habitat would have only negligible effects on foraging, migration, and holding of salmonids that are of the yearling age class or older. These life functions are not dependent on shallow-water habitat for these age classes. Furthermore, the lost habitat is not of particularly high quality. There is abundant similar habitat immediately adjacent along the shorelines of the Columbia River and throughout North Portland Harbor. The lost habitat

represents only a small fraction of the remaining habitat available for miles in either direction. There would still be many acres of habitat for foraging, migrating, and holding.

Physical loss of shallow-water habitat would have only negligible effects on eulachon and green sturgeon for the same reason as above. Resident fish such as dace, threespine stickleback, redside shiners, suckers, and sculpin would be impacted by localized loss of shallow-water habitat, but as is noted above, the lost habitat is not of particularly high quality, and there is abundant similar habitat immediately adjacent along the shorelines of the Columbia River and throughout North Portland Harbor. The lost habitat represents only a small fraction of the remaining habitat available for miles in either direction.

Increase in Overwater Coverage

The project would place several temporary overwater structures in shallow water in the Columbia River and North Portland Harbor. Temporary overwater structures include temporary work platforms, work bridges, oscillator support platforms, and stationary barges. Exhibit 5-60 and Exhibit 5-61 quantify the area and duration of project-related overwater structures in the CRC project area.

Exhibit 5-60. Temporary Overwater Coverage in Shallow-water Habitat in the Columbia River

Structure Type	Area	Duration in Water (days)
Work bridges (P2, P7)	36,000 sq. ft.	150–300 days/pier complex
Barges for Demolition (Existing Piers 10 & 11)	14,350 sq. ft.	Varies up to 30 days/barge
Total Temporary Impact	50,350 sq. ft.	

Exhibit 5-61. Temporary Overwater Coverage in Shallow-water Habitat in North Portland Harbor

Structure Type	Area	Duration in Water
Temporary		
Work Bridges (8 locations)	29,640 sq. ft.	Up to 42 days each
Oscillator Support Platforms (31 locations)	27,900 sq. ft.	Up to 34 days each
Barges for Construction (31 locations)	64,164 sq. ft.	Up to 34 days each
Total Temporary Impact	108,164 sq. ft.	

Temporary structures would not all be present in the project area at the same time. The maximum amount of shade from temporary overwater structures in shallow water in the Columbia River would be no more than about 18,500 square feet at one time. In North Portland Harbor, the maximum amount of shade in shallow water at one time would be about 112,180 square feet.

Effects of overwater coverage on fish and fish habitat are discussed in Section 5.3.2.

Turbidity

The project would also temporarily degrade shallow-water habitat by creating turbidity. Turbidity would pose fairly limited impacts to shallow-water habitat, as the project would restrict the extent of turbidity to distances specified by regulatory permits (anticipated to be 300 feet). The turbidity may make discrete areas temporarily unavailable for foraging, rearing, holding and migration, but

only for short periods of time (as specified by the regulatory permits). Fish would be able to use the abundant, similar-quality shallow-water habitat outside of the areas subject to high turbidity.

Underwater Noise

Underwater noise would temporarily degrade shallow-water habitat, creating disturbance in the Columbia River from RM 101 to 118 and in North Portland Harbor 3.5 miles downstream of the project area and 1.9 miles upstream. In these areas, behavioral disturbance is likely to occur. Additionally, underwater noise is expected to cause significant, though temporary, effects to shallow-water habitat, making these areas unsuitable for foraging, rearing, and holding because fish entering this area would be killed or injured. Underwater noise may also create a temporary barrier to migration for both adults and juveniles in these areas during this time period (Caltrans 2009).

5.3.5.2 Deep-water Habitat

Deep-water habitat occurs only in the Columbia River portion of the project area. Aquatic SOI have mixed use of this deep-water habitat.

Fish Distribution in Deep-Water Habitat

Typically, yearling and subyearling migrant salmonids (CR chum and some LCR Chinook) are restricted to shallow-water habitat in the upper estuary (including the project area) (Carter et al. 2009); however, we cannot discount the possibility that some would occasional stray into the surface layer of deeper waters (Bottom et al. 2005). Larger juveniles commonly use deep-water portions of the navigation channel in high numbers during outmigration, taking advantage of higher velocities there (Carter et al. 2009).

Adult salmonids do not show any specific preference for deep-water habitat over shallow-water habitat (Bottom et al. 2005). While they generally migrate at mid-channel, they may be found at depths of 1 to 50 feet (NMFS 2005b). They commonly use deep-water portions of the project area for foraging and hold in low-velocity areas of deep-water habitat (such as behind bridge piers).

Eulachon adults and juveniles are known to forage at depths of greater than 50 to 600 feet and could be present in deep-water portions of the project area (Hay and McCarter 2000).

Adult and subadult green sturgeon use waters at a depth of 30 feet or less and also could be present in deep-water portions of the project area (73 FR 52084).

Effects to Fish in the CRC Project Area

Impacts to deep-water habitat would affect the following species and life stages of fish:

- Feeding, holding and migration habitat for juveniles and holding and migration habitat
 for adults of the following ESUs/DPSs: LCR coho; CR chum; SR sockeye; LCR, MCR,
 UCR, and SR steelhead; and LCR, UCR spring-run, SR fall-run, and SR spring/summerrun Chinook.
- Rearing habitat for juvenile Chinook (LCR, UCR spring-run, and UWR), LCR coho, LCR steelhead, and CR chum.
- Adult and subadult bull trout migration and holding habitat. (Because of the extremely low numbers of bull trout in this portion of the project area, risk of exposure to this effect is discountable).

- Adult and subadult green sturgeon feeding and migration habitat. (Because of the extremely low numbers of green sturgeon in this portion of the project area, risk of exposure to this effect is discountable).
- All life stages of white sturgeon.
- Adult and larval eulachon spawning and migration habitat.
- Lamprey ammocoetes may be present in the substrate within the project area. Project work would remove or disturb substrate that may contain ammocoetes. Therefore, ammocoetes may be injured or killed. Potential project impacts to larval, juvenile, or adult lamprey in deep water portions of the project area should not be discounted, but because data are lacking, impacts cannot be quantified at this time. Pacific lamprey are discussed in more detail in Appendix A.

The project would have temporary impacts on deep-water habitat in the Columbia River. These impacts include: physical loss of habitat, increase in overwater coverage, turbidity, and in-water noise.

Physical Loss of Deep-water Habitat

A summary of temporary physical impacts to deep-water habitat in the Columbia River is in Exhibit 5-62.

Exhibit 5-62. Physical Impacts to Deep-water Habitat in the Columbia River

Structure Type		Area	Time in Water
Work platforms – construction (P3 – P6) ^a		3,870 sq. ft.	150-300 days each
Tower cranes – construction (P2 – P7)		603 sq. ft.	350 days/crane
Barge moorings – construction (P2 – P6)		226 sq. ft.	120 days /pier complex
Barge moorings - demolition (existing Piers $2-9$)		754 sq. ft.	40 days/pier complex
Coffer dams – demolition (existing Piers 2 – 9)		52,500 sq. ft.	~317 days
	Total	57,953 sq. ft.	

P = Pier Complex

The structures shown in Exhibit 5-62 would not all be in place at the same time. During construction, temporary structures would occupy no more than 1,080 square feet of substrate in deep water at one time. During demolition, temporary structures would occupy no more than 15,100 square feet of substrate in deep water at one time.

Although there would be a temporary net physical loss of deep-water habitat, this is not expected to have a significant impact on fish. None of the fish SOI are particularly dependent on deep-water habitat. The lost habitat is not rare or of particularly high quality, and there is abundant similar habitat in immediately adjacent areas of the Columbia River and for many miles both upstream and downstream. The lost habitat would represent a very small fraction (far less than 1 percent) of the remaining habitat available. Additionally, the in-water portions of the permanent and temporary in-water structures would occupy no more than about 1 percent of the width of the Columbia River. Therefore, the structures would not pose a physical barrier to migration. Due to the small size of the impact relative to the remaining habitat available, this effect would be insignificant.

Increase in Overwater Coverage

The project would place several temporary overwater structures in deep-water portions of the Columbia River including work platforms, tower cranes, and stationary barges. Exhibit 5-63 quantifies the area and duration of temporary overwater structures in deep-water portions of the project area, showing that there would be a net temporary increase in shade in the project area.

Exhibit 5-63. Overwater Coverage in Deep-water Habitat in the Columbia River

Structure Type	Area	Duration in Water (days)
Work Platforms for Drilling Shafts (P 3 – 6)a	112,000 sq. ft.	260 – 315 / platform
Tower Cranes (P 2 – 7)	2,400 sq. ft.	150 - 200 /crane
Barges for Construction (P 3 – 6)	106,432 sq. ft.	300 - 480 / complex
Barges for Demolition (Existing Piers 2 – 9)	14,350 sq. ft.	~320
Total	235,182 sq. ft.	

P = Pier Complex

General effects of overwater coverage on fish are described in detail in Section 5.3.2.1. In summary, overwater coverage creates dense shade that may attract predators and may cause visual disorientation to juvenile fish, which may in turn result in delayed migration and increased vulnerability to predators. Of the juvenile fish that use the project area, rearing juveniles and subyearling-migrant salmonids are highly dependent on shallow-water habitat and therefore are less vulnerable to these effects in deep water. However, as these individuals are not restricted to the nearshore (Bottom et al. 2005), they may stray into deeper water, and there is a small chance of exposure to these effects. Larger juveniles of the yearling age class or older commonly use deep-water habitat during migration, and therefore are likely to be exposed to these effects.

The temporary structures would not create a swath of dense shade completely spanning deep-water habitat. Therefore, even if these structures were to create a shadow line that juvenile salmonids avoid crossing during daylight hours, juveniles could simply circumvent the shadow, resulting in no measurable delay to migration. Nighttime migration would be unaffected. Larval eulachon do not have volitional movement and are therefore not subject to visual disorientation or migration delays.

The increase in the shade footprint increases the amount of suitable habitat for predators and therefore could presumably increase the number of predators in this portion of the project area. This could potentially cause a temporary and/or permanent increase in predation rates on juveniles, although it is not possible to quantify the extent of this effect. All of the juveniles fish SOI that use this portion of the project area could potentially be exposed to this effect, although it is impossible to quantify the number of fish that could be affected.

Turbidity

The project would temporarily degrade deep-water habitat by creating turbidity. Exhibit 5-64 summarizes the activities likely to generate turbidity in deep water.

Exhibit 5-64. Activities Likely to Generate Turbidity in Deep Water in the Columbia River

Activity	Timing ^a	Location ^b	Likely Extent of Turbidity	Duration of Effect (hr/day)	Number of Work Days
Install temporary piles, impact methods	9/15 – 4/15	Adjacent to P 2 – 7	~25 feet	0.66	~138
Install temporary piles, vibratory methods	Year round	Adjacent to P 2 – 7	~25 feet	up to 24	continually over ~928
Remove temporary piles, direct pull or vibratory	Year round	Adjacent to P 2 – 7	Minimal	up to 24	continually over ~928
Install steel casings to drill permanent shafts – vibratory hammer, oscillator, or rotator	Year round	Adjacent to P 2 – 7	~25 feet	8 – 10	60 – 80 days / pier complex
Drill and excavate permanent shafts	Year round	Adjacent to P 2 – 7	None (contained)	N/A	60 – 80 days / pier complex
Demolish existing Columbia River bridge piers (includes installation and demolition of cofferdams)	Year round	Existing Piers 2 – 9	Minimal	8 – 10	~320

a All activities likely to take place throughout the 4-year in-water construction period.

General effects of turbidity are described in detail in Section 5.3.3.4. In summary, turbidity would pose fairly limited impacts to deep-water habitat, as the project would restrict the extent of turbidity to distances specified by regulatory permits. It is anticipated that the regulatory permits would specify a mixing zone of no more than 300 feet. In actuality, many of the activities would restrict the turbidity plume to far shorter distances (Exhibit 5-56). Permits would also restrict the duration of each turbidity plume to approximately 4 to 6 hours at a time.

The turbidity plumes may make discrete areas temporarily unavailable for foraging, holding and migration, but only for short periods of time (as specified by the regulatory permits). Due to the high dilution capacity of the Columbia River, turbidity plumes are expected to disperse relatively quickly and within a short distance of the source. Due to the large size of the water body relative to the small size of the turbidity plume, fish are not likely to become trapped in turbid water. Fish would be able to use the abundant turbidity refugia in deep-water habitat outside of the areas subjected to high turbidity. Both adult and juvenile fish could be exposed to this effect.

Underwater Noise

Underwater noise and vibration would have the same effects on deep-water habitat as described for shallow-water habitat above.

5.4 Effects to Terrestrial Resources

5.4.1 Terrestrial Habitat

Terrestrial habitat is likely to be temporarily impacted for construction access along highway right-of-way. Terrestrial habitat that would be impacted by project construction is likely to be of low quality for terrestrial wildlife because it is likely to be within existing highway right-of-way and/or degraded by proximity to existing urban development. Erosion could occur from construction activities. Appropriate avoidance and minimization methods (silt fencing, no-work

b P = Pier Complex

zones, erosion control BMPs) would reduce potential impacts to the riparian areas. Riparian habitat along the Oregon and Washington banks of the Columbia River would be impacted by construction activities including any of the following: deconstruction of existing structures, construction of new bridge elements, access to work areas, workers on foot, vehicles, survey crews, and other related construction presence along the banks. Riparian vegetation, including herbaceous plants, shrubs, and small trees, may be trampled or removed. Because the condition of the riparian area is currently fairly degraded due to the urban location, construction activities may further compromise riparian function and ability to provide habitat features for terrestrial species including mammals and migratory birds. Mitigation measures would address impacts to the riparian community and are likely to result in a net improvement in riparian function relative to current conditions.

5.4.2 Riparian Habitat

In North Portland Harbor and the Columbia River, effects to riparian habitat would be negligible, as there is very little functioning riparian vegetation in the project area. The project would revegetate disturbed shoreline areas, resulting in a net benefit to riparian habitat in the long term. It has not yet been determined exactly where replanting would take place. However, it is anticipated that replanting would occur on or adjacent to the current sites of the trees where practicable. In any case, the number, type, and size of the replanted trees would be selected to comply with standards outlined in the City of Portland and City of Vancouver tree ordinances.

In Oregon, the project would remove three deciduous trees, all with trunks less than 1 foot in diameter, from the riparian zone on the south bank of the Columbia River. The project would also remove two deciduous ornamental trees from the riparian zone adjacent to North Portland Harbor. These trees are located in a landscaped setting and have trunks of approximately 1 foot in diameter. In Washington, 10 trees with trunks less than 1 foot in diameter would be removed from the riparian zone on the north shore of the Columbia River.

In general, removal of trees from riparian areas results in a reduction of shade in the water column and a concurrent increase in water temperature. However, in the case of the CRC project, only approximately 15 trees would be removed from the Columbia River/North Portland Harbor riparian area. This represents an extremely small amount of shaded water (less than 10,000 square feet, patchily distributed among at least three locations) relative to the thousands of acres of unshaded water located immediately adjacent to the area from which trees would be removed. Because of the small size of the shaded area relative to the large volume of water and because of the high current velocity in these water bodies, it is unlikely that these 15 riparian trees create enough shade to measurably decrease water temperatures in the water column. Thus, the loss of these trees is expected to cause only negligible effects to water temperature, if any.

Additionally, removal of trees from riparian areas may reduce the potential for large woody debris recruitment in a watershed over the long term. However, given the large size of the lower Columbia system and the thousands of remaining riparian trees in this area, removal of 15 trees would not measurably decrease the potential for long-term large woody debris recruitment in the project area or in the lower Columbia system overall.

There would be no excavation, vegetation clearing, or removal of trees from the Columbia Slough riparian area. Therefore, the project would have no effect on Columbia Slough riparian habitat.

The project would not remove any trees from the Burnt Bridge Creek riparian area. Temporary impacts from construction may include some clearing of, or temporary storage in, this area. However, after construction is complete, exposed soil would be revegetated with native vegetation, resulting in no long-term impact.

5.4.2.1 LPA with Highway Phasing

Should the project improvements at SR 500 be deferred under the LPA with highway phasing option, temporary riparian impacts near Burnt Bridge Creek would also be deferred.

5.4.3 Threatened, Endangered, and Proposed Species

Short-term effects to listed and proposed aquatic species are discussed throughout Section 5. No federally listed terrestrial species are known to occur in the project area. Effects to state listed terrestrial species are discussed in the sections below.

5.4.4 Species of Interest

Terrestrial resources, such as protected birds and other SOI, would be impacted because construction activity would create noise disturbance and disruption of potential nesting and/or roosting habitat as the bridge structures are deconstructed or retrofitted. Migratory bird nesting and roosting habitat (e.g., the structures of the existing bridge) would be permanently removed. Construction activities conducted during nesting season could cause excessive disturbance through noise and physical displacement of bridge structures, resulting in nest failure and/or the need to remove active nests.

Although the existing bridge does not provide ideal roosting habitat for bats, several bat species that may pass near the existing bridge and use it for temporary roosting may be affected by construction disturbance. Short-term effects to raccoons, bats, reptiles, and other terrestrial wildlife could result from high levels of noise, clearing/alteration of vegetation, potential impacts to water quality, and other disturbances that could affect breeding, foraging, and dispersal.

5.4.5 Acoustic Impacts to Terrestrial Species

Construction activities conducted during nesting season could cause excessive disturbance through noise and physical displacement of bridge structures, resulting in nest failure and/or the need to remove active nests. Peregrine falcons are known to use the existing bridge and would be directly impacted by noise disturbance if construction activities occurred during nesting and fledging season. Although no other state or federally protected birds (e.g., bald eagles) are known to nest in or near the project area, compliance with applicable regulations such as the federal MBTA would occur. In addition, construction activities would need to comply with city ordinances on noise production.

5.4.6 Wildlife Passage

Given the highly developed character of the project area, wildlife passage is degraded and severely limited in the project area. Passage is most likely to occur along river banks (particularly for waterfowl) and between vegetated areas that offer some cover. Wildlife passage may be even further impaired during construction as construction equipment is mobilized, stored, and used, and as construction activities occur on or near river banks. Effects to wildlife could include altered behavior to avoid construction activities (e.g., moving through more developed areas), and could increase the risks of human/wildlife conflicts and wildlife mortality.

5.5 Effects to Botanical Resources

Temporary impacts to vegetation are anticipated (see discussion above relevant to terrestrial habitat). No listed or otherwise rare plants are known to occur in the project area, and are therefore not expected to be impacted.

6. Proposed Mitigation for Adverse Effects

6.1 Introduction

Mitigation for impacts to aquatic, terrestrial, and botanical resources may include BMPs, conservation measures, and avoidance and minimization measures. Standard construction BMPs and conservation measures would be implemented to avoid or minimize impacts to ecosystem resources from construction activities. Discussions with agencies from both Washington and Oregon are ongoing to determine specific mitigation measures.

6.2 Proposed Mitigation for Long-term Adverse Effects

6.2.1 Aquatic Resources

Impacts to listed salmonids must be addressed through avoidance and minimization measures. The LPA would impact listed fish species through the presence of large piers in the river that could provide habitat for piscivorous fish and birds, lead to a physical loss of substrate, increase the amount of overwater coverage, affect local flow patterns, and impact streambed conditions through sediment deposition. Potential measures to address these impacts include discouraging piscivorous fish and other predator use of piers, promoting aquatic habitat conservation efforts, and ensuring adherence to water quality standards. Riparian fringe habitat may also be altered during construction and as a result of new bridge design. Revegetation of riparian areas and limited use of riprap would be employed to limit long-term effects. Bio-engineered bank protection may also be considered to address impacts to riparian areas and vegetation.

Impact avoidance and minimization are also addressed through project design alternatives that were considered but not advanced due to impacts to ecosystems and other resources. Certain design alternatives have also been modified to reduce impacts to resources. Examples of design alternatives that were not advanced include a dug tunnel between Vancouver and Portland; significant damming of the Columbia River during project construction; and placement of a park and ride facility on Cold Canyon (northwest of Burnt Bridge Creek). Examples of design alternatives that have been modified include minimization of piers in the river from 21 to 12, reducing the number of bridge spans in the Columbia River from 3 to 2, providing a high level of stormwater treatment, and avoiding Vanport wetlands and the Delta Park area.

The project would be required to offset impacts to aquatic habitat by performing compensatory mitigation as required by Section 404 of the Clean Water Act, a WDFW HPA, Oregon Removal/Fill law, and other regulations. Mitigation under City of Portland and City of Vancouver requirements has not yet been determined, but would also address impacts to aquatic habitat. The project proposes two mitigation sites: the Lower Hood River Powerdale Corridor Off-Channel Wetland Reconnection and the Lewis River Confluence Side Channel Restoration.

6.2.1.1 Lower Hood River Powerdale Corridor Off-Channel Wetland Reconnection

The Lower Hood River Powerdale Corridor Off-Channel Wetland Reconnection site contains the following species of fish: LCR Chinook, LCR steelhead, LCR coho, the Columbia River DPS of

bull trout, native resident fish such as sculpins, sticklebacks, and suckers, and potentially Pacific lamprey. This action would remove an earthen berm that currently isolates a historic side channel of the Hood River from the main stem. Long-term beneficial effects of the action would include:

- Increased area of spawning and rearing habitat for salmon and steelhead.
- Native woody vegetation planted throughout the site would provide food chain support, cover, and shade. This would improve foraging, rearing, holding, and migrating habitat for adult and juvenile salmonids.
- Creation of high-flow refuges, improvement of base flows, attenuation of peak flows, and improvements to water quality would result in enhancement of rearing habitat for juvenile salmon and steelhead.
- Placement of large woody debris would create habitat complexity, improving rearing and holding conditions for salmonids.

6.2.1.2 Lewis River Confluence Side-Channel Restoration

The Lewis River Confluence Side-Channel Restoration site contains the following species of fish: LCR Chinook, LCR coho, LCR steelhead, and potentially CR chum, eulachon, green sturgeon, Pacific lamprey, and bull trout. This action would connect a historic side channel with the Lewis River. Long-term beneficial effects are likely to include:

- Increased area of spawning and rearing habitat for salmon and steelhead.
- Native woody vegetation planted throughout the site would provide food chain support, cover, and shade. This would improve foraging, rearing, holding, and migrating habitat for adult and juvenile salmonids.
- Creation of high flow refuges, improvement of base flows, attenuation of peak flows, and improvements to water quality would result in enhancement of rearing habitat for juvenile salmon and steelhead.
- Placement of large woody debris would create habitat complexity, improving rearing and holding conditions for salmonids.

6.2.2 Terrestrial Resources

In general, long-term impacts to terrestrial resources are fairly minimal (see Exhibit 4-18) and would not require extensive mitigation. Long-term impacts to terrestrial resources would be addressed through avoidance and minimization measures, replanting vegetation, and addressing habitat modification for protected birds.

Native migratory birds (e.g., swallows) are not known to consistently utilize the existing bridge structures for nesting or other life stages. Current habitat conditions for migratory birds in the project area, especially along the river banks, are fairly poor and are dominated by urban built environment, with ornamental shrubs and trees providing habitat structure. Opportunities to replant riparian vegetation and to incorporate shrub and tree plantings with improved habitat structure in the project area to improve natural habitat conditions would be identified through ongoing discussions with the regulatory agencies.

Riparian habitat in the project area on both the Oregon and Washington banks is fairly degraded and provides limited habitat for terrestrial wildlife for passage, cover, breeding, feeding, and dispersal. To address the current condition of much of the riparian vegetative community in the project area, as well as the impacts to riparian vegetation from project construction, opportunities

to incorporate the improvement of riparian function and habitat, either on-site or off-site within the basin, would be addressed through ongoing discussions with the regulatory agencies. Mitigation for effects to riparian habitat would meet all applicable local, state, and federal requirements.

Impacts to wildlife passage would be addressed through avoidance and minimization measures. Placement of new structures or replacement of existing structures along the I-5 alignment creates obstructions to movement of wildlife. This is particularly true along riparian zones. Although little intact riparian habitat suitable for passage is currently present along the Columbia River and North Portland Harbor, placement of obstructions would create an additional passage obstacle for several decades, thereby limiting potential future connectivity projects. Efforts to improve riparian conditions through replanting riparian vegetation would be considered and discussed with the regulatory agencies, and would potentially be achieved through local jurisdiction requirements.

6.2.3 Botanical Resources

No long-term impacts requiring mitigation are anticipated for botanical resources. No sensitive, listed, or otherwise rare plant species are known to occur in the project area. Vegetation removal, including riparian vegetation, would be temporary and these areas would be replanted (Section 6.3).

6.3 Proposed Mitigation for Adverse Effects during Construction

6.3.1 Aquatic Resources

The LPA would impact listed fish species through in-water work that could result in increased turbidity and suspended sediments, underwater noise, temporary localized dewatering, potential contaminant spills, loss of substrate, increase in overwater coverage, increase in artificial lighting, increased predation by birds and fish, and an increase in hydraulic shadowing. Avoidance and minimization measures to address these impacts would apply to all phases of construction. Impact avoidance would be addressed to the extent possible by redesigning project components with adverse impacts. Impact minimization would be addressed by implementing BMPs (e.g., sediment and erosion control, no-work zones, appropriate flagging and fencing), and using cofferdams around in-water work sites. Measures to minimize turbidity would be implemented any time that work on the streambed occurs. Monitoring would likely be required to assess impacts to fish from in-water work. Avoidance, minimization, and conservation measures are discussed in more detail below.

6.4 Summary of Avoidance, Minimization, and Conservation Measures

6.4.1 General Measures and Conditions

• A biologist shall re-evaluate the project for changes in design and evaluation methods not previously employed in the BA to assess potential impacts associated with those changes, as well as the status and location of listed species, every 6 months until project construction is completed. Re-initiation of consultation with the Services is required if new information reveals project effects that may affect listed species or critical habitat in a manner or to an extent not previously considered. Re-initiation of consultation is also

- required if the identified action is modified in a manner that causes an effect to species that was not considered in the BA or if a new species is listed or critical habitat is designated that may be affected by the action.
- All work shall be performed according to the requirements and conditions of the
 regulatory permits issued by federal, state, and local governments. Seasonal restrictions,
 e.g., work windows, would be applied to the project to avoid or minimize potential
 impacts to listed or proposed species based on agreement with, and the regulatory permits
 issued by DSL, WDFW, and USACE in consultation with ODFW, USFWS, and NMFS.
- Drilled shafts would be installed while water is still in the cofferdam. The drilled shaft casing would function to contain and isolate the work. Cofferdams would be installed to minimize fish entrapment. Sheet piles would be installed from upstream to downstream, lowering the sheet piles slowly until contact with the substrate. When cofferdams are used, fish salvage would be conducted according to protocol approved by ODFW, WDFW, and NMFS (see Appendix E of the BA) (CRC 2010).
- Contractor shall provide a qualified fishery biologist to conduct and supervise fish
 capture and release activity as to minimize risk of injury to fish, in accordance with
 ODOT Standard Specification 00290.31(i) or its equivalent; and/or the 2009 WSDOT
 Fish Exclusion Protocols and Standards, or its equivalent.
- The contractor shall prepare a Water Quality Sampling Plan for conducting water quality monitoring for all projects occurring in-water in accordance with the specific conditions issued in the Oregon and Washington 401 Water Quality Certifications. The Plan shall identify a sampling methodology as well as method of implementation to be reviewed and approved by the engineer. If, in the future, a standard water quality monitoring plan is adopted by ODOT and/or WSDOT, this plan, with the agreement of NMFS and USFWS, may replace the contractor plan.
- State DOT policy and construction administration practice in Oregon and Washington is
 to have a DOT inspector on site during construction. The role of the inspector would
 ensure contract and permit requirements. ODOT/WSDOT environmental staff would
 provide guidance and instructions to the onsite inspector to ensure the inspector is aware
 of permit requirements.
- If in-water dredging is required outside of a cofferdam, a clamshell bucket shall be used. Dredged material shall be disposed of in accordance with relevant permits and approvals.
- Piles that are not in an active construction area and are in place 6 months or longer would have cones or other anti-perching devices installed to discourage perching by piscivorous birds.
- All pumps must employ a fish screen that meets the following specifications:
 - An automated cleaning device with a minimum effective surface area of 2.5 square feet per cubic foot per second, and a nominal maximum approach velocity of 0.4 foot per second, or no automated cleaning device, a minimum effective surface area of 1 square foot per cubic foot per second, and a nominal maximum approach rate of 0.2 foot per second; and
 - o a round or square screen mesh that is no larger than 2.38 millimeters (0.094") in the narrow dimension, or any other shape that is no larger than 1.75 millimeters (0.069") in the narrow dimension; and
- Each fish screen must be installed, operated, and maintained according to NMFS fish screen criteria.

6.4.2 Spill Prevention/Pollution Control

- The contractor shall prepare a Spill Prevention, Control, and Countermeasures (SPCC) Plan prior to beginning construction. The SPCC Plan shall identify the appropriate spill containment materials; as well as the method of implementation. All elements of the SPCC Plan would be available at the project site at all times. For additional detail, consult ODOT Standard Specification 00290.00 to 00290.90 and/or WSDOT Standard Specification 1-07.15(1). For transit construction in Oregon, consult TriMet Standard Specification 01450{1.04}).
- The contractor would designate at least one employee as the erosion and spill control (ESC) lead. The ESC lead would be responsible for the implementation of the SPCC Plan. The contractor shall meet the requirements of and follow the process described in ODOT Standard Specifications 00290.00 through 00290.30 and/or WSDOT Standard Specification 8-01.3(1)B. The ESC lead shall be listed on the Emergency Contact List as part of ODOT Standard Specification 00290.20(g) and/or WSDOT Standard Specification 1-07.15(1).
- All equipment to be used for construction activities shall be cleaned and inspected prior to arriving at the project site, to ensure no potentially hazardous materials are exposed, no leaks are present, and the equipment is functioning properly. Identify equipment that would be used below OHW. Outline daily inspection and cleanup procedures that would insure that identified equipment is free of all external petroleum-based products. Should a leak be detected on heavy equipment used for the project, the equipment shall be immediately removed from the area and not used again until adequately repaired. Where off-site repair is not practicable, the implemented SPCC Plan would prevent and/or contain accidental spills in the work/repair area to insure no contaminants escape containment to surface waters and cause a violation of applicable water quality standards.
- Operation of construction equipment used for project activities shall occur from on top of floating barge or work decks, existing roads or the streambank (above OHW). Any equipment operating in the water shall use only vegetable-based oils in hydraulic lines.
- All stationary power equipment or storage facilities shall have suitable containment
 measures outlined in the SPCC Plan to prevent and/or contain accidental spills to insure
 no contaminants escape containment to surface waters and cause a violation of applicable
 water quality standards.
- Process water generated on site from construction, demolition or washing activities
 would be contained and treated to meet applicable water quality standards before entering
 or re-entering surface waters.
- No paving, chip sealing, or stripe painting would occur during periods of rain or wet weather.
- For projects involving concrete, the implemented SPCC Plan shall establish a concrete truck chute cleanout area to properly contain wet concrete as part of ODOT Standard Specification 00290.30(a)1 and/or WSDOT Standard Specification 1-07.15(1).

6.4.3 Site Erosion/Sediment Control

• The contractor shall prepare a Temporary Erosion and Sediment Control (TESC) Plan and a Source Control Plan and implemented for the project requiring clearing, vegetation removal, grading, ditching, filling, embankment compaction, or excavation. The BMPs in the plans would be used to control sediments from all vegetation removal or ground-disturbing activities. The engineer may require additional temporary control measures

beyond the approved TESC Plan if it appears pollution or erosion may result from weather, nature of the materials or progress on the work. For additional detail, consult ODOT Standard Specifications 00280.00 to 00280.90 and/or WSDOT Standard Specification 1-07.15. For transit construction, consult TriMet Standard Specification 02276.

- As part of the TESC Plan, contractor shall delineate clearing limits with orange barrier fencing wherever clearing is proposed in or adjacent to a stream/wetland or its buffer and install perimeter protection/silt fence as needed to protect surface waters and other critical areas. Location would be specified in the field, based upon site conditions and the TESC Plan. For additional silt fence detail, consult ODOT Standard Specification 00280.16(c) and/or WSDOT Standard Specification 8-01.3(9)A.
- The contractor shall identify at least one employee as the ESC lead at preconstruction discussions and the TESC Plan. The contractor shall meet the requirements of and follow the process described in ODOT Standard Specifications Section 00280.30 and/or WSDOT Standard Specification 8-01.3(1)B. The ESC lead shall be listed on the Emergency Contact List as part of ODOT Standard Specification 00290.20(g) and/or WSDOT Standard Specification 1-05.13(1). The ESC lead would also be responsible for ensuring compliance with all local, state, and federal erosion and sediment control requirements.
- All TESC measures shall be inspected on a weekly basis. Contractor shall follow
 maintenance and repair as described in ODOT Standard Specifications 00280.60 to
 00280.70 and/or WSDOT Standard Specification 8-01.3(15). Inspect erosion control
 measures immediately after each rainfall, and at least daily during for precipitation events
 of more than 0.5 inches in a 24-hour period.
- For landward construction and demolition, project staging and material storage areas shall be located a minimum of 150 feet from surface waters, in currently developed areas such as parking lots or managed fields, unless a site visit by an ODOT/WSDOT biologist determines the topographic features or other site characteristics allow for site use closer to the edge of surface waters. Excavation activities (dredging not included) shall be accomplished in the dry. All surface water flowing towards the excavation shall be diverted through utilization of cofferdams and/or berms. Cofferdams and berms must be constructed of sandbags, clean rock, steel sheeting, or other non-erodible material.
- Bank shaping shall be limited to the extent as shown on the approved grading plans. Minor adjustments made in the field would occur only after engineer's review and approval. Bio-degradable erosion control blankets would be installed on areas of ground-disturbing activities on steep slopes (1V:3H or steeper) that are susceptible to erosion and within 150 feet of surface waters. Areas of ground-disturbing activities that do not fit the above criteria shall implement erosion control measures as identified in the approved TESC Plan. For additional erosion control blanket detail, consult ODOT Standard Specification 00280.14(e) and/or WSDOT Standard Specification 9-14.5(2)A.
- Erodible materials (material capable of being displaced and transported by rain, wind or surface water runoff) that are temporarily stored or stockpiled for use in project activities shall be covered to prevent sediments from being washed from the storage area to surface waters. Temporary storage or stockpiles must follow measures as described in ODOT Standard Specification 00280.42 and/or WSDOT Standard Specification 8-01.3(1).
- All exposed soils would be stabilized as directed in measures prescribed in the TESC Plan. Hydro-seed all bare soil areas following grading activities, and re-vegetate all temporarily disturbed areas with native vegetation indigenous to the location. For

- additional detail, consult ODOT Standard Specifications 01030.00 to 01030.90 and/or WSDOT Standard Specification 8-01.3(1).
- Where site conditions support vegetative growth, native vegetation indigenous to the location would be planted in areas disturbed by construction activities. Re-vegetation of construction easements and other areas would occur after the project is completed. All disturbed riparian vegetation would be replanted. Trees would be planted when consistent with highway safety standards. Riparian vegetation would be replanted with species native to geographic region. Planted vegetation would be maintained and monitored to meet regulatory permit requirements. For additional detail, consult ODOT Standard Specifications 01040.00 to 01040.90 and/or WSDOT Standard Specification 8-01.3(2)F.

6.4.4 Work Zone Lighting

- Site work shall follow local, state and federal permit restrictions for allowable work hours. If work occurs at night, temporary lighting should be used in the night work zones. The work area and its approaches shall be lighted to provide better visibility for drivers to travel safely travel through the work zone and illumination shall be provided wherever workers are present to make them visible.
- During overwater construction contractor would use directional lighting with shielded luminaries to control glare and direct light onto work area; not surface waters.

6.4.5 Hydroacoustics

6.4.5.1 Minimization Measure 1 - Drilled Shafts for Foundations

Permanent foundations for each in-water pier would be installed by means of drilled shafts. This approach significantly reduces the amount of impact pile driving, the size of piles, and amount of in-water noise.

6.4.5.2 Minimization Measure 2 – Piling Installation with Impact Hammers

Installation of piles using impact driving may only occur between September 15 and April 15 of the following year. On an average work day, six piles could be installed using vibratory installation to set the piles; then impact driving to drive the piles to refusal per project specifications to meet load-bearing capacity requirements. No more than two impact pile drivers may be operated simultaneously within the same water body channel.

In waters with depths more than 0.67 meters (2 feet) deep, a bubble curtain or other sound attenuation measure would be used for impact driving of pilings. If a bubble curtain or similar measure is used, it would distribute small air bubbles around 100 percent of the piling perimeter for the full depth of the water column. Any other attenuation measure (e.g., temporary noise attenuation pile) must provide 100 percent coverage for the full depth of the pile.

A performance test of the noise attenuation device in accordance with the approved hydroacoustic monitoring plan shall be conducted prior to any impact pile driving. If a bubble curtain or similar measure is utilized, the performance test shall confirm the calculated pressures and flow rates at each manifold ring.

6.4.5.3 Minimization Measure 3 – Impact Pile Installation Hydroacoustic Performance Measure

Sound pressure levels from an impact hammer would be measured in accordance with the hydroacoustic monitoring plan. Recording and calculation of accumulated sound exposure levels shall be performed. Exposure factors shall be calculated using the NMFS moving fish model, based on a fish of over 2 grams with a movement rate of 0.1 meter per second (see Appendix K of the CRC BA 2010). Exposure factors shall account for all attenuated and unattenuated impact pile driving in both the Columbia River and North Portland Harbor. The accumulated sound exposure level shall be recorded.

The following thresholds must not be exceeded:

- 1. The maximum weekly exposure factor shall not exceed 0.18649, based on one calendar week. The weekly exposure factor is defined as the proportion of channel affected by impact pile driving as measured by accumulated sound exposure level multiplied by the proportion of a 24-hr day affected multiplied by the proportion of calendar week affected.
- 2. The maximum yearly (calendar year) total exposure factor shall not exceed 0.202181. The maximum yearly exposure factor is the sum of all weekly exposure factors in one calendar year.
- 3. The average yearly exposure factor must not exceed 0.120090 per calendar year of construction. The average yearly exposure factor is the mean value of all yearly total exposure factors.
- 4. A total exposure factor of 0.480359 shall not be exceeded throughout the construction period of the project. The total exposure factor equals the sum of all weekly exposure factors throughout the project.

One 12-hour rest period would occur each work day in which no impact pile driving would occur. In addition, to limit the exposure of migrating fish that may be present in the behavioral disturbance zone,⁴ impact striking of piles that produce hydroacoustic levels over 150 dB_{RMS} would not occur for more than 12 hours per work day. Unattenuated pile striking may occur to meet the requirements of the hydroacoustic monitoring plan or account for malfunction of the noise attenuation device, but would not occur more than 300 impact pile strikes per week in the Columbia River and no more than 150 impact pile strikes per week in North Portland Harbor. To ensure that this measure is not being exceeded, an approved hydroacoustic monitoring plan would be in place to test a representative number of piles installed during the project (see Minimization Measure 5).

If the predicted accumulated sound exposure level exceeds the levels described above, then the Services would be contacted within 24 hours to determine a course of action, so that incidental take estimates are not exceeded. Necessary steps may include modifications to the noise attenuation system or method of implementation.

6.4.5.4 Minimization Measure 4 - Hydroacoustic Monitoring

The project would conduct underwater noise monitoring to test the effectiveness of noise attenuation devices. Testing would occur based on an underwater noise monitoring plan based on

 $^{^4}$ Behavioral disturbance is expressed in dB_{RMS} (root mean square) re: 1 $\mu Pa.$

the most recent version of the Underwater Noise Monitoring Plan Template (http://www.wsdot.wa.gov/Environment/Air/Noise.htm). This template has been developed in cooperation with NMFS, USFWS, and WSDOT and has been approved by NMFS and USFWS for use in Section 7 consultation for transportation projects in Washington.

Testing would occur according to protocols outlined in an Underwater Noise Monitoring Plan (WSDOT 2008). Underwater noise monitoring would occur as follows:

- Hydroacoustic monitoring would occur for a representative number of piles per structure (minimum of five piles installed with an impact hammer).
- Monitoring would occur for piles driven in water depths that are representative of typical water depths found in the areas where piles would be driven.
- Ambient noise would be measured as outlined in the template in the absence of pile driving.

A report that analyzes the results of the monitoring effort would be submitted to the Services as outlined in the monitoring plan template.

Unattenuated impact pile driving for obtaining baseline sound measurements would be limited to the number of piles necessary to obtain an adequate sample size for the project, as defined in the final Hydroacoustic Monitoring Plan.

6.4.5.5 Minimization Measure 5 - Biological Monitoring

A qualified biologist would be present during all impact pile driving operations to observe and report any indications of dead, injured, or distressed fishes, including direct observations of these fishes or increases in bird foraging activity.

6.4.5.6 Minimization Measure 6 - Temporary Pile Removal

Temporary piles shall be removed with a vibratory hammer and shall never be intentionally broken by twisting or bending. Except when piles are hollow and were placed in clean, sand-dominated substrate, the holes left by the removed pile shall be filled with clean native sediments immediately following removal. No filling of holes shall be required when hollow piles are removed from clean, sand-dominated substrates. At locations where hazardous materials are present or adjacent to utilities, temporary piles may be cut off at the mud line with underwater torches.

6.4.6 Marine Mammal Minimization Measures

6.4.6.1 Equipment Noise Standards

To mitigate noise levels and impacts to sea lions, all construction equipment would comply with applicable equipment noise standards of EPA, and all construction equipment would have noise control devices no less effective than those provided on the original equipment.

6.4.6.2 Sound Attenuation Measures

Specific to pile driving, the hydroacoustic minimization measures listed in Section 6.5.4 would be implemented to reduce impacts to sea lions to the greatest extent practicable.

6.4.6.3 Marine Mammal Monitoring

Establishment of Monitoring Zones

For impact pile driving, a safety zone (defined as where SPLs equal or exceed 190 dB RMS) and a disturbance zone (defined as where SPLs equal or exceed 160 dB RMS) would be established. The initial safety and disturbance zones would be established based on the worst-case underwater sound modeled from impact driving of 36- to 48-inch steel pile.

For vibratory pile or vibratory steel casing installation, an initial disturbance zone (defined as where SPLs equal or exceed 120 dB RMS) would be established based on the worst-case sound modeled from vibratory installation of 36- to 72-inch steel pile for pipe piles or the loudest value modeled for sheet piles. Noise levels for vibratory installation of steel sheet or pipe piles are not anticipated to be above the 190 dB RMS thresholds based on literature values; therefore, no safety zone for vibratory installation of steel pile is anticipated. If steel casings for drilled shafts are installed by a vibratory hammer, an initial safety zone of 5 meters would be established.

Once impact or vibratory installation begins, the safety and disturbance zones would either be enlarged or reduced based on actual recorded SPLs from the acoustic monitoring. The zones would be based on actual acoustic monitoring results collected at an approximate 10-meter distance. If new zones are established based on SPL measurements, NMFS requires each new zone be based on the most conservative measurement (i.e., the largest zone configuration).

Exhibit 6-1 and Exhibit 6-2 show initial monitoring distances for safety and disturbance zones in the Columbia River and North Portland Harbor, respectively.

Exhibit 6-1. Initial Underwater Distance to Safety and Disturbance Monitoring Zones in the Columbia River

Pile Type	Hammer Type	Calculated Distance to Monitoring Zones (meters) ^a		
		190 dB RMS⁵ Safety Zone	160 dB RMS Disturbance Zone (impulse noise)	120 dB RMS Disturbance Zone (continuous noise)
18- to 24-inch steel pipe	Impact	9	858	N/A
36- to 48-inch steel pipe	Impact	54	5,412	N/A
48-inch steel pipe	Vibratory	N/A	N/A	20,166 upriver 8,851 downriver
120-inch steel casing	Vibratory	~5°	N/A	20,166 upriver 8,851 downriver
Sheet pile	Vibratory	N/A	N/A	6,962

a Monitoring zones based on worst case modeled values where the attenuation device is not operating. Upriver and downriver distances vary if a landform is encountered prior to noise attenuating to a threshold value.

b $\,$ All values unweighted and relative to 1 μPa .

c No source value available. To obtain a worst case estimate, distance is based on extrapolation of vibratory sound values from 36- and 72-inch piles.

Exhibit 6-2. Initial Underwater Distance to Safety and Disturbance Monitoring Zones in North Portland Harbor

Pile Type	Hammer Type	Calculated Distance to Monitoring Zones (meters) ^a		
		190 dB RMS ^b Safety Zone	160 dB RMS Disturbance Zone (impulse noise)	120 dB RMS Disturbance Zone (continuous noise)
18- to 24-inch steel pipe	Impact	9	858	N/A
36- to 48-inch steel pipe	Impact	54	3,058 upriver 5,412 downriver	N/A
48-inch steel pipe	Vibratory	N/A	N/A	3,058 upriver 5,632 downriver
120-inch steel casing	Vibratory	~5°	N/A	3,058 upriver 5,632 downriver
Sheet pile	Vibratory	N/A	N/A	3,058 upriver 5,632 downriver

a Monitoring zones based on worst case modeled values where the attenuation device is not operating. Upriver and downriver distances vary if a landform is encountered prior to noise attenuating to a threshold value.

Visual Marine Mammal Monitoring and Pile Driving Shutdown Procedure

The CRC project would develop a monitoring plan in conjunction with NMFS that would collect sighting data for marine mammals observed during activities that include impact or vibratory installation of steel pipe pile, sheet pile, or steel casings. A qualified biologist would be present on site at all times during impact or vibratory installation of steel pile or steel casings. In order to be considered qualified, the biologist would meet the following criteria for marine mammal observers:

- Visual acuity in both eyes (correction is permissible) sufficient for discernment of
 moving targets at the water's surface with ability to estimate target size and distance. Use
 of binoculars may be necessary to correctly identify the target.
- Advanced education in biological science, wildlife management, mammalogy, or related fields (Bachelors degree or higher is preferred).
- Experience and ability to conduct field observations and collect data according to assigned protocols (this may include academic experience).
- Experience or training in the field identification of marine mammals (cetaceans and pinnipeds), including the identification of behaviors.
- Sufficient training, orientation, or experience with the construction operation to provide for personal safety during observations.
- Writing skills sufficient to prepare a report of observations that would include information such as the number and type of marine mammals observed; the behavior of marine mammals in the project area during construction, dates and times when observations were conducted; dates and times when in-water construction activities were conducted; dates and times when marine mammals were present at or within the defined safety zone; dates and times when in-water construction activities were suspended to avoid incidental potential injury from construction noise within the defined safety zone; etc.

b All values unweighted and relative to 1 μ Pa.

c No source value available. To obtain a worst case estimate, distance is based on extrapolation of values from 36- and 72-inch piles.

 Ability to communicate orally, by radio or in person, with project personnel to provide real-time information on marine mammals observed in the area as necessary.

The CRC project proposes the following marine mammal monitoring during any impact or vibratory pile driving:

- Monitoring of safety and disturbance zones would occur for all impact pile driving activities. Monitoring of the disturbance zone would occur for all vibratory pipe or sheet pile installation. No SPLs above 190 dB RMS are anticipated for vibratory installation of pipe or sheet piles; therefore, a safety zone would not be established. If hydroacoustic monitoring of vibratory installation of steel casings for drilled shafts indicates SPLs of 190 dB RMS or higher, then a safety zone would be established and monitored for vibratory installation of steel casings.
- Through acoustic monitoring, the CRC project would determine the actual distance to safety or disturbance zones and establish the new zones at that distance.
- Until determination of safety and disturbance zones is accomplished, monitoring would occur for the area within the calculated zones (Exhibit 6-1 and Exhibit 6-2).
- Safety and disturbance zones would be monitored from a work platform, barge, the existing bridge, or other vantage point or by driving a boat along and within the radius of the zones while visually scanning the area. For activities within a safety zone, full observation of the safety zone would occur. If a small boat is used for monitoring, the boat would remain 50 yards from swimming pinnipeds in accordance with NMFS marine mammal viewing guidelines (NMFS 2007a).
- If vibratory installation of steel pipe piles, sheet piles, or casings occurs after dark, the disturbance zone would be monitored with a night vision scope and/or other suitable device. Vibratory installation of steel pipe piles or sheet piles is not expected to produce SPLs at or above 190 dB RMS; therefore, no safety zone would be established or monitored for these activities. If hydroacoustic monitoring of vibratory installation of steel casings for drilled shafts indicates SPLs of 190 dB RMS or higher, then a safety zone would be established and monitored with a night vision scope and/or other suitable device.
- If the safety zone is obscured by fog or poor lighting conditions, pile driving would not be initiated until the entire safety zone is visible.
- The safety zone would be monitored for the presence of sea lions before, during, and after any pile driving activity.
- The safety zone would be monitored for 30 minutes prior to initiating the start of pile driving. If sea lions are present within the safety zone prior to pile driving, the start of pile driving would be delayed until the animals leave the safety zone.
- Monitoring of the safety zone would continue for 20 minutes following the completion of pile driving.
- Monitoring would be conducted using high-quality binoculars. When possible, digital
 video or still cameras would also be used to document the behavior and response of sea
 lions to construction activities or other disturbances.
- Each monitor would have a radio for contact with other monitors or work crews.
- A GPS unit or electric range finder would be used for determining the observation location and distance to sea lions, boats, and construction equipment.

Data collection would include a count of all sea lions observed by species, sex, age class, their location within the zone, and their reaction (if any) to construction activities, including direction of movement, and type of construction that is occurring, time that pile driving begins and ends, any acoustic or visual disturbance, and time of the observation. Environmental conditions such as wind speed, wind direction, visibility, and temperature would also be recorded.

Shutdown Procedure

The safety zone would also be monitored throughout the time required to drive a pile (or install a steel casing if applicable). If a sea lion is observed approaching or entering the safety zone (190 dB RMS isopleth for pinnipeds), piling operations would be discontinued until the animal has moved outside of the safety zone. Pile driving would resume only after the sea lion is determined to have moved outside the safety zone by a qualified observer or after 15 minutes have elapsed since the last sighting of the sea lion within the safety zone.

Acoustical Monitoring

Hydroacoustic monitoring would be conducted for impact driving of steel piles. Acoustic monitoring would be conducted on a representative number of piles as described in the monitoring plan template that has been developed with and approved by NMFS and USFWS for Section 7 consultations. The number, size, and location of piles monitored would represent the variety of substrates and depths, as necessary, in both the Columbia River and North Portland Harbor. Hydroacoustic monitoring would be condcuted during vibratory installation of at least one pile of the largest diameter used by the project to confirm the distance to the 120 dB RMS threshold level. If steel casings are installed with a vibratory hammer, hydroacoustic monitoring would occur for the first casing installed; this would represent a worst case for size, depth, and substate for vibratory installation of casings. For standard underwater noise monitoring, one hydrophone positioned at midwater depth and 10 meters from the pile is used. Some additional initial monitoring at several distances from the pile is anticipated to determine site-specific transmission loss and directionality of noise. This data would be used to establish the radii of the safety and disturbance zones for sea lions.

Marine Mammal Monitoring Reporting

Reports of the data collected during sea lion monitoring would be submitted to NMFS weekly. In addition, a final report summarizing all sea lion monitoring and construction activities would be submitted to NMFS annually.

6.4.7 Steller and California Sea Lion Minimization Measures⁵

6.4.7.1 Timing Windows

Timing restrictions are used to avoid in-water work when ESA listed species are most likely to be present. CRC would comply with all in-water timing restrictions as determined through the ESA Section 7 and included in permit provisions.

⁵ Note: Because seal and sea lion species present in the Columbia River are protected under the Marine Mammal Protection Act (MMPA), an application for a Letter of Authorization under the MMPA section 101(a)(5)(A) is being submitted to NMFS Office of Protected Resources. The project will comply with any additional minimization measures issued for seals and sea lions as part of the authorization.

6.4.7.2 Equipment Noise Standards

To mitigate noise levels and impacts to sea lions, all construction equipment would comply with applicable equipment noise standards of the EPA, and all construction equipment would have noise control devices no less effective than those provided on the original equipment.

6.4.7.3 Sound Attenuation Measures

Specific to pile driving, the hydroacoustic minimization measures listed in Section 7.1.5 would be implemented to reduce impacts to sea lions to the greatest extent practicable.

6.4.8 Terrestrial Resources

The LPA would impact terrestrial resources, such as migratory birds and SOI, through noise impacts and removal or degradation of habitat. Mitigation measures to address these impacts include impact avoidance and impact minimization. Impact avoidance would be addressed through timing vegetation removal to occur outside of nesting season for migratory birds, and/or developing management plans to provide guidance on ways to meet project construction objectives and timeframes while avoiding violation of the MBTA.

Impact minimization would be addressed by implementing BMPs for erosion and sediment control to protect riparian buffers and sensitive terrestrial habitats (e.g., for riparian species such as pond turtles), appropriate flagging and signing, preservation of native plant species onsite, and other relevant conservation measures. Canada geese and swallows are known to nest on the concrete piers, but not on steel structure portions of the existing bridge; use of the new bridge structures would likely be similar. The I-5 bridge could be inspected at least one full year prior to commencement of construction activities to determine whether any SOI or migratory birds are using the bridge for nesting or roosting. If such species are present, exclusionary devices may be installed on the bridge during the non-nesting season to prevent the bridge from being used for nesting or roosting when demolition activities begin. If high disturbance activities must take place during nesting season, the CRC project team would coordinate with USFWS, ODFW, and WDFW to establish work buffer zones around the nest during nesting season.

To address temporary loss of riparian vegetation resulting from project impacts, mitigation measures could include streambank revegetation and reshaping to restore habitat function, removal of noxious weeds in certain areas, and revegetation of disturbed areas with native species.

6.4.9 Botanical Resources

No sensitive, listed, or otherwise rare plant species are known to occur in the project area. Vegetation removal, including riparian vegetation, is addressed above (Terrestrial Resources).

7. Permits and Approvals

7.1 Federal

The project activities described in this document will be subject to the following federal regulations relevant to protecting fish, wildlife, and their habitat:

- Endangered Species Act. 1973. 16 USC 1531-1544, as amended.
- Migratory Bird Treaty Act. 1936. 16 USC 703-712, as amended.
- Bald and Golden Eagle Protection Act. 1940. 16 USC 668a-d, as amended.
- Magnuson-Stevens Fishery Conservation Management Act. 1976. Public Law 94-265, as amended.
- Marine Mammal Protection Act. Title I. 1972. 16 USC 1361-1389, 16 USC 1401-1407, 1411-1417, and 1421-1421h, as amended.
- Clean Water Act. 1977. 33 USC 1251-1376, as amended.
- Fish and Wildlife Coordination Act. 16 U.S.C. 661-667e, as amended.

7.1.1 Endangered Species Act

The Endangered Species Act (ESA) prohibits the incidental take of any federally listed species. Take is defined in the law to include harass and harm; harm is further defined to include any act which actually kills or injures federally listed species, including acts that may modify or degrade habitat in a way that significantly impairs essential behavioral patterns of the species. Under Section 7 of the ESA, any federal agency that permits, funds, carries out, or otherwise authorizes an action is required to ensure that the action will not jeopardize the continued existence of listed species or result in the destruction or adverse modification of designated critical habitat.

An incidental take permit, obtained through a formal Section 7 consultation with NMFS and/or USFWS, is required if there is potential for the project to adversely impact federally listed species or their critical habitat. Informal consultations occur for projects that result in a "not likely to adversely affect" determination; formal consultations occur for projects that are "likely to adversely affect" listed species.

Formal consultation under Section 7 of the ESA is ongoing with NMFS to analyze effects to listed species and EFH. A Biological Opinion is expected to be issued in early 2011. Informal consultation with USFWS has been completed. A concurrence letter for a "not likely to adversely affect" determination for bull trout and designated critical habitat was issued by USFWS on August 27, 2010.

7.1.2 Migratory Bird Treaty Act

The Migratory Bird Treaty Act (MBTA) regulates the unauthorized taking of migratory bird eggs, young, or adults. Under the MBTA, a permit is required from USFWS if active nests (i.e., those with eggs or young) of migratory birds are destroyed during the breeding season. The breeding season in the project area is approximately March through August, although some birds may breed outside this period. Taking the necessary steps to deter nesting, if possible, in order to

preclude the need for a permit to remove active nests and/or eggs, is generally preferable to obtaining a permit.

7.1.3 Bald and Golden Eagle Protection Act

Administered by the USFWS, this law provides for the protection of the bald eagle and the golden eagle (*Aquila chrysaetos*) by prohibiting, except under certain specified conditions, the taking, possession and commerce of such birds. Golden eagles are not likely to occur within the project area.

Bald eagles, now delisted, are primarily protected under the Bald and Golden Eagle Protection Act (BGEPA). The BGEPA prohibits unregulated take and makes it illegal to kill, wound, pursue, shoot, shoot at, poison, capture, trap, collect, molest or disturb bald or golden eagles. If disturbance will occur in potential violation of the act, a permit to authorize take of eagles is required. This permit authorizes incidental take of bald and golden eagles, as well as incidental take of bald eagles that complies with the terms and conditions of a previously granted Section 7 incidental take statement. Projects permitted under the BGEPA do not need a permit under the MBTA.

There are no documented bald eagle nests within or near the project area; therefore, no permits under the BGEPA are expected to be necessary for this project.

7.1.4 Magnuson-Stevens Fishery Conservation Management Act

The Magnuson-Stevens Act (MSFCMA) affords protection to EFH, which may include streams, lakes, ponds, wetlands, other currently viable water bodies, and most of the habitat historically accessible to salmon. Under MSFCMA, NMFS is required to provide EFH conservation and enhancement recommendations to federal and state agencies for actions that adversely affect EFH. Consultation with NMFS on effects to EFH has been done in conjunction with the Section 7 ESA consultation.

7.1.5 Marine Mammal Protection Act

The Marine Mammal Protection Act (MMPA) is administered by NMFS and provides for the protection of marine mammals by prohibiting, except under certain specified conditions, the taking, possession, and commercial use of such mammals. Under the MMPA, "take" includes to harass, hunt, capture, or kill, or attempt to harass, hunt, capture or kill any marine mammal. Previous analysis of the impact area suggests that marine mammals do utilize this portion of the Columbia River.

CRC submitted an application for a Letter of Authorization (LOA) to NMFS for authorization to incidentally "take" marine mammals (sea lions and harbor seals) over the course of the project. NMFS published the receipt for application in the Federal Register on December 15, 2010, thereby opening the 30-day public comment period (75 FR 78228). The LOA is expected to be issued in mid-2011.

7.1.6 Clean Water Act

Impacts to jurisdictional wetlands or other waters will require a Section 404 permit from the United States Army Corps of Engineers (USACE). For activities that may result in discharge to waters of the U.S., Section 401 of the CWA requires certification that the project will comply with water quality requirements and standards. Dredging, filling, and other activities that alter a waterway require a Section 404 permit and Section 401 certification. The appropriate state

agency must also certify that the project meets state water quality standards and does not endanger waters or wetlands of the state or the United States. Certifications are issued by DEQ in Oregon and by Ecology in Washington.

7.1.7 Fish and Wildlife Coordination Act

This Act authorizes the Secretaries of Agriculture and Commerce to provide assistance to, and cooperate with, Federal and State agencies to protect, rear, stock, and increase the supply of game and fur-bearing animals, as well as to study the effects of domestic sewage, trade wastes, and other polluting substances on wildlife.

7.2 State

7.2.1 Oregon

Work associated with the CRC will be subject to the following Oregon state regulations relevant to protecting fish, wildlife, and their habitat:

- Oregon Endangered Species Act. 2003. Oregon Revised Statutes (ORS) 496.171-192 and Oregon Administrative Rule (OAR) 635-100. Salem, OR.
- Fish Passage; Fishways; Screen Devices; Hatcheries Near Dams. 2001. ORS 509.580-910 and OAR 635-412-0005 to 0040. Salem, OR.
- Goal 5: Natural Resources, Scenic and Historic Areas, and Open Spaces. 1973. OAR 660-15-0000 (5). Salem, OR.
- Oregon's Removal-Fill Law. 2002. ORS 196.800 to 990 and ORS 196.600 to 692.
 Issuance and Enforcement of Removal-Fill Authorizations, OAR 141-085-0005 to 141-089-0615 and Water Quality Standards, 340-041. Salem, OR.

7.2.1.1 Oregon Endangered Species Act

The Oregon ESA applies to actions of state agencies on state-owned or leased lands. In general, the Oregon ESA is much more limited in scope than the federal ESA. The ODFW is responsible for fish and wildlife protected under the Oregon ESA, and the ODA is responsible for plants. The ODFW or ODA may issue a permit to any person for the incidental take of a state-listed threatened or endangered species if it determines that such take will not adversely impact the long-term conservation of the species or its habitat. The department may issue the permit under such terms, conditions, and time periods necessary to minimize the impact on the species or its habitat. An incidental take permit may be issued for individuals of more than one state-listed species. An incidental take permit for state-listed species not covered under the federal ESA would be required from ODFW or ODA.

7.2.1.2 Fish Passage; Fishways; Screen Devices; Hatcheries Near Dams

Oregon's fish passage law has several triggers that initiate compliance requirements. All new culverts, bridges, and dams must meet the current ODFW guidelines for fish passage. If passage is not possible, the law allows for waivers or exemptions to be approved by the ODFW fish passage coordinator or the Oregon Fish and Wildlife Commission, depending on the amount of habitat that will be removed from fish usage. Waivers allow for fish passage to be accomplished off-site, but still within the watershed if a net benefit to fish is shown. Exemptions allow the applicant not to provide passage at the specific site, but passage could be required in the future if watershed conditions change.

A fish passage plan will be submitted to ODFW and approved when project designs are further advanced. The project will meet all fish passage criteria for the state of Oregon.

7.2.1.3 Goal 5: Natural Resources, Scenic and Historic Areas, and Open Spaces

Goal 5 requires local governments in Oregon to protect natural resources and conserve scenic and historic areas and open spaces by adopting programs to protect these resources. Permitting may be required through local government Goal 5 ordinances. Goal 5 planning related to ecosystem resources within the I-5 CRC project includes the following:

- Fish and wildlife areas and habitats should be protected and managed in accordance with the Oregon Fish and Wildlife Commission's fish and wildlife management plans. The nearest ODFW Wildlife Management Area is on Sauvie Island, adjacent to the downstream extent of the project area (ODFW 2010).
- Stream flow and water levels should be protected and managed at a level adequate for fish, wildlife, pollution abatement, recreation, aesthetics and agriculture.
- Significant natural areas that are historically, ecologically or scientifically unique, outstanding or important, including those identified by the State Natural Area Preserves Advisory Committee, should be inventoried and evaluated. The study area includes numerous "significant natural areas"; however, at this time it does not include any areas specifically identified by the Oregon Natural Heritage Advisory Council (ONHAC 2010).
- Plans should provide for the preservation of natural areas consistent with an inventory of scientific, educational, ecological, and recreational needs for significant natural areas.

7.2.1.4 Oregon's Removal-Fill Law

Impacts to jurisdictional wetlands or other waters of the state (e.g., fill or removal activities below the bankfull stage or the line of non-aquatic vegetation, whichever is higher) require a removal-fill permit from DSL. This permit would typically be obtained in conjunction with a federal Section 404 permit (Section 11.1.6) via a joint permit application for impacts to wetlands and jurisdictional waters; a wetland delineation and conceptual mitigation plan would also be required.

7.2.1.5 Wildlife Policy

It is the policy of the State of Oregon that wildlife would be managed to prevent serious depletion of any indigenous species. An in-water blasting permit is required from ODFW if the project alternatives include in-water blasting. This permit is required if explosives are used when removing any obstruction in any waters of this state, in constructing any foundations for dams, bridges or other structures, or in carrying on any trade or business. ODFW issues in-water blasting permits only if they contain conditions for preventing injury to fish and wildlife and their habitat.

No in-water blasting is expected to be necessary during the course of this project; therefore, no permit is likely to be required under this policy.

7.2.2 Washington

Work associated with the CRC will be subject to the following Washington state regulations relevant to protecting fish, wildlife, and their habitat:

- State Environmental Protection Act (SEPA). 1971. Revised Code of Washington (RCW) 43.21C, and Washington Administrative Code (WAC) 197-11 and WAC 468-12. Olympia, WA.
- Habitat buffer zones for bald eagles. 1984. RCW 77.12.655. Bald eagle protection rules. 1986. WAC 232-12-292. Olympia, WA.
- Shoreline Management Act of 1971. 1971. RCW 90.58, WAC 173-18-100 and WAC 173-22. Olympia, WA.
- Hydraulic Code. 1949. Chapter 77.55 RCW. Olympia, WA.
- Fishways, flow, and screening. 1949. RCW 77.57, as amended. Olympia, WA.
- Clean Water Act certification.

7.2.2.1 State Environmental Protection Act

SEPA requires all governmental agencies to consider the environmental impacts of a proposed action before making decisions. An environmental impact statement (EIS) must be prepared for all proposals with probable significant adverse impacts on the quality of the environment. State and local agencies may approve an EIS prepared under NEPA to fulfill the SEPA evaluation requirement.

7.2.2.2 Bald Eagle Protection Rules

The current bald eagle protection rules are only applicable and enforceable when the bald eagle is listed under state law as threatened or endangered. The bald eagle is not currently listed under state law as threatened or endangered. However, landowners must still comply with the federal Bald and Golden Eagle Protection Act, and should contact the USFWS to determine if a permit is required when proposing land use activities within 660 feet of an eagle nest.

7.2.2.3 Shoreline Management Act of 1971

Under the Shoreline Management Act (SMA), each city and county is required to adopt a shoreline master program that is based on state guidelines and that may be tailored to the specific geographic, economic, and environmental needs of the community (Ecology 2009c). A permit will be required from the City of Vancouver for project activities occurring along the shoreline of the Columbia River or Burnt Bridge Creek.

7.2.2.4 Hydraulic Code

The Hydraulic Code is intended to ensure that required construction activities are performed in a manner to prevent damage to the state's fish, shellfish, and their habitat. An HPA from WDFW will be required for work occurring within waters of the state (defined as all salt and fresh waters waterway of the OHW and within the territorial boundary of the state).

7.2.2.5 Fishways, Flow, and Screening

Washington's fish passage regulations describe requirements for fish screens or bypasses when a lake, river, or stream containing game fish will be diverted, and for fishways, if an obstruction will be placed in a stream. An HPA will be required (Hydraulic Code), and a permit from Ecology will be required if water is diverted.

7.2.2.6 Clean Water Act Certification

This certification would typically be obtained from Ecology in conjunction with a federal Section 404 permit and a 401 certification (Section 11.1.6) via a joint permit application for impacts to wetlands and jurisdictional waters; a wetland delineation and conceptual mitigation plan would also be required.

7.3 Local

7.3.1 Oregon

Work on the Columbia River Crossing will be subject to the following Oregon local regulations relevant to protecting fish, wildlife, and their habitat:

- Environmental Zones. 1994. City of Portland Code (CPC) 33.430, as amended. Portland, OR
- Tree Cutting. 2002. CPC 20.42. Portland, OR.
- Nature in Neighborhoods. 2005. Metro Code Sections 3.07.130 3.07.1370) Title 13.
- Floodplain Overlay District. 2009. City of Gresham Code 5.0120. Gresham, Oregon.
- Special Purpose Overlay District. 2006. City of Gresham Code 10.221. Gresham, Oregon.

7.3.1.1 Environmental Zones

Permits are required for development or disturbance within environmental zones. Applicable permits will be completed for the CRC project as project designs and timelines are finalized.

The environmental zones provide for fish habitat protection through the designation of environmental protection or conservation zones. Development within these zones requires a permit application and additional information. Natural resource management plans (NRMPs) may be developed and approved, and may contain regulations that supersede or supplement the environmental zone regulations. These regulations will apply when a building permit or development permit application is requested within the resource area of the environmental conservation zone and is subject to the Development Standards of Section 33.430.110-170. These regulations do not apply to building or development permit applications for development that has been approved through environmental review. Environmental review is overseen by the City of Portland Land Use Review process.

7.3.1.2 Tree Cutting

A permit to cut trees on private or public property within the project area may be required from the City of Portland. Urban Forestry also regulates the cutting and planting of trees on public property, including street trees located on the public right-of-way. Permits are required to plant, prune, remove, or cut the roots of any tree located on public property.

7.3.1.3 Nature in Neighborhoods

The purpose of Nature in Neighborhoods is to conserve, protect, and restore a continuous ecologically viable streamside corridor system that is integrated with upland wildlife habitat and the surrounding urban landscape.

7.3.1.4 Floodplain Overlay District

According to City of Gresham code, within the floodplain overlay district of Fairview Creek, proposed developments need to comply with the guidelines and recommendations of the Fairview Creek master storm drain plan and would need to be accompanied by documentation prepared by a registered civil engineer demonstrating that the development would not result in an increase in floodplain area on other properties, reduce natural flood storage volumes, or result in an increase in erosive velocity of the stream that may cause channel scouring or reduced slope stability downstream of the development.

7.3.1.5 Special Purpose Overlay District

Sites specified by the Inventory of Significant Natural Resources and Open Spaces as having particular importance as fish and wildlife habitat areas shall be designated on the Community Development Special Purpose District Map as Natural Resource (NR) districts. The NR districts shall function as a special purpose overlay district.

Measures shall be adopted in the Community Development Code and Standards document to restrict development proposed within or adjacent to an NR district site. These measures shall require any such development to take place in a manner which minimizes adverse impacts on the resource site. Findings of public need and lack of alternative sites shall be required in connection with any proposed development activity within an NR district site.

7.3.2 Washington

Work on the Columbia River Crossing will be subject to the following Washington local regulations relevant to protecting fish, wildlife, and their habitat:

- Critical Areas Protection Ordinance. 2005. City of Vancouver Vancouver Municipal Code (VMC) 20.740; Fish and Wildlife Habitat Conservation Areas. 2005. VMC 20.740.110. Vancouver, WA.
- Shoreline Management Area. 2005. VMC 20.760. Vancouver, WA.
- Critical Areas and Shorelines. 2005. Clark County Code. Title 40.4. Vancouver, WA.
- SEPA Regulations. 2004. VMC 20.790.
- Street Trees. VMC 12.04; and Tree Conservation. VMC 20.770. Vancouver, WA.
- Water Resources Protection. VMC 14.26.

7.3.2.1 Critical Areas Protection Ordinance (City of Vancouver)

The CAO applies to habitat for any life stage of state or federally designated endangered, threatened, or sensitive fish or wildlife species, priority habitats and habitats of local importance, riparian management areas and riparian buffers, and water bodies. Critical Areas Protection also regulates development in the floodplain and in erosion hazard areas, both of which occur in the project area. A critical areas report will be required as part of the submittal for a Critical Areas Permit, which is required for project activities occurring on properties containing critical areas or buffers. A Critical Areas Report for a riparian management area or riparian buffer must include an evaluation of habitat functions using the Clark County Habitat Conservation Ordinance Riparian Habitat Field Rating Form or another habitat evaluation tool approved by the WDFW.

7.3.2.2 Shoreline Management (City of Vancouver)

A Substantial Development Permit will be required for project activities occurring within areas regulated by the Shoreline Management Master Program (see discussion above in the Washington state section).

7.3.2.3 Critical Areas and Shorelines (Clark County)

Clark County has designated Critical Areas in accordance with the Growth Management Act (GMA). A permit may be required if the project occurs in habitat conservation areas, wetlands protected by Clark County Code, or along unincorporated Clark County shorelines.

7.3.2.4 State Environmental Protection Act

The NEPA EIS will be submitted to state and local agencies which may adopt the NEPA EIS to fulfill SEPA requirements (see discussion above in the Washington state section).

7.3.2.5 Street Trees

Street Trees and Tree Conservation municipal codes require permits if the project alternative results in the cutting of trees on public or private property. There are two kinds of permits required for trees in the City: one for street trees and one for private trees. If the tree is in the public right-of-way, a street tree permit is required.

7.4 Regional and Local Resource Protection

Work on the Columbia River Crossing will be subject to Oregon and Washington regional planning zones and guidelines relevant to protecting fish, wildlife, and their habitat.

7.4.1 Oregon

7.4.1.1 City of Portland

The City of Portland applies two environmental overlay zones—protection and conservation—to various sites throughout the city to protect natural resources. The conservation overlay zone is intended to conserve important natural resources and their functions. This zone applies to areas where natural resources can be protected while allowing environmentally sensitive development. Environmental zoning is applied to all development and site disturbance activities. The Columbia River, North Portland Harbor, and Columbia Slough are zoned *conservation*.

The environmental protection overlay zone offers the highest level of protection for the city's sensitive natural resources. This zone typically covers a stream, streamside area, wetland, or large forested area, and is essentially a *no-build* zone because development in these areas would degrade Portland's most important and sensitive natural resources. Some projects may be allowed if there is a clear public benefit (trails and interpretive facilities) or if there is no feasible project location outside of the protection zone (access). No lands in the project area are in designated protection zones.

7.4.1.2 Metro

In 2004, Metro updated its December 2002 inventory of riparian and upland habitat. Metro defines riparian habitats as land and vegetation located near rivers, streams, lakes and wetlands; upland habitats are natural areas providing wildlife with food and shelter and allowing movement

from one habitat to another. Based on this inventory, Metro identified regionally significant habitat. These areas were then mapped with a ranking of low, medium, or high based on their capacity to protect fish and wildlife (Metro 2005).

7.4.1.3 City of Gresham

The City of Gresham inventoried wetland, riparian areas, and upland areas in the fall of 1987. The findings of this survey are summarized by the Inventory of Significant Natural Resources and Open Spaces that was adopted by the City of Gresham as an appendix to the Community Development Plan (City of Gresham 2005, 2006). This survey was oriented primarily toward wildlife habitat values of lowland and upland natural areas within the City of Gresham. The resource areas included in the inventory are significant wildlife habitats and noteworthy scenic features that perform a variety of useful natural functions, including retention of soils, pollution control, groundwater recharge, and flood control. Forty-five sites having potential significance as natural resource areas were identified within the City of Gresham and include wetlands, riparian corridors, upland areas, and greenways (City of Gresham 2005).

7.4.2 Washington

7.4.2.1 Priority Habitats

The Washington Department of Fish and Wildlife (WDFW) has established priority habitat areas within the state. Priority habitats are those habitats with "unique or significant value to a variety of different species" (WDFW 2008), and may consist of a unique vegetation type or dominant plant species, a described successional stage, or a specific structural element. Washington has identified 18 priority habitat types. Within the project area, established priority habitats include Riparian, Urban Natural Open Space, and Oak Woodland. These priority habitats were not field-verified during the September 2005 surveys.

7.4.2.2 Riparian

Riparian habitats are those areas adjacent to aquatic systems with flowing water that contain elements of both aquatic and terrestrial ecosystems that mutually influence each other. In riparian systems, perennial or intermittent water bodies influence the vegetation, water tables, soils, microclimate, and wildlife of terrestrial ecosystems. The biological and physical properties of the aquatic ecosystems are influenced by adjacent vegetation, nutrient and sediment loading, terrestrial wildlife, and organic and inorganic debris. Riparian habitats begin at the OHW and extend to the portion of the terrestrial landscape influenced by, or directly influencing, the aquatic ecosystem. Riparian habitat includes the entire extent of the floodplain and riparian areas of wetlands directly connected to stream courses (WDFW 2006).

The criteria used by WDFW for establishing priority riparian habitats include high fish and wildlife density, high fish and wildlife species diversity, important fish and wildlife breeding habitat, important wildlife seasonal ranges, important fish and wildlife movement corridors, high vulnerability to habitat alteration, and unique or dependent species (WDFW 2006).

7.4.2.3 Urban Natural Open Space

Urban Natural Open Spaces are isolated remnants of natural habitat larger than 4 hectares (ha) (10 acres) and surrounded by urban development, although local considerations may be given to smaller open space areas (WDFW 2006). Natural open spaces in urban areas are priority habitat due to the limited amount of such habitat. One or more priority species may reside within or

adjacent to the open space and use it for breeding and/or feeding or the open space may function as a corridor connecting other priority habitats, especially those that would otherwise be isolated.

7.4.2.4 Oak Woodland

Oak Woodland priority habitats are those habitats with stands of pure oak or oak/conifer associations where canopy coverage of the oak component of the stand is at least 25 percent, or where total canopy coverage of the stand is less than 25 percent but where oak accounts for 50 percent or more of the canopy coverage present (oak savannah). In urban areas, single oaks or stands less than 0.4 ha (1 acre) are considered a priority when valuable to fish and wildlife. The criteria for this priority habitat are comparatively high fish and wildlife density, high fish and wildlife species diversity, limited and declining availability, high vulnerability to habitat alteration, and dependent species.

7.4.2.5 Critical Areas

The GMA requires cities and counties to designate and protect "critical areas," including fish and wildlife habitat, wetlands, flood hazard areas, geologic hazard areas, and critical aquifer recharge areas. Both Clark County and the City of Vancouver have passed ordinances designating critical areas. The City of Vancouver has jurisdiction only over critical areas within its boundary. Clark County has jurisdiction over critical areas in the unincorporated areas of the County.

7.4.2.6 City of Vancouver

The City of Vancouver protects priority habitat areas through its Critical Areas Protection Ordinance. Critical areas include fish and wildlife habitat conservation areas, wetlands, frequently flooded areas, critical aquifer recharge areas, and geologic hazard areas as defined by the GMA. Fish and wildlife habitat conservation areas include, but are not limited to, habitat for any life stage of state-designated or federally designated endangered, threatened, and sensitive fish or wildlife species, priority habitats and habitats of local importance, riparian management areas and riparian buffers, and water bodies. The City of Vancouver also applies the WDFW priority habitat designations.

7.4.2.7 Clark County

In Clark County, mapped critical areas include Riparian Priority Habitat, Other Priority Habitats and Species (PHS), and Locally Important Habitats and Species. Locally Important Habitats and Species areas are areas legislatively designated and mapped by the County because of unusual or unique habitat that warrants protection due to qualitative species diversity or habitat system health indicators. Such areas are designated as critical, sensitive, or both critical and sensitive.

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